

Vertical Flux of Trace Elements Associated With Lithogenic and Biogenic Carrier Phases in the Southern Ocean

Stéphane Blain, H. Planquette, I. Obernosterer, A. Guéneuguès

► To cite this version:

Stéphane Blain, H. Planquette, I. Obernosterer, A. Guéneuguès. Vertical Flux of Trace Elements Associated With Lithogenic and Biogenic Carrier Phases in the Southern Ocean. Global Biogeochemical Cycles, 2022, 36 (5), 10.1029/2022GB007371. hal-03684981

HAL Id: hal-03684981 https://hal.science/hal-03684981

Submitted on 1 Jun2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	Vertical flux of trace elements associated with lithogenic and biogenic carrier			
2	phases in the Southern Ocean			
3	S. Blain ¹ , H. Planquette ² , I. Obernosterer ¹ , A. Guéneuguès ¹			
4				
5	¹ Sorbonne Université, CNRS, Laboratoire d'océanographie microbienne (LOMIC), 1 avenue			
6	Pierre Fabre, 66650 Banyuls sur mer, France.			
7	² CNRS, IRD, Ifremer, LEMAR, University of Brest, Plouzané, France.			
8				
9	Corresponding author: Stéphane Blain (stephane.blain@obs-banyuls.fr)			
10				
11	Key Points:			
12	• Export fluxes of 15 trace elements reveal contrasted seasonal patterns between			
13	lithogenic and biological carriers.			
14	• Basalt particles are the major lithogenic carrier phase of 9 trace elements.			
15	• Fecal pellets, diatom vegetative cells and spores are each carriers of distinct trace			
16	elements.			

18 Abstract

Trace elements (TE) are tracers of multiple biotic and abiotic processes in the ocean and some 19 of them are essential for marine life. Vertical export by particles is a major removal process of 20 a large fraction of TE from the surface ocean. However, the seasonal export dynamics and its 21 controlling factors, critical for the understanding of the internal TE cycling, remain poorly 22 constrained. Here, we report and discuss the seasonal export of 15 TE in sinking particles 23 collected by a sediment trap deployed in a highly productive region of the Southern Ocean. 24 25 Basalt material was the main carrier phase for the export flux of 9 TE, and its dynamic was characterised by a strong decrease over time. TE export driven by biological carriers such as 26 diatom spores and vegetative cells added pulsed seasonal dynamics to the lithogenic signal, 27 while the contribution of fecal pellets was less variable over the season. For each TE, we were 28 able to decipher the biological carrier phases that represent the most dominant export 29 30 pathway. We discuss this partitioning with regards to the known metabolic functions of the different trace metals or TE of biological interest. 31

32

34

35 **1 Introduction**

36 During the past decade the international project GEOTRACES has undertaken an

37 unprecedented effort to improve our knowledge of the distribution and biogeochemical cycles of trace elements and their isotopes (TEI) in the ocean (Anderson 2020). Trace metals are 38 required in many metabolic functions (Sunda 2012) and as such, biogenic particles that are 39 generated in the upper ocean are one of the main players regulating the internal cycling of so 40 called "bioactive" trace elements (TE). TE are incorporated into particles through biological 41 42 uptake and/or passive adsorption; they can then be remineralized, desorbed and/or ultimately exported. Particle sinking is one of the mechanisms of downward transport of chemical 43 44 elements (Boyd et al. 2019). Large biogenic particles such as phytoplankton aggregates or 45 fecal pellets are major vectors in the downward transport of particulate organic carbon (POC), but lithogenic particles can also play a role by ballasting aggregates and reducing 46 remineralisation (Lemaitre et al. 2020). Due to their role in the control of atmospheric CO_2 47 (Antia et al. 2001), carbon vertical export fluxes have been extensively studied, yet TE export 48 fluxes have been considerably less investigated. 49

50 Vertical fluxes of particulate TE can be determined in different ways (McDonnell et al. 2015). Among them, moored sediment traps have been used since the 1970s to measure vertical 51 fluxes of sinking material in the ocean. Important characteristics of vertical fluxes were 52 revealed by this approach, but possible biases and limitations were also identified leading to 53 the delivery of best practises (Buesseler et al. 2007). Taking these recommendations into 54 account for deployment of the moorings and their design, moored sediment traps are powerful 55 56 tools and are quite unique to capture long term variability (months to years) of sinking fluxes including TEs (Kremling and Streu 1993; Huang and Conte 2009; Kuss et al. 2010; Conte et 57 al. 2019). In the northern Sargasso Sea, the oceanic flux program (OFP) provides the longest 58 time series for elemental composition of export fluxes at three depths. Data collected between 59

2000 and 2015 were used to build mean seasonal cycles for 19 elements with a monthly 60 61 temporal resolution (Conte et al. 2019). Based on the assumption that the elemental composition of the upper continental crust approximated lithogenic material composition in 62 the traps, elemental fluxes were partitioned between different phases (organic matter, 63 carbonates, lithogenic, authigenic). Two components have been identified as main drivers of 64 the seasonal dynamics of the elemental fluxes. One of them was coupled to the seasonal cycle 65 66 of primary production and surface export. The other one was related to internal processes associated with chemical scavenging and particle aggregation (Conte et al. 2019). 67 Based on a 13 year time series of elemental flux composition at 2000 m in the North East 68 69 Atlantic, the mean annual cycles (monthly resolution) of 13 elements were reported (Pullwer and Waniek 2020). Overall, depending on the element considered, weak or no seasonality was 70 detected. This was likely due to a small biological signal at the seasonal level, which was 71 72 further damped by interannual variability of environmental conditions in surface waters. The depth of the traps also likely contributed to mask a clear seasonal signal at this site. 73 74 In the Southern Ocean, moored sediment traps were also widely used to investigate carbon export dynamics (Honjo et al. 2000). However, studies of the seasonal dynamics of TE export 75 are rare. The seasonality of particulate export fluxes of 8 TE was studied in the polynya of 76 Pridz Bay (Sun et al. 2016). Seasonal variations of Cu, Zn and Cd were mainly driven by ice 77 coverage and biological production whereas fluxes of Al, Fe and Mn mainly derived from 78 continental debris were controlled by ice melting and freezing processes. 79 80 Further investigations of the seasonal export dynamics of TE, with high temporal resolution are therefore required. We have addressed this challenge in the Kerguelen Plateau, a 81 productive region of the Southern Ocean, where iron fertilisation leads to a marked seasonal 82 pattern of carbon export (Rembauville et al. 2015a; Blain et al. 2020). In this context, we 83 aimed to study and understand the various processes impacting the stoichiometry and the 84

magnitude of these export fluxes, including the seasonal dynamics that can facilitate the
partitioning of TE export between different carrier phases.

87

88 2 Material and methods

89 **2.1 Sediment trap and sensors.**

90 The sediment trap mooring was deployed during the SOCLIM cruise

91 (doi/10.17600/16003300) on October 13th 2016, at a station located on the central Kerguelen

92 plateau (50°38'344 °S, 71°59'854 °E) (Figure 1A) in the core of a productive region which is

naturally iron fertilized (Blain et al. 2007). The bottom depth is 527 m. Two consecutive

94 diatom blooms occur annually (Blain et al. 2020) with peaks in chlorophyll concentrations

95 within the mixed layer in November and in December (Figure 1B).

We used a Technicap PPS3 sediment trap (0.125 m² collecting area, 4.75 aspect ratio) located 96 97 at 292 m below the surface. The cups were prepared using trace metal clean protocols in a clean room. The cups were washed with warm solution alkalin detergent (Extran) for 24 98 hours, rinsed 3 times with distilled water, then soaked in 2M HCl (analytical grade) for one 99 100 week, rinsed 3 times with MQ water, soaked in 0.2 M HCl (ultrapure) for 1 week and finally rinsed 3 times with MQ water. The cups were then stored in plastic bags until used. Following 101 102 the protocol of the preparation of the trace metal clean AQUIL medium (Price et al. 1989), the hypersaline formalin solution buffered at pH=8 with sodium tetraborate was passed through a 103 Chelex resin to remove trace metal contamination. Trace metal concentrations of this solution 104 105 can be found in Table S2. Just prior to the deployment, the 12 cups (250mL) were filled with the 5% preservative solution and mounted on the sediment trap carousel. The collection time 106 for each cup was 11 days (Table 1). A current meter (Aquadopp) and an inclinometer were 107 attached to the sediment trap to record measurements of current speeds and inclination of the 108 trap at a frequency of 1 h^{-1} . 109

After recovering the sediment traps on April 3rd 2017, 1 mL of the supernatant of the cups 110 was immediately replaced by fresh hypersaline formalin buffered (pH=8) solution before 111 storage at room temperature until further processing. Four months later, at the home 112 laboratory, samples were first transferred to a Petri dish and examined under a 113 stereomicroscope (Leica MZ8, x10 to x50 magnification) to remove swimmers (i.e. organisms 114 for which the structure was well preserved and that actively entered the cup). Then the 115 116 samples were split into eight aliquots using a Jencons peristaltic splitter (Rembauville et al. 117 2015a).

118

119 **2.2 Bulk chemical analysis.**

Aliquots for chemical analyses were centrifuged for 5 min at 3000 rpm. After this step, the
supernatant was withdrawn and replaced by Milli-Q-grade water to remove salts.
This rinsing step was repeated three times. The remaining pellet was freeze-dried (SGD-

123 SERAIL, 0.05–0.1 mbar, -30 to 30°C, 48 h run) and weighed three times (Sartorius MC

124 210 P balance, precision of 10^{-4} g) to calculate the total mass. The particulate material was

then ground to a fine powder and used for further elemental analysis.

126

127 2.2.1 Mass, total POC/PON, BSi, CaCO₃.

128 For particulate organic carbon (POC) and particulate organic nitrogen (PON) analyses, 3 to 5

129 mg of the freeze-dried powder was weighed directly into pre-combusted (450°C, 24 h) silver

130 cups. Samples were decarbonated by adding 20 µL of 2M analytical-grade HCl (Sigma-

- 131 Aldrich). Samples were dried overnight at 50°C. POC and PON were measured with a CHN
- analyser (Perkin Elmer 2400 Series II CHNS/O elemental analyser) calibrated with glycine.
- 133 Samples were analysed in triplicate with an analytical precision of less than 0.7 %.

134 For BSi analysis, 2 to 8 mg of material was used. For BSi sample digestion we followed the

135 protocol from (Ragueneau et al. 2005) and the silicic acid concentrations in the solutions were

determined manually following Aminot and Kérouel (2007). The precision of BSi

- 137 measurement was 10% (Ragueneau et al. 2005).
- 138 For bulk CaCO₃ analyses, 5 mg of freeze-dried material was weighed into Teflon vials for the

139 mineralization. One mL of 65% (v/v) HNO₃ (Sigma analytical grade) was added and samples

140 were placed in an ultrasonication bath for 20min. Samples were then dried overnight at 130

141 °C, then 0.5 mL of 40% (v/v) HF (Sigma analytical grade) and 5 mL of 65% HNO₃ were

added. The samples were ultra-sonicated a second time and dried overnight. The resulting

residue was dissolved in 10 mL of 0.1N HNO₃ and the calcium (Ca) content was analyzed by

144 inductively coupled plasma – optical emission spectrometry (ICP-OES, Perkin-

145 ElmerOptima2000). The efficiency of the mineralization procedure was estimated using the

reference material GBW-07314. The efficiency was 96% and the precision of the Ca

147 measurement was 2% (Rembauville et al. 2016). Based on a Ca/Ti ratio in basalt of Kerguelen

148 (3.3 mol/mol) we estimate that the contribution of Ca of lithogenic origin to the total Ca flux

149 was low. We therefore equalled the Ca flux to the biological CaCO₃ flux.

150

151 2.2.2 Elemental analysis by SF-ICP-MS

For elemental analysis by SF-ICP-MS, between 12 and 40 mg of dried material were transferred into clean PFA vials and were digested in a mixture of 8.0M HNO₃ (Merck ultrapur) and 2.9M HF (Merck suprapur). Vials were tightly capped and heated to 130° C for 4 hours. The remaining solution was then evaporated to near dryness, then 400 µL of concentrated HNO₃ (Merck ultrapur) was added to drive off the fluorides and was then evaporated. Finally, samples were redissolved with 3mL of 3% HNO₃ (Merck Ultrapur) and kept in acid-cleaned 15mL polypropylene tubes (Corning[®]) until analysis by SF-ICP-MS (see details below). This procedure has been proven adequate for digestion of all particulate trace metals (Planquette andSherrell 2012).

All archive solutions were analyzed by SF-ICP-MS (Element XR) following the method of 161 162 Planquette and Sherrell (2012). Final concentrations of samples and procedural blanks were calculated from In-normalized data. Analytical precision was assessed through replicate 163 samples (every 10th sample) and accuracy was deduced from analysis of Certified Reference 164 Materials (CRMs) of plankton (BCR-414) and sediments (PACS-3 and MESS-4) 165 (Supplementary Table S1). Dissolved Mn, Fe, Cu, and Co concentrations of the saline solution 166 were determined before deployment and after the recovery in an aliquot collected after the 167 168 centrifugation, by SF-ICP-MS after preconcentration using the SeaFast (Supplementary Table S2) following the method described previously (Tonnard et al. 2020). Based on these results 169 we calculated the percentage of dissolution of the particulate material within the cups 170 (Supplementary Table S2). We did not correct particulate flux for dissolution as the values are 171 generally low (<10%) with the exception of Mn (23.5%) in cup #11 and Cu in cups #7 (11.4 172 173 %), #8 (10.1%) and #11 (30.8%).

174

175 **2.3 Carbon export fluxes of diatoms and faecal pellets**.

Microscopic observations were conducted within four months after recovery of the moorings. 176 For the identification of diatoms, counting and size measurements, we followed the protocol 177 178 described in Rembauville et al. (2015a) that allows the separate consideration of full and empty cells. For diatom counting, the samples were processed as follows. Two mL of one-eighth 179 aliquot was diluted with 18 mL of artificial seawater and decanted in a Sedgewick Rafter 180 counting chamber. Full diatoms were enumerated and identified under an inverted microscope 181 with phase contrast (Olympus IX170) at 400x magnification. The morphometric measurements 182 were done using high resolution images (Olympus DP71 camera) and Fiji image processing 183

184 software. The biovolume was calculated from morphometric measurements (Hillebrand et al.185 1999).

186 The export flux of diatoms (Cell $m^{-2} d^{-1}$) was calculated using the equation:

187
$$Cell flux = N_{diat} \times d \times 8 \times V_{aliquot} \times \frac{1}{0.125} \times \frac{1}{11} \times k$$
 (Eq. 1)

188 Where N_{diat} (cell mL⁻¹) is the number of cells counted in one chamber, d is the dilution factor, 189 8 relates to measurements being made on a one eighth aliquot of the sample, $V_{aliquot}$ (mL) is the 190 volume of the aliquot, 1/0.125 relates to the trap surface area (in m²), 1/11 relates to the sample 191 collection time (in days), and k is the fraction of the chamber counted.

192 The diatom flux was then converted to POC flux for each taxon using allometric equations

reported in the literature (Menden-Deuer and Lessard 2000; Cornet-Barthau et al. 2007) and taking into account specific relationships for spores (Rembauville et al. 2015a) (Table S3). The spore and vegetative cell carbon fluxes were then obtained by summing up the contribution of the different taxa. The sum of both fluxes corresponded to the carbon flux associated with diatoms.

To enumerate fecal pellets, an entire one-eighth aliquot of each sample cup was placed in a 198 gridded Petri dish and observed under a stereomicroscope (Zeiss Discovery V20) coupled to a 199 camera (Zeiss Axiocam ERc5s) at 10X magnification. Fecal pellets were classified into three 200 types according to their shape: spherical, cylindrical, ovoid/ellipsoid (Table S3) (Gleiber et al. 201 202 2012). Size measurements were used to calculate the volume of each fecal pellet according to their shape that was then converted to carbon using a factor of 0.036 mg C mm⁻³ (González and 203 Smetacek 1994). The fecal pellets carbon fluxes (F_{fp} (mg C m⁻² d⁻¹)) in the different size classes 204 205 were calculated using the equation:

206
$$F_{fp} = C_{fp} \times 8 \times \frac{1}{0.125} \times \frac{1}{11}$$
 (Eq. 2)

where C_{fp} (mg C per fecal pellets for each type) is the concentration of carbon in each fecal pellet type. Others terms in the equation have the same definition as in Eq. 1). The F_{fp} were finally summed to provide the total fecal carbon fluxes. Although the calculation of total POC
flux is associated with large uncertainties (around 50%, (Rembauville et al. 2015a), the linear
regression between POC_{calculated} and POC_{measured} was as follows:

212 POC_{calculated} = (0.84 ± 0.05) x POC_{measured} + (0.2 ± 0.35) with R²=0.9621)

213

214 **2.4 Statistics tools and data visualisation**.

Statistical analysis (cross-correlation, Principal Component Analysis (PCA) and Partial least
Square Regression (PLSR)) were performed using scikit-learn packages python 2.7. Scipy.stats

217 package python 2.7 was used to conduct ANOVA after checking for homoscedasticity with a

levene test. Data visualisation was realised with python 2.7 matplot library.

219

220 **3 Results**

3.1 Physical conditions at the depth of the sediment trap

The average depth of the sediment trap was 293 ± 2 m (n=3152) with a few short and episodic deepening events below 300 m (Fig. 2A). The mean inclination angle of the sediment trap was $0.8^{\circ} \pm 1^{\circ}$ (Fig. 2B). Inclination angles above 2° were rare and associated with deepening events of the trap and current speeds exceeding 0.2 m s⁻¹. The mean current speed was 0.13 ± 0.07 m s⁻¹. The short-term variability of current speed and direction (Fig. 2C and 2D) was driven by tide (see Rembauville et al. 2015a) for a detailed study at the same site and depth) and a window of 26 h is adequate for filtering this short-term variability.

229

3.2 Seasonal changes of mass flux and biological export.

231 The seasonal variations of export fluxes were determined for particle mass, total POC and PON,

for CaCO₃ and BSi. POC was further partitioned between different biological carrier phases:

total diatoms (POC_{diat}), separated into diatom spores (POC_{spore}) and diatom vegetative cells

(POCveg), and fecal pellets (POCfp)(Fig. 3). Seasonal variations of particle mass, POC, POCdiat, 234 235 and BSi fluxes were characterised by two peaks of export. The first export event occurred between 12 Nov 2016 and 15 Dec 2016 and was recorded in cups #3 to #5. A second export 236 237 event occurred between 17 Jan 2017 and 08 Feb 2017 and was recorded in cups #9 and #10. The export of POC_{spore} took place largely during the first event, while POC_{veg} and CaCO₃ 238 exports were mainly observed during the second event. High export fluxes of POC_{fp} were also 239 observed during these two main events (cups #4, #5 and #9). On a seasonal basis, the POC 240 export was largely dominated by fecal pellets (89 %) while the relative contribution of diatoms 241 (vegetative cells and spores) to total POC never exceeded 11%. 242

243

To better understand the seasonal variability of the export via different biological carrier phases 244 we used principal component analysis (PCA) (Fig. 4). The first two principal components (PC) 245 246 explained respectively 72.1 % and 20.8 % of the total variance. The first PC separated the cups in two categories. Positive values of PC1 correspond to cups with material collected during 247 248 both major export events (#3, #4, #5 and #9, #10), and negative values of PC1 were related to 249 the other cups. The highly correlated variables (mass, BSi, POC, PON, POC_{fp}, POC_{diat} and POC_{spore}) (Figure S1) mainly contribute to PC1. The second component (PC2) separated mainly 250 251 the first (negative values, #3, #4, #5) from the second export event (positive values, #9, #10). The variables contributing mainly to PC2 are POC_{spore} associated with the first export event and 252 POC_{veg} and CaCO₃ associated with the second export event. 253

254

255 **3.3 Seasonal changes of TE export**

The 12 TE fluxes varied by almost 7 orders of magnitude (Figure 5). The highest flux was recorded for Al with a maximum of 160 μ mol m⁻² d⁻¹ and the lowest for Th with a minimum of 3.5 x 10⁻² nmol m⁻² d⁻¹. Examination of the seasonal changes revealed two qualitatively different temporal patterns. High export fluxes of P, Cd, Ba, Mo, Cu, Ni and V were associated with one or both of the main export events described in the previous section. For most of these elements, the export fluxes were higher during the first than during the second event, but more subtle differences appear. For example, the flux of Cd export was much more pronounced during the first than the second event, whereas for V the fluxes were almost identical during both events. For the other elements (Y, Mn, Zr, Co, Ti, Cr, Th, Fe and Al) the highest fluxes were measured in the first cups (#1 to #5) and the lowest in the remaining cups (#6 to #12).

266

For TE fluxes, a PCA confirms the partitioning between the two main groups mentioned in the 267 previous section based on the qualitative analysis of the seasonal changes (Figure 6). The two 268 first components of the PCA explain 98.2 % of the seasonal variation of the TE export. For 269 PC1, positive scores correspond to the beginning of the season (#1 to #5) and negative scores 270 271 to the remaining cups (#6 to #12). The highest positive scores for the PC2 are typical of cups corresponding to the two export events (#4 #5 and #9 #10). Within the group of elements 272 273 characterised by two marked peaks of export (P, Cd, Ba, Mo, Cu, Ni and V), the PCA shows 274 three possible subgroups where seasonal variations of these elements are highly correlated. These consist of P, Cd and Ba, then Mo and Cu, and finally Ni and V (Figure S2). 275

276

277 **4 Discussion**

Our parallel observations of the seasonal changes in the export fluxes of different biological carrier phases, as defined hereafter, and of trace metals, provide the opportunity to identify the the main factors that control their export in this iron fertilized region of the Southern Ocean.

The bulk composition of particles is usually partitioned between different pools identified as particulate organic matter (POM), biogenic silica (BSi), calcium carbonate (CaCO₃), lithogenic and authigenic material (Lam et al. 2015). The partitioning of TEs between these different pools

relies on two main hypotheses. First, one assumes that it is possible to identify a chemical 284 285 element or a chemical form of the element that largely dominates one of the pools and has a minor contribution to the others. For the lithogenic fraction, Al has been extensively used, 286 287 although Ti has recently gained interest in this context when the potential source material and its chemical composition are clearly identified. For the POM fraction, beside POC that is widely 288 used, phosphorus (P) is also selected as the reference element, because it is a major contributor 289 and has a mineral form (e.g. apathite) with low abundance in seawater. In addition, P is 290 291 measured simultaneously with metals by analytical methods like Sector Field Induced Coupled Plasma Mass Spectrometry (SF-ICP-MS) or X-ray Fluorescence (XRF) synchrotron (Twining 292 293 et al. 2003).

294

The second assumption is that for any given element, the ratio with the reference element of a 295 296 given fraction must be known or postulated. For the lithogenic pool, the elemental composition of a representative material can be used in order to determine enrichment factors. These 297 298 enrichment factors provide information on the extent to which TE are associated with particles 299 of lithogenic origin. In most studies, global crustal composition or upper crustal compositions (Taylor and McLennan 1995) are used, but the composition of local mineral sources like desert 300 301 dust are also valuable (Kremling and Streu 1993). For the biogenic fractions, CaCO₃ and BSi determinations are straightforward, but there are few experimental data to constrain the ratio of 302 a given TE to CaCO₃ (TE/ CaCO₃) and BSi (TE/BSi) and therefore to derive directly the amount 303 304 of metal transported by these fractions. The issue is even more complicated for POM due to the 305 diverse composition of this fraction. When POM is dominated by phytoplankton, an extension of the Redfield ratio to metals can be considered, but there are large uncertainties in the 306 307 determination of phytoplankton TE/P ratios (Twining and Baines 2013). Moreover, elemental ratios of dead microorganisms can largely differ from those measured in living cells due to the 308

dissolution and remineralization rates that vary between elements. Consequently, TE/P ratios 309 in two important vectors of TE export, phytoplankton aggregates (Twining et al. 2015) or fecal 310 pellets (Fowler 1977) cannot easily be inferred. For example, different types of particulate 311 312 organic matter (Lam et al. 2015) could influence surface adsorption of TEs (Balistrieri et al. 1981) and ultimately the TE stoichiometry. Together, these considerations result in a complex 313 dynamic of TEs hosted in dissolved and particulate pools. This is further complicated by the 314 315 fact that the magnitude of external sources and individual processes are subjected to strong variations throughout the year (Sternberg et al. 2007; Hayes et al. 2015). 316

317

318 In the following, we will discuss our findings from several points of view. First, we will use an approach classically found in the literature and summarised above that provides an estimate of 319 the lithogenic contribution to the TE flux. This approach allows the derivation of the flux not 320 321 supported by lithogenic carriers which can approximate the biological, scavenged, and authigenic contributions. Secondly, we will consider simultaneously several possible carrier 322 323 phases to extract the ones most probably associated with the individual elements. This second 324 approach will be used to investigate further the role of different biological carriers. We will confront these results with recent findings on the biological role of TE in both autotrophic and 325 326 heterotrophic microorganisms, as revealed by laboratory or in situ omics-based studies.

327

328 **4.1 Basalt is the main lithogenic carrier phase.**

Using the results of the PCA for TE (Figure 6), we show that Y, Mn, Cr, Ti, Co, Th, Fe, Al and Zr have a similar seasonal export pattern. Some of these elements are well known as representatives of lithogenic matter (i.e. Ti, Cr, Zr, Y, Th and Al), while others like Mn, Co and Fe are also involved in biological processes. The quantitative estimate of the lithogenic fraction of the fluxes relies on both the choice of a reference element and a reference material of a known

elemental composition. Al and Ti have been both used previously as reference elements. In the 334 335 present study we will not use Al because it is known to be associated with diatom frustules (Ren et al. 2013) and a previous study above the Kerguelen plateau has shown that diatoms dominate 336 337 during spring and summer (Blain et al. 2020). Therefore, Al is likely present in diatoms exported directly via aggregates or indirectly via fecal pellets. Using Al as the reference could 338 therefore lead to an overestimation of the lithogenic fraction, while Ti can provide a more 339 340 conservative estimate. We thus consider Ti as a reference element for the lithogenic fraction (Ohnemus and Lam 2014) and calculated the mass ratios F_{TE}/F_{Ti} where F_{TE} is the export flux of 341 a given TE and F_{Ti} is the export flux of Ti collected in the same cup (Table 2). 342

343

The choice of the elemental ratio is also critical for the calculation of the lithogenic contribution 344 to TE export fluxes. In the present study, for most of the elements associated mainly with a 345 346 lithogenic carrier phase, Al, Fe, Cr, Co, Y, Zr, (Figure 6) the TE/Ti ratios (Table 2) are not significantly different (p=0.01) from the composition of basalt rocks collected around the 347 348 Kerguelen plateau and islands (Kerguelen archipelago, Heard and Mac Donald Islands) (Weis 349 et al. 1993; Yang et al. 1998). However, with the exception of Cr and Y, the TE/Ti ratios measured in the sediment trap differed largely from that typical of upper continental crust 350 (UCC) (Taylor and McLennan 1995) (Table 2). Yet, the large differences for Fe/Ti and Al/Ti 351 ratios resulted very likely from the high Ti content of island basalt (Prytulak and Elliott 2007). 352 We also note that the Mn/Ti ratios are not significantly different from Kerguelen basalt, if a few 353 354 basalt samples with low Ti (ratio<2) are excluded from this analysis. Therefore, derived basalt particles are likely the main contributors to the lithogenic export fluxes, although alteration of 355 rocks and subsequent transformation during transport in terrestrial and marine environments 356 357 could modify the chemical composition of lithogenic particles.

We calculated for individual elements the average TE/Ti based on: i) all cups and ii) only the first two cups and compared with the TE/Ti ratios in the UCC and in the Kerguelen basalt (Table 2). We then estimated the lithogenic contribution to TE export fluxes using equation 3:

362 $F_{\text{TElith}} = (\text{TE}/\text{Ti}) \times F_{\text{Ti}}$ (Eq. 3),

where TE/Ti is the average ratio for the two first cups. In Figure 4, the projection of cups #1 363 and #2 presented the most negative score along PC1 suggesting that TE export fluxes collected 364 365 in this trap were mainly driven by non-biological carriers. Moreover, the PCA of TE export fluxes (Figure 6) shows that the projections of cups#1 and #2 were located in the quarter of 366 space that was related to a suite of TEs typically associated with basalt. This analysis of both 367 PCAs clearly identify these cups as mainly associated with lithogenic material and suggests 368 they are therefore the most appropriate to estimate a lithogenic elemental ratio for the sediment 369 370 trap material. We note that including cups #3, #4 and #5 in the calculation of the individual elemental ratio would have resulted in a biased estimate due to the contribution of biological 371 372 fluxes. We also calculated the residual export flux, which is not associated with lithogenic 373 material for each element using equation 4:

374
$$F_{xs} = F_{TE} - F_{TElith}$$
 (Eq. 4)

and these residual fluxes are represented on Figure 7.

Our observations clearly underscore that the residual export fluxes of 6 elements (Zr, Co, Cr, Th, Fe, Al) estimated using this ratio are occasionally or consistently negative throughout the season (Figure 7). Regarding Fe, the order of magnitude of an expected biogenic flux based on the export flux of P (Pxs) can be estimated. Considering the highest values of P_{xs} of 50 µmol m⁻² d⁻¹ in cup#5, and using a high estimate for the Fe quota (Fe:P=5 mmol mol⁻¹ (Twining and Baines 2013)), one would expect 0.25 µmol m⁻² d⁻¹ of biogenic Fe at the time of the peak flux. This represents around 0.5 % of the total flux of Fe measured (Figure 5), confirming that such low contributions cannot be detected using the calculation of the residual fluxes (Fe_{xs} Eq. 4).
This result, together with the negative values of residual export fluxes, highlight that any
contribution of a carrier phase (e.g. biological) other than basalt derived particles cannot be
detected using this approach.

387

4.2 Role of different biological carriers in the export of TE.

389 In the following, we consider the 9 elements (P, Cd, Ba, Mo, Cu, Ni, V, Y, Mn) for which F_{xs} are positive throughout the season (Figure 7). Among these, 7 elements (P, Mo, Cd, Cu, Ni, V, 390 Mn) have known biological functions and can therefore be directly associated with biological 391 carrier phases. Overall, this is confirmed by the seasonal dynamics of their F_{xs} that presented 1 392 or 2 maxima corresponding to the cups that collected sinking material during the first (cups# 3, 393 4, 5) or second (cups# 9, 10) export event. To go a step forward, we took advantage of the 394 395 detailed description of biological matter export provided by microscopic observations (Rembauville et al. 2015a) in the same cup material. An important aspect was to quantify the 396 397 carbon content of exported diatoms and all types of fecal pellets. Microscopic observations revealed that diatoms dominated the phytoplankton community (Blain et al. 2020), and that 12 398 different taxa contributed significantly (>1% of total biomass) to both the surface carbon 399 biomass and carbon export. However, the concentrations of TE with a biological role certainly 400 varied throughout the season in surface waters due to intense uptake and remineralisation as 401 observed for Fe above the Kerguelen plateau (Blain et al. 2008; Bowie et al. 2015). Similarly, 402 TE quota are likely to vary over time in surface diatoms, with consequences on TE composition 403 of the fecal pellets. The absence of data on seasonal changes in TE concentrations in the water 404 column and the large uncertainty of the TE transfer efficiency between phytoplankton and 405 406 zooplankton led us to make a rather conservative choice of only three biological carriers,

407 vegetative cells, spores and fecal pellets. Additionally, we considered the total particle mass,
408 POC, PON, POC_{diat}, and used CaCO₃ as a tracer of calcifying organisms.

409

410 We first investigated the role of these different biological carriers using PCA (Figure S3) based on F_{xs} and the different biological carriers mentioned above. However, this approach did not 411 prove informative on the association of a given TE with a biological carrier, except for V, which 412 413 was strongly associated with vegetative cells and/or calcifying organisms exported during the second bloom. The strong association of Mn with the first bloom, as revealed by the PCA, is 414 not meaningful, because F_{xs} of Mn is high only in cup #3 albeit the export of this bloom is 415 416 collected by cups #4 and #5 as well (Blain et al. 2020). For Ba and Y, the PCA does not provide any clues on their association with a particular biological carrier. Co-linearity between the 417 different biological descriptors may have hampered the emergence of more significant 418 419 relationships for other elements.

420

421 We have therefore analysed the data set using a different statistical tool, the Partial Least Square 422 Regression (PLSR), also referred to as Projection of Latent Structure Regression (Abdi 2010). This method considers a set of predictors (X) and descriptors (Y) and extracts a single set of 423 424 scores from both simultaneously. The method can be seen as a simultaneous PCA on X and Y which achieves the best relationships between X and Y. The method is efficient even when the 425 variables are possibly correlated and when the number of variables is large compared to the 426 number of observations. This method has been successfully applied to determine the ecological 427 vectors associated with sinking carbon flux (Rembauville et al. 2015), to predict the partitioning 428 of carbon within plankton assemblages based on bio-optical properties (Rembauville et al. 429 2017) or to link biological diversity and carbon fluxes (Guidi et al. 2016). To apply PLSR we 430 considered the total export flux of the 15 elements (descriptors) and the different biological 431

vectors (predictors) mentioned above, and we considered Ti as an overall predictor of lithogenic 432 433 material. It is important to note that with this approach the search for relationships between elements and the lithogenic carrier phase does not require the use of an elemental ratio. To 434 435 summarize the results of the PLSR analysis we present the projections of both descriptors and predictors in a three-dimensional space defined by the three first latent variables (Figure 8) 436 which represent 57.5%, 22.1% and 8.6 % of the covariance, respectively. The three 437 438 corresponding 2D dimensional projections in the latent vectors space are provided in Figure S4. Three different groups of TE emerge from this analysis. 439

440

TEs associated with the lithogenic carrier phase. The PLSR, clearly identifies a group of
TEs (Al, Zr, Cr, Fe, Th, Co, Mn and Y) for which the seasonal dynamics are strongly related to
the lithogenic carrier phase, represented by Ti. This result is in line with the conclusions of the
PCA and TE_{xs} analysis (Figure 6).

445

TEs associated with fecal pellets and diatom spores. The export of Cd, P and Ba was strongly 446 associated with POC_{fp} and to a lesser extend to POC_{sp}. For Ba, this result is not surprising 447 considering that particulate Ba is largely found as authigenic mineral barite in the ocean 448 (Dehairs et al. 1980), formed by precipitation from dissolved Ba in microenvironments where 449 it becomes supersaturated with respect to barite. Such environments are fecal pellets (Alldredge 450 and Cohen 1987; Ploug 2001) or aggregates like marine snow which in our study contained 451 large quantities of spores (Blain et al. 2020). Strong correlations of Cd and P export fluxes have 452 already been observed with sediment traps deployed in the upper water column (> 1500m) 453 whereas this relationship vanished at greater depth (Ho et al. 2011; Conte et al. 2019). In the 454 455 present study, the export of Cd was mainly driven by spores during the first bloom and by fecal pellets throughout the season, while vegetative cells and calcifying organisms present during 456

the second bloom played a minor role for Cd export. Cd, but also Co, can substitute for Zn in 457 458 the carbonic anhydrase (CA) enzyme, Cd-CA and Zn-CA, respectively (Morel et al. 2020). This has been demonstrated for diatoms under low Zn conditions (Lane and Morel 2000). Cd-459 CA is present in Thalassiosira antarctica, Chaetoceros dichaeta, Proboscia alata and 460 Proboscia inermis (Morel et al. 2020), species that were well represented in our sediment traps 461 462 (Blain et al. 2020). Interestingly T. antarctica and C. dichatea are small and spore forming diatoms which dominated during the first bloom, while the genus Proboscia contains large 463 diatoms exported as vegetative cells that thrived during the second bloom. Cd utilisation by 464 different diatoms in surface waters could explain the seasonal variation of Cd export in the 465 466 sediment traps. Cd can also be coincidentally taken up by the divalent transporter under Felimited conditions (Lane et al. 2008; Horner et al. 2013). At the beginning of the season, the 467 reservoir of Zn and Fe was large above the Kerguelen plateau (Wang et al. 2019), but the rapid 468 469 development of the massive bloom of small diatoms could lead to a rapid decrease in Zn to levels at which the substitution of Zn by Cd in CA occurred and/or Cd being taken up by the 470 471 divalent transporter. No strong signal of particulate Cd was associated with the second bloom suggesting that the substitution of Zn by Cd in CA or divalent transport uptake are not dominant 472 processes at the end of the productive season, either due to increased Zn or Fe concentrations 473 474 provided by remineralisation after the first bloom or due to lower requirements of large diatom cells which do not need Cd for CA activities. Although calcifying organisms including 475 coccolithophorids, present during the second bloom, have high Cd requirements (Ho et al. 2003; 476 477 Sunda 2012), their contribution was likely hidden behind the large fluxes associated with fecal pellets. 478

479

TEs associated with lithogenic and biological carrier phases. V, Mo, Cu and Ni export fluxes
are both driven by lithogenic and biological carrier phases. This is a consequence of both their

482 significant contribution to Kerguelen basalt composition (Table 2) and their biological role in 483 microorganisms. Using a similar approach to that used for Cd, we examine the seasonal 484 dynamics of the export of these four metals by first summarizing a few recent insights on their 485 biological role for microorganisms relevant for our study. We then discuss how these 486 observations can provide clues to understand the seasonal dynamics of their export.

The main non-lithogenic V export event coincided with the flux of large vegetative diatoms and 487 488 calcifying organisms after the second bloom (Figure 6). Due to the similar seasonal patterns of these biological carrier phases, it is not possible, based on PLSR, to make a clear preferential 489 association with either of them. The current knowledge on the biological role of V is mainly 490 491 related to diatoms, thus our discussion on the temporal changes of V export focuses on this phytoplankton group. V is a cofactor of haloperoxidase enzymes (VHPO) that produce organo-492 halogens (Moore et al. 1996; Murphy et al. 2000; Hill and Manley 2009). Haloperoxidase 493 494 activity by diatoms could alter the quorum sensing of prokaryotes and therefore protect diatoms against algicidal prokaryotes (Amin et al. 2012). In contrast to the seasonal dynamics of all 495 496 other elements, the export of V associated to the biological fraction was higher during the second than during the first bloom (Figure 7). Seasonal observations of diatom and prokaryotic 497 communities in the surface layer revealed compositional changes and strong associations 498 499 (positive and negative) between diatom species and prokaryotic taxa (Liu et al. 2020). Positive associations could result from interactions based on the exchange of metabolites between 500 diatoms and prokaryotes for resource acquisition, but negative associations are more difficult 501 to interpret. The seasonal dynamics of non-lithogenic particulate V, if related to VHPO activity, 502 503 could suggest that some diatoms efficiently reduce the growth of targeted prokaryotic taxa with 504 algicidal activity in the phycosphere.

The prevalence of non-lithogenic particulate V during the second bloom could be related to 506 seasonal changes of the bioavailability of Fe. Haloperoxidase can contain Fe-heme as prosthetic 507 group instead of V. Fe-heme containing enzymes could dominate the haloperoxidase activity 508 509 of diatoms when the bioavailable Fe stock is high such as at the beginning of the season. However, as biological uptake during the first bloom consumed a large part of the bioavailable 510 Fe, haloperoxidase activity of diatoms dominating during the second bloom may have switched 511 to VHPO, which requires the uptake of vanadate, an anion that is always present at non-limiting 512 513 concentrations in seawater. V can also be found in nitrogenase (*nif*) involved in the fixation of dinitrogen (N₂) where it substitutes Mo. A recent study illustrated that Mo/Fe containing nif 514 genes are overexpressed by prokaryotic communities on marine particles (Debeljak et al. 2021). 515 Therefore, the association of V or Mo with vegetative diatoms could partly be explained by N₂ 516 517 fixing prokaryotes attached to particles and their downward transport could be a biological carrier phase for Mo and V. 518

519

520 The dominant biological carrier phase for Cu was different to that of V and Mo. Cu export was 521 mainly related to diatom spores and to a lesser extend to fecal pellets whereas no clear association with vegetative cells and CaCO₃ was observed (Figure 8 and Figure S4). Cu is a co-522 factor of a large number of oxidative enzymes involved in different metabolic pathways 523 including Fe acquisition (Maldonado and Price 2001) and nitrogen cycling (Kuypers et al. 524 2018). Another noticeable feature of these Cu proteins is that most of them are located outside 525 eukaryotic cells or in the periplasm of prokaryotes (Silva and Williams 2001). A possible 526 527 consequence can be that Cu enzymes are prone to rapid degradation and release of Cu following cell death. Fecal pellets or spores could provide a protected environment during export, which 528 529 could explain our observations.

The biological carrier phases for Ni were mainly diatoms (spores or vegetative cells) and fecal 531 532 pellets had a minor role. Among the many biological pathways, Ni is involved in the assimilation of urea (Oliveira and Antia 1984) and is also the cofactor of an enzyme of the 533 534 superoxide dismutase (SOD) family which can substitute for Fe-superoxide dismutase in low Fe environments (Dupont et al. 2010; Cuvelier et al. 2010). These requirements for Ni likely 535 lead to high Ni quota of diatoms relative to other phytoplankton groups (Twining et al. 2012). 536 537 It was, however, also noted that 50% of Ni contained in diatoms is associated with the frustule with an unknown function. These Ni dependent enzymes suggest that diatom spores, vegetative 538 cells and fecal pellets that contained mainly diatoms are all potential biological vectors of Ni 539 540 export. If true, the lack of a marked difference between the first and the second bloom dominated by spores and vegetative cells, respectively, is surprising. A larger contribution of 541 diatoms to Ni export would be expected during the second bloom for two reasons. First, the 542 543 assimilation of urea is likely only noticeable when the switch from NO_3^- to NH_4^+ uptake has occurred, thus after the first bloom. Second, since Fe bioavailability was lower during the 544 545 second bloom, Fe-SOD is likely to be replaced by Ni-SOD. We suggest an additional process 546 to significantly contribute to the biological export of Ni. Methanogenic Archaea utilise different enzymes belonging to the hydrogenase, reductase or CO dehydrogenase families where Ni is 547 present as co-factor (Mulrooney and Hausinger 2003). Methanogenic Archaea have been 548 detected in different marine particles like marine snow or fecal pellets (Maarel et al. 1999) 549 where they could thrive within anoxic niches (Alldredge and Cohen 1987; Ploug 2001). A time 550 series of the composition of the particulate matter in the surface layers would certainly provide 551 new data required to decipher between these different hypotheses. 552

553

554 **5 Conclusion**

Our observations of the seasonal particulate TE export in a productive region of the Southern 555 556 Ocean have revealed that the identification of the carrier phases is critical for our understanding of the export dynamics of individual TE. The lithogenic and biological carrier phases identified 557 558 in our study had distinct temporal patterns. Basalt particles, the main lithogenic carrier phase dominated the export flux early in the season and strongly decreased over time, reflected in the 559 560 particulate export pattern of TE representative of lithogenic matter (Ti, Cr, Zr, Y, Th and Al) 561 and of TE with a defined biological role (Mn, Co and Fe). The biological carrier phases, diatom vegetative cells and spores, revealed two pulsed export events, while vertical transport via fecal 562 pellets remained stable over time. TE with known biological functions (Cd, Ba, Mo, Cu, Ni and 563 564 V) were associated with one or both of these main export events.

A further look into the seasonal variability of stocks of bioavailable TE is necessary to better understand how these influence the phytoplankton assemblage, inherent enzyme strategies, and subsequent TE utilisation and exports. Finally, future studies should investigate TE composition of individual fecal pellets produced by different zooplankton species feeding on distinct food sources. This could provide insight to help decipher the contribution of each zooplankton species to TE export.

571

572 Acknowledgments, Samples, and Data

We thank the captains and the crew of the R/V Marion Dufresne for their support during the two cruises. We thank E. de Saint Léger, F. Pérault from DT-INSU, and people of IPEV (Institut Polaire Paul Emile Victor) for the technical support during preparation, deployment and recovery of moorings. We thank Nathalie Leblond (Laboratoire Océanographie de Villefranche sur mer) for processing the samples and performing chemical analysis. We thank Mathieu Rembauville for his help during deployment of the clean traps. We thank the anonymous reviewers and the associated editor for their careful reading of the manuscript and for their

580	comments and suggestions that have improved of our manuscript. This work is part of the
581	project SOCLIM supported by the Climate Initiative of the foundation BNP Paribas, the French
582	research program LEFE-CYBER of INSU-CNRS, IPEV, Sorbonne Université, and the Flotte
583	Océanographique Française. This work was also supported by the project SEATRAK funded
584	by the French research program LEFE-CYBER. The authors declare no conflict of interest.
585	Data are available at the SEANOE database https://www.seanoe.org/data/00606/71768/
586	

588 **References**

589

- 590 Abdi, H. 2010. Partial least squares regression and projection on latent structure regression (PLS
- 591 Regression). WIREs Comp Stat **2**: 97–106. doi:10.1002/wics.51
- 592 Alldredge, A. L., and Y. Cohen. 1987. Can Microscale Chemical Patches Persist in the Sea?
- 593 Microelectrode Study of Marine Snow, Fecal Pellets. Science **235**: 689–691.
- 594 doi:10.1126/science.235.4789.689
- Amin, S. A., M. S. Parker, and E. V. Armbrust. 2012. Interactions between Diatoms and Bacteria.
 Microbiology and Molecular Biology Reviews 76: 667–684. doi:10.1128/MMBR.00007-12
- 597 Aminot, A., and R. Kérouel. 2007. Dosage automatique des nutriments dans les eaux marines :
- 598 méthodes en flux continu, Ifremer.
- 599 Anderson, R. F. 2020. GEOTRACES: Accelerating Research on the Marine Biogeochemical Cycles of
- 600 Trace Elements and Their Isotopes. Annu. Rev. Mar. Sci. **12**: 49–85. doi:10.1146/annurev-601 marine-010318-095123
- Antia, A. N., W. Koeve, G. Fischer, and others. 2001. Basin-wide particulate carbon flux in the Atlantic
- 603 Ocean: Regional export patterns and potential for atmospheric CO ₂ sequestration. Global

604 Biogeochem. Cycles **15**: 845–862. doi:10.1029/2000GB001376

Balistrieri, L., P. G. Brewer, and J. W. Murray. 1981. Scavenging residence times of trace metals and

surface chemistry of sinking particles in the deep ocean. Deep Sea Research Part A.

607 Oceanographic Research Papers **28**: 101–121. doi:10.1016/0198-0149(81)90085-6

Blain, S., B. Quéguiner, L. Armand, and others. 2007. Effect of natural iron fertilisation on carbon

sequestration in the Southern Ocean. Nature **446**: 1070–1075. doi:doi:10.1038/nature05700

- 610 Blain, S., M. Rembauville, O. Crispi, and I. Obernosterer. 2020. Synchronized autonomous sampling
- 611 reveals coupled pulses of biomass and export of morphologically different diatoms in the
- 612 Southern Ocean. Limnol Oceanogr Ino.11638. doi:10.1002/Ino.11638

- 613 Blain, S., G. Sarthou, and P. Laan. 2008. Distribution of dissolved iron during the natural iron-
- 614 fertilization experiment KEOPS (Kerguelen Plateau, Southern Ocean). Deep Sea Research Part
 615 II: Topical Studies in Oceanography 55: 594.
- Bowie, A. R., P. van der Merwe, F. Quéroué, and others. 2015. Iron budgets for three distinct
- biogeochemical sites around the Kerguelen Archipelago (Southern Ocean) during the natural
- 618 fertilisation study, KEOPS-2. Biogeosciences **12**: 4421–4445. doi:10.5194/bg-12-4421-2015
- 619 Boyd, P. W., H. Claustre, M. Levy, D. A. Siegel, and T. Weber. 2019. Multi-faceted particle pumps
- 620 drive carbon sequestration in the ocean. Nature **568**: 327–335. doi:10.1038/s41586-019-
- 621 1098-2
- Buesseler, K. O., A. N. Antia, M. Chen, and others. 2007. An assessment of the use of sediment traps
 for estimating upper ocean particle fluxes. j mar res 65: 345–416.
- 624 doi:10.1357/002224007781567621
- 625 Conte, M. H., A. M. Carter, D. A. Koweek, S. Huang, and J. C. Weber. 2019. The elemental
- 626 composition of the deep particle flux in the Sargasso Sea. Chemical Geology **511**: 279–313.
- 627 doi:10.1016/j.chemgeo.2018.11.001
- 628 Cornet-Barthau, V., L. Armand, and B. Quéguiner. 2007. Biovolume and biomass estimates of key
- 629 diatoms in the Southern Ocean. Aquatic Microbial Ecology **48**: 295–308.
- 630 Cuvelier, M. L., A. E. Allen, A. Monier, and others. 2010. Targeted metagenomics and ecology of
- 631 globally important uncultured eukaryotic phytoplankton. Proceedings of the National

632 Academy of Sciences **107**: 14679–14684. doi:10.1073/pnas.1001665107

- 633 Debeljak, P., S. Blain, A. Bowie, P. Merwe, B. Bayer, and I. Obernosterer. 2021. Homeostasis drives
- 634 intense microbial trace metal processing on marine particles. Limnol Oceanogr 66: 3842–
- 635 3855. doi:10.1002/lno.11923
- 636 Dehairs, F., R. Chesselet, and J. Jedwab. 1980. Discrete suspended particles of barite and the barium
- 637 cycle in the open ocean. Earth and Planetary Science Letters **49**: 528–550. doi:10.1016/0012-
- 638 821X(80)90094-1

- 639 Dupont, C. L., K. N. Buck, B. Palenik, and K. Barbeau. 2010. Nickel utilization in phytoplankton
- assemblages from contrasting oceanic regimes. Deep Sea Research Part I: Oceanographic
 Research Papers 57: 553–566. doi:10.1016/j.dsr.2009.12.014
- 642 Fowler, S. W. 1977. Trace elements in zooplankton particulate products. Nature **269**: 51–53.
- 643 doi:10.1038/269051a0
- 644 Gleiber, M., D. Steinberg, and H. Ducklow. 2012. Time series of vertical flux of zooplankton fecal
- 645 pellets on the continental shelf of the western Antarctic Peninsula. Mar. Ecol. Prog. Ser. **471**:
- 646 23–36. doi:10.3354/meps10021
- 647 González, H., and V. Smetacek. 1994. The possible role of the cyclopoid copepod Oithona in retarding
- 648 vertical flux of zooplankton faecal material. Mar. Ecol. Prog. Ser. **113**: 233–246.
- 649 doi:10.3354/meps113233
- Guidi, L., S. Chaffron, L. Bittner, and others. 2016. Plankton networks driving carbon export in the
 oligotrophic ocean. Nature 532: 465–470. doi:10.1038/nature16942
- Hayes, C. T., J. N. Fitzsimmons, E. A. Boyle, D. McGee, R. F. Anderson, R. Weisend, and P. L. Morton.
- 653 2015. Thorium isotopes tracing the iron cycle at the Hawaii Ocean Time-series Station
- 654 ALOHA. Geochimica et Cosmochimica Acta **169**: 1–16. doi:10.1016/j.gca.2015.07.019
- 655 Hill, V. L., and S. L. Manley. 2009. Release of reactive bromine and iodine from diatoms and its
- possible role in halogen transfer in polar and tropical oceans. Limnol. Oceanogr. **54**: 812–822.
- 657 doi:10.4319/lo.2009.54.3.0812
- Hillebrand, H., C.-D. Dürselen, D. Kirschtel, U. Pollingher, and T. Zohary. 1999. Biovolume calculation
- 659 for pelagic and benthic microalgae. Journal of Phycology **35**: 403–424. doi:10.1046/j.1529660 8817.1999.3520403.x
- Ho, T.-Y., W.-C. Chou, H.-L. Lin, and D. D. Sheu. 2011. Trace metal cycling in the deep water of the
- 662 South China Sea: The composition, sources, and fluxes of sinking particles. Limnol. Oceanogr.
- 663 **56**: 1225–1243. doi:10.4319/lo.2011.56.4.1225

- Ho, T.-Y., A. Quigg, Z. V. Finkel, A. J. Milligan, K. Wyman, P. G. Falkowski, and F. M. M. Morel. 2003.
- The elemental composition of some marine phytoplanton. Journal of Phycology **39**: 1145–
 doi:10.1111/j.0022-3646.2003.03-090.x
- 667 Honjo, S., R. Francois, S. Manganini, J. Dymond, and R. Collier. 2000. Particle fluxes to the interior of
- the Southern Ocean in the Western Pacific sector along 170[deg]W. Deep Sea Research Part
 II: Topical Studies in Oceanography 47: 3521.
- 670 Horner, T. J., R. B. Y. Lee, G. M. Henderson, and R. E. M. Rickaby. 2013. Nonspecific uptake and
- 671 homeostasis drive the oceanic cadmium cycle. Proceedings of the National Academy of
- 672 Sciences **110**: 2500–2505. doi:10.1073/pnas.1213857110
- Huang, S., and M. H. Conte. 2009. Source/process apportionment of major and trace elements in
- 674 sinking particles in the Sargasso sea. Geochimica et Cosmochimica Acta **73**: 65–90.
- 675 doi:10.1016/j.gca.2008.08.023
- 676 Kremling, K., and P. Streu. 1993. Saharan dust influenced trace element fluxes in deep North Atlantic
- 677 subtropical waters. Deep Sea Research Part I: Oceanographic Research Papers 40: 1155–
- 678 1168. doi:10.1016/0967-0637(93)90131-L
- 679 Kuss, J., J. J. Waniek, K. Kremling, and D. E. Schulz-Bull. 2010. Seasonality of particle-associated trace
- 680 element fluxes in the deep northeast Atlantic Ocean. Deep Sea Research Part I:
- 681 Oceanographic Research Papers **57**: 785–796. doi:10.1016/j.dsr.2010.04.002
- Kuypers, M. M. M., H. K. Marchant, and B. Kartal. 2018. The microbial nitrogen-cycling network. Nat
 Rev Microbiol 16: 263–276. doi:10.1038/nrmicro.2018.9
- Lam, P. J., D. C. Ohnemus, and M. E. Auro. 2015. Size-fractionated major particle composition and
- 685 concentrations from the US GEOTRACES North Atlantic Zonal Transect. Deep Sea Research
- 686 Part II: Topical Studies in Oceanography **116**: 303–320. doi:10.1016/j.dsr2.2014.11.020
- 687 Lane, E. S., K. Jang, J. T. Cullen, and M. T. Maldonado. 2008. The interaction between inorganic iron
- 688 and cadmium uptake in the marine diatom Thalassiosira oceanica. Limnol. Oceanogr. 53:
- 689 1784–1789. doi:10.4319/lo.2008.53.5.1784

- Lane, T. W., and F. M. M. Morel. 2000. A biological function for cadmium in marine diatoms.
- 691 Proceedings of the National Academy of Sciences **97**: 4627–4631.

692 doi:10.1073/pnas.090091397

- 693 Lemaitre, N., H. Planquette, F. Dehairs, and others. 2020. Particulate Trace Element Export in the
- 694 North Atlantic (GEOTRACES GA01 Transect, GEOVIDE Cruise). ACS Earth Space Chem. 4:
- 695 2185–2204. doi:10.1021/acsearthspacechem.0c00045
- Liu, Y., S. Blain, crispi, olivier, and I. Obernosterer. 2020. Seasonal dynamics of prokaryotes and their
- 697 associations with diatoms in the Southern Ocean as revealed by an autonomous sampler.

698 Environmental Microbiology. doi:10.1111/1462-2920.15184

- Maarel, M. J. E. C., W. Sprenger, R. Haanstra, and L. J. Forney. 1999. Detection of methanogenic
- 700 archaea in seawater particles and the digestive tract of a marine fish species. FEMS
- 701 Microbiology Letters **173**: 189–194. doi:10.1111/j.1574-6968.1999.tb13501.x
- 702 Maldonado, M. T., and N. M. Price. 2001. Reduction and transport of organically bound iron by
- thalassiosira oceanica (bacillariophyceae). Journal of Phycology **37**: 298–310.
- 704 doi:10.1046/j.1529-8817.2001.037002298.x
- 705 McDonnell, A. M. P., P. J. Lam, C. H. Lamborg, and others. 2015. The oceanographic toolbox for the
- collection of sinking and suspended marine particles. Progress in Oceanography **133**: 17–31.
- 707 doi:10.1016/j.pocean.2015.01.007
- 708 Menden-Deuer, S., and E. J. Lessard. 2000. Carbon to volume relationships for dinoflagellates,

709 diatoms, and other protist plankton. Limnol. Oceanogr. **45**: 569–579.

- 710 doi:10.4319/lo.2000.45.3.0569
- 711 Moore, R. M., M. Webb, R. Tokarczyk, and R. Wever. 1996. Bromoperoxidase and iodoperoxidase
- 712 enzymes and production of halogenated methanes in marine diatom cultures. J. Geophys.
- 713 Res. **101**: 20899–20908. doi:10.1029/96JC01248

- Morel, F. M. M., P. J. Lam, and M. A. Saito. 2020. Trace Metal Substitution in Marine Phytoplankton.
- 715 Annu. Rev. Earth Planet. Sci. 48: annurev-earth-053018-060108. doi:10.1146/annurev-earth716 053018-060108
- Mulrooney, S. B., and R. P. Hausinger. 2003. Nickel uptake and utilization by microorganisms. FEMS
 Microbiol Rev 27: 239–261. doi:10.1016/S0168-6445(03)00042-1
- Murphy, C. D., R. M. Moore, and R. L. White. 2000. Peroxidases from marine microalgae. Journal of
 Applied Phycology 12: 507–513. doi:10.1023/A:1008154231462
- 721 Ohnemus, D. C., and P. J. Lam. 2014. Cycling of lithogenic marine particles in the US GEOTRACES
- 722 North Atlantic transect. Deep Sea Research Part II: Topical Studies in Oceanography.
- 723 doi:10.1016/j.dsr2.2014.11.019
- 724 Oliveira, L., and N. J. Antia. 1984. Evidence of nickel ion requirement for autotrophic growth of a
- marine diatom with urea serving as nitrogen source. British Phycological Journal 19: 125–
 134. doi:10.1080/00071618400650131
- 727 Planquette, H., and R. M. Sherrell. 2012. Sampling for particulate trace element determination using
- water sampling bottles: methodology and comparison to in situ pumps. Limnol. Oceanogr.
- 729 Methods **10**: 367–388. doi:10.4319/lom.2012.10.367
- Ploug, H. 2001. Small-scale oxygen fluxes and remineralization in sinking aggregates. Limnol.
- 731 Oceanogr. **46**: 1624–1631. doi:10.4319/lo.2001.46.7.1624
- 732 Price, N. M., G. I. Harrison, J. G. Hering, R. J. Hudson, P. M. Nirel, B. Palenik, and F. M. Morel. 1989.

733 Preparation and chemistry of the artificial algal culture medium Aquil. Biological

- 734 Oceanography **6**: 443–461.
- 735 Prytulak, J., and T. Elliott. 2007. TiO2 enrichment in ocean island basalts. Earth and Planetary Science
- 736 Letters **263**: 388–403. doi:10.1016/j.epsl.2007.09.015
- Pullwer, J., and J. J. Waniek. 2020. Particulate trace metal fluxes in the center of an oceanic desert:
- 738 Northeast Atlantic subtropical gyre. Journal of Marine Systems **212**: 103447.
- 739 doi:10.1016/j.jmarsys.2020.103447

740	Ragueneau, O., N. Savoye, Y. Del Amo, J. Cotten, B. Tardiveau, and A. Leynaert. 2005. A new method
741	for the measurement of biogenic silica in suspended matter of coastal waters: using Si:Al
742	ratios to correct for the mineral interference. Continental Shelf Research 25 : 697–710.
743	doi:10.1016/j.csr.2004.09.017
744	Rembauville, M., S. Blain, L. Armand, B. Quéguiner, and I. Salter. 2015a. Export fluxes in a naturally
745	iron-fertilized area of the Southern Ocean – Part 2: Importance of diatom resting spores and
746	faecal pellets for export. Biogeosciences 12 : 3171–3195. doi:10.5194/bg-12-3171-2015
747	Rembauville, M., N. Briggs, M. Ardyna, and others. 2017. Plankton Assemblage Estimated with BGC-
748	Argo Floats in the Southern Ocean: Implications for Seasonal Successions and Particle Export:

749 Journal of Geophysical Research: Oceans **122**: 8278–8292. doi:10.1002/2017JC013067

750 Rembauville, M., J. Meilland, P. Ziveri, R. Schiebel, S. Blain, and I. Salter. 2016. Planktic foraminifer

and coccolith contribution to carbonate export fluxes over the central Kerguelen Plateau.

752 Deep Sea Research Part I: Oceanographic Research Papers **111**: 91–101.

753 doi:10.1016/j.dsr.2016.02.017

754 Rembauville, M., I. Salter, N. Leblond, A. Gueneugues, and S. Blain. 2015b. Export fluxes in a naturally

iron-fertilized area of the Southern Ocean – Part 1: Seasonal dynamics of particulate organic

carbon export from a moored sediment trap. Biogeosciences **12**: 3153–3170.

757 doi:10.5194/bg-12-3153-2015

758 Ren, H., B. G. Brunelle, D. M. Sigman, and R. S. Robinson. 2013. Diagenetic aluminum uptake into

759 diatom frustules and the preservation of diatom-bound organic nitrogen. Marine Chemistry

760 **155**: 92–101. doi:10.1016/j.marchem.2013.05.016

Silva, J. J. R. F. da, and R. J. P. Williams. 2001. The biological chemistry of the elements: the inorganic
 chemistry of life, 2nd ed. Oxford University Press.

763 Sternberg, E., C. Jeandel, J.-C. Miquel, B. Gasser, M. Souhaut, R. Arraes-Mescoff, and R. Francois.

764 2007. Particulate barium fluxes and export production in the northwestern Mediterranean.

765 Marine Chemistry **105**: 281–295. doi:10.1016/j.marchem.2007.03.003

- 766 Sun, W.-P., Z.-B. Han, C.-Y. Hu, and J.-M. Pan. 2016. Source composition and seasonal variation of
- particulate trace element fluxes in Prydz Bay, East Antarctica. Chemosphere 147: 318–327.
 doi:10.1016/j.chemosphere.2015.12.105
- 769 Sunda, W. G. 2012. Feedback Interactions between Trace Metal Nutrients and Phytoplankton in the
- 770 Ocean. Frontiers in Microbiology **3**. doi:10.3389/fmicb.2012.00204
- Taylor, S. R., and S. M. McLennan. 1995. The geochemical evolution of the continental crust. Rev.
- 772 Geophys. **33**: 241. doi:10.1029/95RG00262
- 773 Tonnard, M., H. Planquette, A. R. Bowie, and others. 2020. Dissolved iron in the North Atlantic Ocean
- and Labrador Sea along the GEOVIDE section (GEOTRACES section GA01). Biogeosciences **17**:
- 775 917–943. doi:10.5194/bg-17-917-2020
- Twining, B. S., and S. B. Baines. 2013. The Trace Metal Composition of Marine Phytoplankton. Annual
 Review of Marine Science 5: 191–215. doi:10.1146/annurev-marine-121211-172322
- 778 Twining, B. S., S. B. Baines, N. S. Fisher, J. Maser, S. Vogt, C. Jacobsen, A. Tovar-Sanchez, and S. A.
- 779 Sañudo-Wilhelmy. 2003. Quantifying Trace Elements in Individual Aquatic Protist Cells with a
- 780 Synchrotron X-ray Fluorescence Microprobe. Anal. Chem. **75**: 3806–3816.
- 781 doi:10.1021/ac034227z
- 782 Twining, B. S., S. B. Baines, S. Vogt, and D. M. Nelson. 2012. Role of diatoms in nickel biogeochemistry
- 783 in the ocean: DIATOMS AND NICKEL BIOGEOCHEMISTRY. Global Biogeochem. Cycles 26: n/a-
- 784 n/a. doi:10.1029/2011GB004233
- 785 Twining, B. S., S. Rauschenberg, P. L. Morton, and S. Vogt. 2015. Metal contents of phytoplankton
- and labile particulate material in the North Atlantic Ocean. Progress in Oceanography **137**:
- 787 261–283. doi:10.1016/j.pocean.2015.07.001
- 788 Wang, R.-M., C. Archer, A. R. Bowie, and D. Vance. 2019. Zinc and nickel isotopes in seawater from
- the Indian Sector of the Southern Ocean: The impact of natural iron fertilization versus
- 790 Southern Ocean hydrography and biogeochemistry. Chemical Geology **511**: 452–464.
- 791 doi:10.1016/j.chemgeo.2018.09.010

- 792 Weis, D., F. A. Frey, H. Leyrit, and I. Gautier. 1993. Kerguelen Archipelago revisited: geochemical and
- isotopic study of the Southeast Province lavas. Earth and Planetary Science Letters **118**: 101–

794 119. doi:10.1016/0012-821X(93)90162-3

- Yang, H.-J., F. A. Frey, D. Weis, A. Giret, D. Pyle, and G. Michon. 1998. Petrogenesis of the Flood
- 796 Basalts Forming the Northern Kerguelen Archipelago: Implications for the Kerguelen Plume.
- 797 Journal of Petrology **39**: 711–748. doi:10.1093/petroj/39.4.711

798

Cup number	Opening date	Closing date
1	20/10/2016	01/11/2016
2	01/11/2016	12/11/2016
3	12/11/2016	23/11/2016
4	23/11/2016	04/12/2016
5	04/12/2016	15/12/2016
6	15/12/2016	26/12/2016
7	26/12/2016	06/01/2017
8	06/01/2017	17/01/2017
9	17/01/2017	28/01/2017
10	28/01/2017	08/02/2017
11	08/02/2017	19/02/2017
12	19/02/2017	01/03/2017

Table 1 : Sampling dates for sediment trap

Table 2 : **TE/Ti ratio.** In the first column, the star associated with a given trace element indicates that the mean TE/Ti in Kerguelen basalt and in sediment trap matter is not statistically different (p=0.01). ^(a) mean elemental ratios in continental crust (CC) from (Taylor and McLennan 1995). ^(b) Mean elemental ratios derived from composition of basalt rocks of Kerguelen island (Weis et al. 1993; Yang et al. 1998). ^(c) Mean elemental ratios considering all cups. ^(d) Ratio elemental ratios considering the two first cups.

	Upper CC ^(a)	Basalt ^(b)	cups#1-12 ^(c)	cups#1-2 ^(d)
	TE/Ti	TE/Ti	TE/Ti	TE/Ti
TE	(g/g)	(g/g)	(g/g)	(g/g)
Ti	1	1	1	1
Al*	26.7	4.2 ± 1.6	5.3 ± 0.7	6.4
Fe*	11.7	5.0 ± 1.7	4.2 ± 0.7	4.4
Mn	0.2	(8.6 ± 2.4) 10 ⁻²	(6.3± 0.8) 10 ⁻²	5.2 10 ⁻²
Р	0.23	(2.5 ± 1.0) 10 ⁻⁵	6.9 ± 4.9	0.68
Ва	0.18	(1.6 ± 0.9) 10 ⁻³	0.9 ± 0.6	0.18
V	2 10 ⁻²	(1.4 ± 0.3) 10 ⁻²	(2.9 ± 1.8) 10 ⁻²	1.1 10-2
Cr*	1.2 10-2	(0.7 ± 1.4) 10 ⁻²	(1.2 ± 0.2) 10 ⁻²	1.1 10-2
Co*	3.3 10 ⁻³	(2.4 ± 1.2) 10 ⁻³	(1.8 ± 0.2) 10 ⁻³	1.7 10 ⁻³
Ni	6.6 10 ⁻³	(4.9 ± 7.8) 10 ⁻³	(2.7 ± 1.5) 10 ⁻²	1.1 10 ⁻²
Υ*	7.3 10 ⁻³	(1.8 ± 0.6) 10 ⁻³	(2.3 ± 0.6) 10 ⁻³	1.5 10 ⁻³
Zr*	6.3 10 ⁻²	(1.4 ± 0.7) 10 ⁻²	(6.9 ± 1.0) 10 ⁻²	6.8 10 ⁻²
Th	3.5 10 ⁻³	(1.7 ± 1.2) 10 ⁻⁴	(5.8 ± 1.2) 10 ⁻⁵	7.0 10 ⁻⁴

800 Figure captions:

801

Figure 1: Kerguelen plateau bloom. a) Monthly composite of chlorophyll surface concentration (mg m⁻³) for November 2016. The white dot denotes the location of the sediment trap mooring. b) Seasonal variation of chlorophyll (mg m⁻³). The blue line corresponds to the 8-day composite chlorophyll concentrations over the season of sediment trap deployment. The green line and light green area represent the climatology and standard deviation respectively. The white rectangles along the x axis denote the 12 periods of sediment trap collection.

Figure 2: Physical environment of the sediment trap. For all panels, the grey line shows the raw data acquired every 30 minutes. The black line denotes the running average with a time window of 26 hours. A) depth of the sediment trap, B) Inclination angle of the sediment trap (vertical reference = 0°), C) current speed measured 3 m below the sediment trap. D) Current direction and intensity.

Figure 3: Export fluxes of biological vectors. Each panel shows the seasonal variations of the export flux of the parameter indicated in the upper left corner. Within each panel, vertical bars represent the export fluxes determined in the 12 cups. POC_{diat} is the flux associated with diatoms, POC_{spore} is the flux associated with diatom spores, POC_{veg} is flux associated with diatom vegetative cells and POC_{fp} is the flux associated with fecal pellets.

Figure 4: PCA correlation biplots of biological fluxes. Black dots denote the cups associated with their labels from 1 to 12 (1 corresponds to the first cup collected). Blue arrows represent the projection of the descriptors into the two first principal component plan (for clarity their lengths were multiplied by 2). The definition of arrow labels are C_{tot} (flux of total POC), N_{tot} (flux of total PON), C_{veg} (flux of POC associated with vegetative diatoms), C_{spore} (flux of POC from diatom spores), C_{fp} (flux of POC from fecal pellets), $CaCO_3$ (flux of CaCO₃), m (flux of total particle mass) and BSi (flux of biogenic silica). Figure 5: Export fluxes of phosphorus and 11 trace elements. Each individual panel shows the seasonal variability of the export flux of the element with its unit indicated in the left upper corner. Within each panel, vertical bars represent the export fluxes collected in the 12 cups and the vertical lines show the standard deviation based of analytical precision.

Figure 6: PCA correlation biplots of trace elements: Black dots denote the cups associated with their labels from 1 to 12 (1 corresponds to the first cup collected). Blue arrows represent the projection of the descriptors into the two first principal component plan (for clarity their lengths were multiplied by 2).

Figure 7: Residual export of trace elements. Each individual panel shows the seasonal variability of the residual export flux F_{xs} (see text for definition) of the element with its unit indicated in the left upper corner.

Figure 8: Partitioning of total trace element export fluxes between different carrier
phases. The plot presents the projections of both predictors (in black) and descriptors (in blue)
in a 3-dimensional space formed by the 3 first latent variables resulting from PLSR analysis
which explained 57.5%, 22.1% and 8.6% of the covariance.















Figure 4







-1

0 PC1 (70.6 %)

1

٧i

۰Y

2

Mn,Zr, Cr Ti, Co, Th Fe

-2 -

-2

Figure 7



Figure 8

