

A plant trait-based response-and-effect framework to assess vineyard inter-row soil management

E Kazakou, G Fried, J Richarte, O Gimenez, C Violle, A Metay

▶ To cite this version:

E Kazakou, G Fried, J Richarte, O Gimenez, C Violle, et al.. A plant trait-based response-and-effect framework to assess vineyard inter-row soil management. Botany Letters, 2016, 163, pp.373 - 388. 10.1080/23818107.2016.1232205. hal-03499357

HAL Id: hal-03499357 https://hal.science/hal-03499357

Submitted on 25 Dec 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1	A plant trait-based response-and-effect framework to assess vineyard inter-row soil					
2	management					
3 4 5 6 7	Kazakou, E. ^{1*} , Fried, G. ² , Violle C. ³ , Gimenez O ³ ., Richarte J. ¹ , Metay A. ⁴					
8 9	¹ Montpellier SupAgro, Centre d'Ecologie Fonctionnelle et Evolutive (UMR 5175), CNRS -					
10	Université de Montpellier - Université Paul-Valéry Montpellier - EPHE, campus CNRS, 19					
11	route de Mende, 34293 Montpellier cedex 5, France					
12	² Anses , Laboratoire de la Santé des Végétaux, Unité Entomologie et Plantes invasives,					
13	CBGP- Campus International de Baillarguet, CS 30016 F-34988 Montferrier-sur-Lez cedex,					
14	France					
15	³ Centre d'Ecologie Fonctionnelle et Evolutive (UMR 5175), CNRS - Université de					
16	Montpellier - Université Paul-Valéry Montpellier - EPHE, campus CNRS, 1919 route de					
17	Mende, 34293 Montpellier cedex 5, France					
18	⁴ Montpellier SupAgro, UMR System (UMR 1230), 2 place Viala, 34060 Montpellier cedex 2,					
19	France					
20						
21	*Corresponding author: tel: +33 (0) 467613334; e-mail: elena.kazakou@cefe.cnrs.fr					
22						
23 24	Running title: Responses and effects of weed communities in vineyards inter-row soil management					

26	Authors' bibliography
27	Elena Kazakou is a plant functional ecologist and an associate professor at Montpellier
28	SupAgro and at the Centre d'Ecologie Fonctionelle et Evolutive (UMR CEFE 5175).
29	Contribution: wrote the manuscript, performed the floristic survey, designed statistical
30	analyses.
31	Guillaume Fried is a weed ecologist working at the Plant Health Laboratory of Anses
32	(French Agency for Food, Environmental and Occupational Health & Safety).
33	Contribution: performed the floristic survey, helped in designing the analysis and reviewed
34	the manuscript.
35	Cyrille Violle is a plant ecologist working at the Centre d'Ecologie Fonctionelle et Evolutive
36	(UMR CEFE 5175). Contribution: helped with the statistical analyses and reviewed the
37	manuscript.
38	Olivier Gimenez is a biostatistician at the Centre d'Ecologie Fonctionelle et Evolutive (UMR
39	CEFE 5175). Contribution: helped with the statistical analysis and reviewed the manuscript.
40	Jean Richarte is a technician, expert in botany working in Montpellier SupAgro.
41	Contribution: performed the floristic survey.
42	
43	Aurélie Metay is an agronomist scientist and an associate professor at Montpellier SupAgro.
44	Contribution: performed the agronomic analysis and reviewed the manuscript.

46 Abstract

47 Biodiversity impacts ecosystem properties and ecosystem services provided by those 48 ecosystems. In this respect, weeds have an important role in maintaining field biodiversity, 49 when balanced with their potential negative impact on crop yield and quality. Trait-based 50 approaches, originally developed in the field of comparative ecology, allow describing weed 51 species responses to management practices in annual crop systems. Here, we aimed at 52 extending the trait-based approach to the spontaneous vegetation of perennial cropping 53 systems, such as vineyards. We adapted the response-effect framework for perennial cultures 54 and especially the vegetation for different soil management practices (tillage, cover crops, 55 spontaneous vegetation) considered as alternatives to the use of herbicides. Soil management 56 practices act as environmental filters that determine the composition and structure of 57 vegetation, which, in turn, modify grapevine growth conditions in the vineyard. We tested this 58 approach in a Mediterranean vineyard where we characterized during 2 years the responses of 59 different components of weed communities (specific and functional composition and 60 dynamics) in three inter-row management practices (tillage, cover crops and mowing 61 spontaneous vegetation) and their effects on several grapevine processes (vine yield, vine leaf 62 water potential and assimilable nitrogen in must). 63

⁶⁵ Keywords: cover crop, functional traits, tillage, vineyard, weed communities

66 Introduction

67 Biodiversity impacts ecosystem properties and ecosystem services provided by those 68 ecosystems (Cardinale et al. 2012; Diaz and Cabido 1997; Chapin et al. 2000; Tilman et al. 69 1997). The underlying hypothesis is that species differ in the efficiency of resource use 70 (resource use complementarity) and/or in how they modify the surrounding environment in 71 ways that impact other species (Wood et al. 2015; Cardinale et al. 2012). As a result, 72 improving plant diversity (specific or genetic) in agricultural systems has been a key issue in 73 agriculture over recent years (Litrico and Violle 2015; Tilman, Wedin, and Knops 1996; 74 Cardinale et al. 2012). In this context, weeds have an important role in maintaining field 75 biodiversity, when it is balanced with their potential negative impact on crop yield and quality 76 (Storkey 2006). More precisely, weeds are a major problem in crop production either through 77 competing for resources or reducing crop quality (Naylor and Lutman 2002). At the same time, weeds can in some cases provide additional services to provisioning service of yield 78 79 (such as pollination, limitation of soil erosion, "traps" for disease agents etc.). The use of key 80 ecological concepts is an important requirement for quantifying the positive contribution of 81 weeds to maintaining biodiversity and ecosystem services without compromising yield must 82 therefore be set in the context of these dis-services associated with weeds (Storkey 2006). 83 Related to this, to manage co-occurring to cultivated plant diversity (weeds, spontaneous 84 vegetation or cover crops) it seems essential to identify assembly rules of complex weed 85 communities and to recognize groups of species that similarly respond to a set of management 86 practices, interact with biotic and biotic components of the agro-system or affect the 87 ecosystem functioning (Navas 2012). 88 Several studies have focused mostly on the taxonomic characterization of the composition and

89 the structure of weed communities occurring in cropping systems to predict their response to 90 agricultural management practices (Fried, Norton, and Reboud 2008; Storkey and Westbury

2007). However, functional trait diversity, rather than the diversity of species *per se*, is a facet
of biodiversity most directly related to species responses on different management practices
and thus consequences to ecosystem functioning (Naeem and Wright 2003; Cadotte,
Carscadden, and Mirotchnick 2011). Trait-based approaches, originally developed in the field
of comparative ecology, have permitted to describe weed species responses to management
practices in annual crop systems (Booth and Swanton, 2002; Gaba, Fried, Kazakou, Chauvel
and Navas 2013; Garnier and Navas 2011; Gunton, Petit and Gaba 2011).

98 The aim of this paper is to extend the trait-based approach to the spontaneous 99 vegetation of perennial cropping systems, namely vineyards. Only a few studies have dealt 100 with the taxonomic characterization of weed communities in those systems (Gago, Cabaleiro 101 and Garcia 2007; Monteiro and Lopes 2007; Steenwerth, Calderón-Orellana, Hanifin, Storm 102 and Mcelrone 2016; Tesic, Keller and Hutton 2007) and none of them have used the trait-103 based approach. According to the response-effect framework (Fig. 1a; Lavorel and Garnier 104 2002; Lavorel, 2013; Suding et al., 2008), environmental drivers act as filters sorting species 105 according to the value of traits (response traits), which results in a functional structure of 106 communities impacting ecosystem functioning (effect traits). The ultimate goal of response-107 effect analyses should be the formulation of parsimonious quantitative relationships 108 expressing the different components of each ecosystem function in relation to particular traits 109 (Lavorel and Garnier 2002). These relationships would make it possible to use traits to scale 110 from individual plants and the communities they form to the ecosystem level (Dawson and 111 Chapin 1993). So, to apply this framework, it is important to use arguments about scaling 112 through the community level by integrating two components: (1) how a community responds 113 to changes, and (2) how that modified community affects ecosystem processes (Suding et al. 114 2008). This response-effect framework has been tested and successfully applied in grasslands 115 to understand how the different levels of land-use impact ecosystem services through the

changes in vegetation but not yet tested in cultivated systems (Gross et al. 2008; Minden andKleyer 2011).

118 Moreover, this approach provides a mechanistic understanding of the linkages 119 between biodiversity and ecosystem functioning (Renting et al. 2009). As some trait values 120 vary with environmental conditions and agricultural management practices (response traits) 121 and can affect ecosystem functioning (effect traits), this framework could be used to develop 122 particular trait-based management strategies that can be implemented in farming systems to 123 increase multiple ecosystem services as well as to manage trade-offs among ecosystem 124 services in agriculture (Wood et al. 2015). However, most research focused on using traits to 125 understand how biodiversity in agricultural systems responds to management practices, rather 126 than on understanding how biodiversity impacts agroecosystem services. In this paper, we 127 adapt the response-effect framework for perennial cultures and more specifically the 128 vegetation under different soil management practices (tillage, cover crops, spontaneous 129 vegetation) considered as alternatives to the use of herbicides between rows. We hypothesized 130 that soil management practices act as environmental filters, which determine the composition 131 and structure of vegetation, which in turn modify grapevine growth conditions in the vineyard 132 (Fig. 1b).

Next, we tested this approach in a Mediterranean vineyard where we characterized
during 2 years the responses of different components of weed communities (taxonomic and
functional composition and dynamics) in three inter-row management practices (tillage, cover
crops and managed spontaneous vegetation) and their effects on several grapevine processes
(vine yield, vine leaf water potential and assimilable nitrogen in must).

138

139 Insert Figure 1, here

141 Vineyard management inter-row practices as a gradient of disturbance and competition

142 intensity

143 Before the 1970s, vegetation between vine rows was traditionally managed by mechanical 144 weeding based on soil tillage. Following the generalized use of chemical weed control and the 145 disappearance of tillage, important shifts in weed community composition were observed 146 between the 1970s and the 1990s (Barralis, Cloquemin and Guérin, 1983; Maillet, 1980). 147 However, health and environmental concerns about the impact of chemicals and deep tillage 148 have recently caused changes in weed management practices (Monteiro and Moreira, 2004; 149 Moreira, 1994). Currently, grape growers can choose between two main weed control 150 methods, alternative to chemical control, used exclusively or in combination: reduced tillage 151 in inter-rows and/or the use of a plant cover (temporary or permanent, spontaneous or sown, 152 in rows or inter-rows) (Gago, Cabaleiro, and García 2007). The practice of cover cropping is 153 currently increasing in vineyards as it provides various ecosystem services in relation to the 154 soil (erosion), the crop (control of vegetative development, and the resulting conditions of 155 yield formation and disease development) and the environment (limited use of pesticides as 156 herbicides, or fungicides) (see Ripoche et al. 2011 for related references). But introducing a 157 second crop can lead to undesirable competition for soil resources such as water and nitrogen 158 (Celette, Gaudin, and Gary 2008) and result in a problem of trade-off between provisioning 159 and regulating ecosystem services. For this reason, vineyards represent a relevant model, in 160 which inter-row soil management ranges from regular soil cultivation (tillage), as in annual 161 crops, to mowing spontaneous vegetation, as in semi-natural permanent grasslands.

Gaba et al. (2013) proposed a comparative description of the environmental gradients created mainly by annual cropping system. Here, we represent three inter-row vineyard management practices (tillage, cover crop and spontaneous vegetation) as a double gradient of soil disturbance and competition intensity with tillage being the most disturbed habitat based

166 on Grime's theory (Grime 1979). Tillage corresponds to high disturbance, defined according 167 to White & Pickett (1985) as "any relatively discrete event in time that disrupts ecosystem, 168 community, or population structure and changes resources, substrate availability or the 169 physical environment" (Fig. 2a). The outcome of tillage varies with respect to both 170 characteristics of the operation, such as depth, number of passes and the characteristics of the 171 soil that is being tilled (see Gaba et al. 2013 for a detailed description). At low levels of 172 disturbance strong competitors exclude competitively inferior species and communities are 173 dominated by a few species. Intermediate levels of disturbance, however, disrupt competitive 174 hierarchies by increasing levels of mortality and thus making free space available for 175 recruitment of competitively inferior species (Connell 1978). These patterns are also affected 176 by spatio-temporal variability in disturbance: how often a disturbance occurs (i.e. frequency), 177 how large the disturbance is (i.e. area or extent) and time since the last disturbance (i.e. time). 178 When the extent of disturbance is considered, areas that are too large will eliminate all 179 species, areas that are too small will have little or no impact, whereas intermediate areas may 180 disrupt competitive exclusion and allow establishment of new species in the disturbed patches 181 Wilson (1994).

182 Studies in annual crops systems have described tillage as a filter that influences weed 183 species composition and weed seed distribution in the soil seed bank (for exemple Cardina, 184 Herms and Doohan, 1991; Cousens and Moss 1990). According to Grime's theory, tillage 185 treatment will result in less diverse communities dominated by a small number of species 186 whereas an intermediate disturbance (corresponding to vegetation cover treatment frequently 187 mowed) will result in more diverse and equitable communities (Grime 2006). Under low 188 disturbance, corresponding to the spontaneous cover, competitive exclusion by the dominant 189 species is expected to occur (Grime 2006), due to light, nutrient or water competition (Fig. 190 2a). At low intensities or frequencies of disturbance there is a balance between competitive

191 exclusion and loss of competitive dominant species by disturbance. As indicated in Fig. 2a,

192 we assumed that the intensity of soil tillage (from no tillage treatments to conventional tillage

193 treatments) corresponds to a disturbance gradient, whereas the degree of vegetation cover,

194 from spontaneous vegetation (with varying and irregular cover according to the weed species)
195 to sowed cover crop (with the sowed species being well implanted and thus dominant species)

196 corresponds to a competition intensity gradient (for light, nutrient or soil water content).

- 197
- 198 Insert Figure 2 here
- 199

200 Effects of different inter-row management practices on communities' composition and

201 structure and ecosystem functioning

202 The effects of different inter-row soil management practices in vineyards on communities' 203 composition and structure and ecosystem processes have been discussed in the literature 204 (synthesis in Table 1) (we focus on non chemical management methods). Conflicting results 205 were found considering the effects of tillage on plant diversity and composition, possibly as a 206 result of community fluctuations rather than deterministic changes in community composition 207 (for example Derksen, Lafond, Thomas, Loeppky and Swanton, 1993; Legere, Stevenson, & 208 Vanasse, 2011; Gago, Cabaleiro, and García 2007 found that plant cover crop showed a much 209 lower number of weed species, mostly therophytes (Beuret and Neury 1990; Maillet 1980) 210 irrespective to soil management technique). A limitation of using plant cover (spontaneous or 211 sown) is the competition for resources including water, soil nutrients and light, which can 212 compromise vineyard vigor. However recent studies on intercropping in vineyards have 213 shown that in some situations, water stress may not be greater than in bare soil vineyards 214 (Celette, Gaudin, and Gary 2008). The advantage of some cover crops is the possibility to use 215 them to manage weeds through several mecanisms. First, competition between weeds and cover crops for light and soil resources will occur to varying degrees based on the vineyard
environment and management (Fredrikson, Skinkis, and Peachey 2011). Second, allelopathic
suppression of weeds has been observed upon decomposition of legume residues, such as
clovers (Dyck and Liebman 1994, Liebman and Davis 2000) and non leguminous residues,
such as cereal rye (Weston 1996).

On the other hand spontaneous cover treatment can be chosen as a costless tradeoff for the winegrowers between improving soil properties, limiting mechanical work and maintaining vine production. According to the objectives of winegrowers (limiting weed competition when resources are limiting or improving soil structure and improving soil bearing capacity), the management of the spontaneous treatment should be adapted each year in agreement with the climatic constraints: for example, the number and dates of mowing could be adapted and the spontaneous treatment may be retained or not.

228

Insert Table 1 here

230 Functional structure and species response traits to different management practices and

231 ressources availability

The functional characterization of weed responses to specific components of management hasbeen successfully developed in several recent studies (Booth and Swanton, 2002; Fried,

Kazakou and Gaba 2012; Gunton et al. 2011; Navas 2012; Storkey 2004; 2006) showing that

for annual crop species, the phenological traits of weeds (timing of emergence) are one of the

key drivers of weed community assembly in responses to crop sowing dates and harvest dates.

- 237 However, in the case of perennial cropping systems only taxonomical diversity indices of
- 238 weed species have been used to compare the effect of different management practices so far
- 239 (e.g. Barralis et al., 1983; Dastgheib and Frampton, 2000 for vineyards) while functional traits
- 240 were only used very recently to assess the services of sown cover crop (Gamour et al. 2015).

241 According to the response-effect framework, response traits to environmental filters 242 change community's functional structure and diversity and thus impact ecosystem processes 243 via changes in the representation of ecosystem-effect traits (Suding et al. 2008; Fig. 1a). It has 244 been assumed that the most abundant species are often more functionally important simply 245 due to greater representation (Grime 1998; Garnier et al. 2004; Balvanera, Kremen, and 246 Martinez-Ramos 2005). This assumption forms the basis of the dominance hypothesis 247 proposed by Grime (1998) under the name of the mass-ratio hypothesis, suggesting that 248 community effects on ecosystem functioning are mainly determined by the traits of the 249 dominant species. Garnier et al. (2004) described the functional structure of a community 250 through value and range of traits by an estimator known as the Community Weighted Mean 251 (CWM). CWM represents the average trait value for a unit of biomass within a community. In 252 other cases, species functional effects may not scale as directly with abundance due to non-253 additive interactions as suggested by the niche complementarity hypothesis (Petchey and 254 Gaston 2006). According to this hypothesis, environmental filtering may affect functional 255 trait diversity (i.e. trait convergence or divergence) within communities through mechanisms 256 such as complementarity resource use (Petchey and Gaston 2006). 257 In Table 2 we provide a summary of the response of traits according to different inter-row 258 soil management and the corresponding sources. Based on previous results in other 259 environments (for a review see Garnier, Navas and Grigulis 2016), highly disturbed habitats 260 select species with rapid completion of the life cycle and high fecundity (low seed mass, onset 261 of flowering at the end of favorable season, low reproductive height), corresponding to the 262 ruderal species strategy (R) (Grime 1979; Table 2). An intermediate disturbance condition 263 favors the coexistence of competitive species and stress-tolerant species (due to the resources 264 limitation) (Mackey and Currie 2001). Unproductive habitats select perennial species with 265 slower plant growth, longer life spans, denser tissues, in which resources are conserved more

266 efficiently (stress-tolerant S). Stress-tolerant species, found in low stress, low disturbance 267 habitats, allocate resources to maintenance and defenses, such as anti-herbivory (Grime 268 2001). Finally productive habitats (no disturbed and no stressed habitats) select mostly annual 269 species for the ability to pre-empt resources by foraging (competitors, C) (Table 2). 270 Competitors are primarily composed of species with high relative growth rate, short leaf-life, 271 relatively low seed production, and high allocation to leaf construction. They persist in high 272 nutrient and low disturbance environments. Thus, in contrast to the denser tissues, low 273 specific leaf area and concomitantly slow growth of stress-tolerators (Poorter and Jong 1999; 274 Poorter and van der Werf 1998; Weiher et al. 1999), both competitors and ruderal species are 275 characterized by high specific leaf area and faster relative growth rates (resulting from greater 276 internal conductivity and lesser investment in structural tissues), with ruderal species 277 investing more in the reproductive phase of the life cycle (Table 2).

278 We assume that in tillage treatments, tolerance and escape are the two types of responses 279 that species will adopt resulting in small size, with high growth rates (as found for example in 280 Storkey 2006; Fried et al. 2012). Plants growing in nutrient-poor or dry environments will 281 increase their ability to access soil resources (by increasing root biomass or their capacity to 282 fix atmospheric nitrogen). In cover crop treatments, it is expected that the less competitive 283 weed species will be excluded, while only some competitive species that can rapidly reach a 284 high size (relative to the cover crop), high specific root length and depth, and efficient 285 nutrient use strategies would be able to maintain Alternatively, we assume that spontaneous 286 species having different traits and using resources differently than the cover crop could also 287 maintain (e.g. early flowering weed species able to produce seeds before the closure of the 288 canopy of the cover crop).

In addition, we hypothesize that disturbance (represented by tillage treatment) act as a
filter resulting in the convergence of traits within communities (i.e. reduction in trait variation

with increased disturbance), whereas cover crop and spontaneous vegetation will result to less
similar species in the communities according to the limiting similarity hypothesis (Fig. 2b).
This means that the functional divergence (i.e. the degree to which the distribution of species
abundances in niche space maximizes total community variation in functional characters) of
the different weed species is the result of species separate niches and thus complementary in
resource use.

- 297
- 298 Insert Table 2 here
- 299

300 Effect traits of weed communities and services on vineyard ecosystems

301 Several studies have shown that many ecosystem properties (for example above-ground net 302 primary productivity, litter decomposition, soil nitrogen, soil water content, digestibility...) 303 are controlled by the traits of dominant species or/and functional diversity of communities 304 (see Garnier et al. 2016 for a detailed review). These numerous studies were conducted in 305 diverse ecosystems (secondary succession, grasslands, forests or serpentine ecosystems) but 306 none of them considered perennial cropping. Recently, Gamour et al. (2015) discussed how 307 the trait-based approach could be used to assess the services delivered in cover cropped 308 banana cropping systems and identified which effect traits are related to these services.

Here we adapt the synthesis about traits and ecosystem services proposed by de Bello et al. (2010) to assess the contribution of vegetation species traits to different ecosystem services related to vineyards (Table 3). Following de Bello et al. (2010), we consider that there is a group of traits such as growth form, canopy density and plant size as well as the root system that are involved in water regulation and soil stability. We assume that leaf and litter traits influence vine growth and soil fertility *via* an increase in the decomposition and mineralization processes (as shown for tree species by Wardle et al. 2002). In order to expand

the trait-services approach to services such as pollination (with different trophic groups are involved) we should consider different trophic levels (Lavorel et al. 2013). To date, this approach has never been tested in vineyards or any other agricultural system (except in grasslands). The challenge for future works is to establish the linkages between traits and services to provide a basis for practical agroecosystem management and decision-making.

321

322 Insert Table 3 here

323

A case study: Short-term dynamics of weed communities in response to different soil management and effects in a vineyard performance of southern France

326 In Mediterranean regions, adaptation to climate fluctuations is a recurrent management 327 problem in agricultural production (Ripoche et al. 2011). In perennial systems, like vineyards, 328 adaptation is possible through canopy management (Smart et al., 1991), fertilization, 329 irrigation, or soil surface and intercropping management (Celette and Gary, 2006; Chifflot et 330 al., 2006). However only 14% of vineyards in the Mediterranean regions are intercropped, 331 (Mezière et al. 2009) as wine growers in those regions fear occasional episodes of strong 332 competition for water between the two crops and are reluctant to introduce cover crops 333 despite the regulating services they would provide (Ripoche et al. 2011). In this study, we 334 compare response traits of associate plant diversity in three management practices and test 335 their links with agroecosystem properties. This first attempt to adapt the trait-response framework to vineyards may permit to demonstrate how management practices can achieve 336 337 target traits and those traits will achieve goals for the rates of ecosystem properties and 338 services (as suggested at the theoretical framework of Wood et al. 2015). More precisely we 339 tested the effect of three management practices (tillage, managed spontaneous vegetation and 340 cover crop) commonly used in Mediterranean vineyards on different components of 341 taxonomic and functional diversity.

342 The field experiment is detailed in (Guilpart, Metay, and Gary 2014). Grapevines 343 (Vitis vinifera L. cv.Shiraz) were planted in 2002, in rows oriented NW-SE at a density of 344 3333 stocks per hectare (2.5 m \times 1.2 m). Three treatments were designed to create a gradient 345 of soil resources (water and nitrogen): (a) A first treatment was obtained by sowing a mix of 346 annual medics (Medicago truncatula, M. rigidula, M. polymorpha) in the inter-row during 347 autumn 2009 (cover crop treatment hereafter); (b) A second treatment with bare soil was 348 obtained by mechanical weeding in the inter-row with three operations in spring (in April, 349 May and June in 2010 and in March, April and June in 2011), (tillage treatment hereafter); (c) 350 permanent natural plant coverage between rows (spontaneous cover treatment) mowed twice 351 a year. There was no fertilization or irrigation in these treatments. Treatments were applied as 352 strips. Cover crop and Tillage treatments were composed of 185 vine stocks (37 vinestocks 353 per row and 5 rows) whereas Spontaneous cover treatment composed of 74 vine stocks (37 354 vinestocks per row and 2 rows).

355 Depending on the regime of disturbance due to practical operations in the vineyard, 356 weed sampling covers was conducted at several dates during a year, especially before and 357 after treatments and at least two years to record weed communities responses (measured traits 358 are detailed in Figures 3 and 4). We also monitored vine yield and growth, and water and 359 nitrogen grapevine status, as previous research has documented significant effects of cover 360 crops on these key parameters (Ingels et al. 2005; Tesic, Keller, and Hutton 2007b; King and 361 Berry 2005). After calculation of several indices, their distribution was tested for normality 362 and log₁₀ transformed when required. A repeated ANOVA was performed to test the effect of 363 three treatments, the effect of different sampling dates and their interaction as the different 364 measurements made on the same plots at different dates were not independent. We tested 365 main effect of date and treatment, the additive effect and their interaction. The model best 366 supported by the data was selected based on the Akaike Information Criterion (AIC; Burnham

and Anderson 2002). Post-hoc tests were performed to test the effect of treatment or date on
different variables. Statistical analyses were performed under R environment (R Development
Core Team 2010) using the lme4, plotrix and multcomp packages.

370

371 *Hypothesis 1: Inter-row management treatments as a gradient of disturbance and competition*372 *intensity*

373 We first hypothesized that the three management practices represent a double gradient of 374 disturbance and competition intensity, which acts as filters affecting species richness and 375 abundance. We expected that tillage treatment would favor a small number of dominant 376 species (low Shannon indices) whereas vegetation cover will favor a higher number of 377 species, equally abundant (high Shannon indices). Our results showed that diversity indices 378 varied among the three management practices, and the sampling date (Fig. 3). More precisely, 379 spontaneous treatment showed the highest number of species with an average of 14 species 380 per m² observed in all the sampling dates while tillage treatment harbours the lowest number 381 of species and the highest variation across dates (Fig. 3a). Shannon diversity index was 382 highest in the spontaneous treatment and lowest in the tillage treatment, indicating that in this 383 treatment only a limited number of species were very abundant (Fig. 3b). High disturbance, 384 corresponding to tillage treatment, resulted in less diverse communities dominated by a small 385 number of species whereas an intermediate disturbance resulted in more diverse and equitable 386 communities. On the contrary, there are some species that are very tolerant to all treatments 387 e.g. Poa annua, Crepis sancta and Senecio vulgaris in April 2010 and Malva sylvestris in 388 May 2011 which have also been shown to be among the most generalist weed species able to 389 withstand a large range of management and ecological conditions in annual arable fields 390 (Fried, Petit, and Reboud, X. 2010). Weed and cover crop biomass did not vary between the 391 three treatments in April 2010 (Fig. 3c). Spontaneous and cover crop treatments did not

392 present significant differences during the three dates of measurements, however it is

important to notice that cover crop treatment presented a big variability across the different

394 plots (certainly due to the establishment difficulties of the cover crop in the first year). In the

tillage treatment, biomass was destroyed just before measurements in May and June.

396

397 Insert Figure 3 here

398

399 *Hypothesis 2: Functional structure of weed communities responds to management treatments* 400 The second hypothesis was about species response traits to different management practices: 401 that tillage treatment will favor species with traits corresponding to ruderal strategy (high 402 growth rate, traits favoring resources acquisition), whereas cover crop will favor competitive 403 species. To test this hypothesis we characterized weeds functional composition in the three 404 inter-row treatments. We choose traits reflecting species morphology, phenology and 405 reproduction (traits values were obtained in standardized databases: Leda, BiolFlor and 406 Badoma). We then calculated community-weighted means (CWM; Garnier et al. 2004) and 407 Functional richness index which represents the amount of functional space filled by the 408 community (Villéger, Mason, and Mouillot 2008) for plant communities in the three 409 treatments for all the sampling dates. Our results confirm the hypothesis that tillage treatment 410 favored species with trait values associated to extensive exploitation of productivity-related 411 resources and fast growth (especially high specific leaf area values). CWM of all the traits 412 varied significantly among the three treatments. Species from the tillage treatment showed the 413 lowest reproductive height values, the highest specific leaf area (SLA) values and late onset 414 of flowering. The opposite pattern was found for species from the cover treatment (Fig. 4). 415 This pattern can be explained in part if we consider the early onset of flowering observed in 416 plants for the cover crop treatment: species in this treatment tend to flower and produce seeds

417 earlier in season than Medicago plants. Additionally, species in the cover plant treatment 418 showed the higher reproductive height values as a response to competition for light. The 419 results of this study agree with the findings of (Laliberté and Tylianakis 2012; Guerrero et al. 420 2014; Flynn et al. 2009) who showed that the intensification of land use reduced functional 421 diversity and redundancy. We also tested whether tillage treatment (highly disturbed habitat), 422 act as a filter resulting in the convergence of traits within communities whereas plant cover 423 (sown or spontaneous) vegetation will result to less similar species in the communities with 424 large trait variability. We found that spontaneous vegetation treatment showed highest 425 functional richness whereas in tillage treatment functional richness decreased even if species 426 richness increased, thus implying functional convergence (Fig. 4d). Communities in 427 spontaneous vegetation treatment showing high functional divergence, which indicates a high 428 degree of niche differentiation, and thus low resource competition. Thus communities with 429 high functional divergence may have increased ecosystem function, especially improvement 430 of soil fertility and nutrient cycling, as a result of more efficient resource use (N. W. H. 431 Mason et al. 2005).

432

433 Insert Figure 4 here

434

435 Hypothesis 3: Species traits affect ecosystem properties

We hypothesized that different soil management practices in the vineyards will affect vine growth and performance. Although, spontaneous cover is supposed to have a higher impact on vine yield compared to tillage; it is expected that the effect of the legume cover crop is intermediate after two years, since competition with the vines is compensated by an increase in nitrogen supply and that the presence or the vigor of competitive weeds that may be present in spontaneous cover is reduced. However, spontaneous and tillage treatments presented

442 significantly higher grapevine yields than the cover crop treatment around 4000 g/vine (which corresponds to about 12 t of grapes ha⁻¹ according the density of vines per ha) in spontaneous 443 and tillage treatments versus around 2500 g/ vine (corresponding to 7.5 t grapes ha⁻¹) in the 444 445 cover treatment) (Fig. 5a). This result can be explained by the non significant differences in 446 biomass of spontaneous and cover crops and the higher functional trait divergence (especially 447 of SLA, which corresponds to resource use). Predawn leaf water potential, a dynamic 448 indicator of the water stress undertaken by the vine, shows that cover treatment always 449 created higher water stress conditions for the vine throughout the summer (-0.64 MPa in late 450 august versus -0.57 MPa and -0.58 for tillage and spontaneous treatment respectively) 451 (Fig.5b). Assimilable nitrogen in must was significantly higher at harvest at the tillage treatment (168 mg L^{-1} versus 80 mg L^{-1} and 105 mg L^{-1} for spontaneous and cover treatment 452 453 respectively). For the latter two, the assimilable nitrogen content in must is below the threshold of 140 mg L^{-1} generally considered as nitrogen deficiency threshold (Casalta, 454 455 Sablayrolles, and Salmon 2013). Cover treatment based on legumes did not show any 456 improvement in vine nitrogen nutrition even if there is a slight increase (Fig. 5c) probably 457 because of the water stress created by the cover crop. In this case, the most limiting factor is 458 water (Fig. 5b).

459 Lavorel and Grigulis (2012) proposed a framework based on alpine grasslands, 460 differing in their management regime, corresponding to a gradient of management intensity: 461 intensive management practices (like mowing and fertilization), favor species with traits 462 values associated to high resources acquisition (high specific leaf surface for example), and 463 thus influencing ecosystem processes and services (such as net primary productivity and rapid 464 biochemical cycles). In our study, we found a positive relationship between specific leaf area 465 CWM and vineyard yield (Fig. 6): communities dominated by species with high specific leaf 466 area are linked to higher yield than communities dominated by species with low specific leaf

467 area. This result, found for the first time in perennial crop systems like vineyards, allows 468 testing directly the trait-service link and the hypothesis presented in Fig. 1b. This hypothesis 469 should be tested on other ecosystem properties such as decomposition and mineralization as 470 we hypothesize that intensive management practices, like tillage, acting like filter, for species 471 traits favoring high growth rates, rapid nutrient mobilization and then rapid decomposition 472 and mineralization. As a consequence, resources availability is high and vineyard yield 473 increases. Our results did not clearly distinguish tillage treatment and spontaneous vegetation 474 cover. There is a high variability in each treatment, which requires characterizing resources in 475 more details for each community to interpret the results.

- 476
- 477

Insert Figure 6 here.

478

479 Conclusion and perspectives

480 In conclusion, both spontaneous and tillage treatments appeared acceptable as far as 481 yield was maintained. The advantage of the tillage treatment on one hand was that it seemed 482 efficiently able to limit weed growth (see biomass data in Fig. 3c) and consequently maintain 483 both an adequate water status for the vine and a high level of assimilable nitrogen. On the 484 other hand, spontaneous vegetation treatment can be chosen as a costless tradeoff for the 485 winegrowers between improving soil properties, limiting mechanical work and maintaining 486 vine production. This case study was the first demonstration of a direct relationship between 487 functional traits and ecosystem services in perennial crop systems. This result confirms our 488 initial hypothesis that the functional characterization of communities' responses to different 489 management practices is a key issue to the management of inter-row communities, as traits 490 with certain functional properties (e.g., ability to fix N₂) can be chosen in order to select or 491 promote spontaneous species with those traits.

492	In this paper, we introduced a trait-based framework to assess vineyard inter-row soil							
493	management. As demonstrated in our case study, this framework has the potential to improve							
494	the understanding of weed community assembly and allow an adaptation of different soil							
495	management practices. According to this framework, it is essential to:							
496	(i) Characterize different management practices as disturbance and resource gradient							
497		in order to identify which traits may respond to these gradients;						
498	(ii)	Quantify resources availability and disturbance intensity; this will inform on which						
499		factors determine the abundance and species richness. For associated diversity, it						
500		is important to take the periodicity of disturbance turnover into account;						
501	(iii)	Identify and measure traits that are closely related to functions and processes						
502		across the gradients (response traits) as different functional traits are important for						
503		different services and ecosystem processes (see Table 3);						
504	(iv)	Measure the processes and services;						
505	5 (v) Compare the diversity and ecosystem function(s)/service(s) of different							
506		management practices;						
507	(vi)	Test the linkages between traits and services (effect traits) and propose						
508		quantitative relationships. These relationships would make it possible to use traits						
509		to scale from individual plants and the communities they form to the ecosystem						
510		level function.						
511								
512	Given the	e results obtained, different "ideotypes", i.e. species with required traits, providing						
513	several services, could be proposed considering both grapegrower's objectives. The choice of							

- those "ideotypes" will contribute to design more sustainable grapevine systems. More
- 515 precisely based on Table 3 we propose species with:

516 (i) High specific leaf area and low dry matter content as well as N fixing species
517 which will improve soil fertility and nutrient cycling;
518 (ii) High root density in order to support water regulation;
519 (iii) Rosette and herbaceous species with high growth rate which will support soil
520 stability.
521
522

523 **References**

- 524 Balvanera, P., C. Kremen, and M. Martinez-Ramos. 2005. "Applying Community Structure
- 525 Analysis to Ecosystem Function: Examples from Pollination and Carbon Storage."
- 526 *Ecological Applications* 15 (1): 360–75.
- 527 Barralis, G., G. Cloquemin, and A. Guérin. 1983. "Evolution de La Flore Adventice Du
- 528 Vignoble de Côte-d'-Or Sous La Pression Des Techniques D'Entretien Des Cultures."
 529 Agronomie, EDP Sciences 3 (6): 585–94.
- Beuret, E., and G. Neury. 1990. "L'entretien Du Sol En Viticulture." *Revue Suisse de Viticulture Arboriculture Horticulture* 22 (1): 63–68.
- Booth, B D, and C J Swanton. 2002. "Assembly Theory Applied to Weed Communities." *Weed Science* 50 (1): 2–13.
- 534 Braun-Blanquet, J. 1964. *Pflanzensoziologi. Plant Sociology. Pflanzensoziologi. Plant*535 Sociology.
- 536 Cadotte, M.W., K. Carscadden, and N. Mirotchnick. 2011. "Beyond Species: Functional
- 537 Diversity and the Maintenance of Ecological Processes and Services." *Journal of*538 *Applied Ecology* 48: 1079–87.
- 539 Cardina, J., C.P. Herms, and D.J. Doohan. 1991. "Long-Term Tillage Affects on Seed Banks
 540 in Three Ohio Soils." *Weed Science* 50: 448–60.
- 541 Cardinale, B.J., J.E. Duffy, A. Gonzalez, D.U. Hooper, C.Perrings, P. Venail, A. Narwani, et
- al. 2012. "Biodiversity Loss and Its Impact on Humanity." *Nature* 486 (7401): 59–67.
- 543 Casalta, E., J.M. Sablayrolles, and J.M. Salmon. 2013. "LWT Food Science and Technology
- 544 Comparison of Different Methods for the Determination of Assimilable Nitrogen in
- 545 Grape Musts." *LWT Food Science and Technology* 54 (1): 271–77.
- 546 Celette, F., R. Gaudin, and C. Gary. 2008. "Spatial and Temporal Changes to the Water

- 547 Regime of a Mediterranean Vineyard due to the Adoption of Cover Cropping." *Europ. J.*548 *Agronomy* 29: 153–62.
- Chapin, F.S., E.S. Zavaleta, V.T. Eviner, R.L. Naylor, P.M. Vitousek, H.L. Reynolds, D.U.
 Hooper, et al. 2000. "Consequences of Changing Biodiversity." *Nature* 405 (6783): 234–
 42.
- 552 Chifflot, V., Bertoni, G., Cabanettes, A. and Gavaland, A. 2006. "Beneficial Effects of
 553 Intercropping on the Growth and Nitrogen Status of Young Wild Cherry and Hybrid
 554 Walnut Trees." *Agroforestry Systems*: 66(1): 13-21.
- 555 Connell, J. H. 1978. "Diversity of tropical rainforests and coral reefs". Science 199: 1304–
 556 1310.
- 557 Cousens, R.D., and S.R. Moss. 1990. "A Model of the Effects of Cultivation on the Vertical
 558 Distribution of Weed Seeds within the Soil." *Weed Research* 30: 61–70.
- 559 Dastgheib, F., and C. Frampton. "Weed Management Practices in Apple Orchards and
- 560 Vineyards in the South Island of New Zealand." J. Crop Hort. Sci. 28: 53–58.
- 561 Dawson, T. E., and F. S. I. Chapin. 1993. "Grouping plants by their form-function
- 562 characteristics as an avenue for simplification in scaling between leaves and
- 563 landscapes". In *Scaling Physiological Processes: Leaf to Globe* (eds. J. R. Ehleringer
- and C. B. Field), pp. 313-319. Academic Press Inc., San Diego
- de Bello, F., S.Lavorel, S.Díaz, R.Harrington, J.H.C. Cornelissen, R.D. Bardgett, M.P. Berg,
- t al. 2010. "Towards an Assessment of Multiple Ecosystem Processes and Services via
 Functional Traits." *Biodiversity and Conservation* 19 (10): 2873–93.
- 568 Delabays, N., J.L. Spring, and G. Mermillod. 2006. "Study of Cover Crops in Vineyards with
- 569 Weakly Competitive Species: Botanical and Weed Aspects." *Revue Suisse de Viticulture*,
- 570 *Arboriculture et Horticulture* 38 (6): 343–54.

- 571 Derksen, D.A., G.P. Lafond, A.G. Thomas, H.A. Loeppky, and C.J. Swanton. 1993. "Impact
 572 of Agronomic Practices on Weed Communities: Tillage Systems." *Weed Science* 41:
 573 409–17.
- 574 Diaz, S., and M.R. Cabido. 1997. "Plant Functional Types and Ecosystem Function in
 575 Relation to Global Change." *Journal of Vegetation Science* 8: 463–74.
- 576 Flynn, D.F.B., M.Gogol-Prokurat, T. Nogeire, N. Molinari, B.Trautman Richers, B.B. Lin,
- 577 N.Simpson, M.M. Mayfield, and F. De Clerck. 2009. "Loss of Functional Diversity
- under Land Use Intensification across Multiple Taxa." *Ecology Letters* 12 (1): 22–33.
- 579 Fredrikson, L., P.A. Skinkis, and E. Peachey. 2011. "Cover Crop and Floor Management
- 580 Affect Weed Establishing Coverage and Density in an Oregon Vineyard." *Hort*581 *Technology* 21: 208–16.
- 582 Fried, G., B. Chauvel, and X. Reboud. 2009. "A Functional Analysis of Large-Scale
- Temporal Shifts from 1970 to 2000 in Weed Assemblages of Sunflower Crops in
 France." *Journal of Vegetation Science* 20:49-58
- 585 Fried, G., E. Kazakou, and S. Gaba. 2012. "Trajectories of Weed Communities Explained by
- 586 Traits Associated with Species' Response to Management Practices." *Agriculture*,
 587 *Ecosystems and Environment* 158: 147–55.
- 588 Fried, G., L. Norton, and X. Reboud. 2008. "Environmental and Management Factors
- 589 Determining Weed Species Composition and Diversity in France." *Agriculture*,
- 590 *Ecosystems & Environment* 128 (1-2): 68–76.
- Fried, G., S. Petit, and Reboud, X. 2010. "A Specialist-Generalist Classification of the French
 Arable Flora and Its Response to Changes in Agricultural Practice." *BMC Ecology* 10
 (20): 10-20..
- 594 Gaba, S., G. Fried, E.Kazakou, B.Chauvel, and M.L. Navas 2015. "Agroecological Weed

- 595 Control Using a Functional Approach : A Review of Cropping Systems Diversity."
- 596 *Agronomy for Sustainable Development* 34: 103–19.
- 597 Gago, P., C. Cabaleiro, and J. García. 2007. "Preliminary Study of the Effect of Soil
- 598 Management Systems on the Adventitious Flora of a Vineyard in Northwestern Spain."
- 599 *Crop Protection* 26 (4): 584–91.
- 600 Gamour, G., E. Garnier, M.L. Navas, M. Dorel, and J.M. Risède. 2015. "Using Functional
- 601Traits to Assess the Services Provided by Cover Plants: A Review of Potentialities in
- Banana Cropping Systems." *Advances in Agronomy* 134:81-133.
- 603 Garnier, E., M.L. Navas, and K. Grigulis. 2016. Planr Functional Diversity Organism Traits,
- 604 *Community Structure and Ecosystem Properties*. Oxford Uni. Oxford.
- 605 Garnier, E., J. Cortez, G. Billès, M.L. Navas, C.Roumet, M. Debussche, G.d Laurent, et al.
- 606 2004. "Plant Functional Markers Capture Ecosystem Properties During Secondary
 607 Succession." *Ecology* 85 (9): 2630–37.
- 608 Garnier, E., and M.L. Navas. 2011. A Trait-Based Approach to Comparative Functional Plant
- Ecology: Concepts, Methods and Applications for Agroecology. A Review. Agronomy for
 Sustainable Development. 32(2): 365-399
- 611 Grime, J.P. 1979. "Plant Strategies and Vegetation Processes." *Journal of Ecology* 68 (2):
 612 704–6.
- 613 Grime, J.P. 1998. "Benefits of Plant Diversity to Ecosystems: Immediate, Filter and Founder
- 614 Effects." *Journal of Ecology* 86 (6): 902–10.
- 615 Grime, J.P. 2006. "Trait Convergence and Trait Divergence in Herbaceous Plant
- 616 Communities: Mechanisms and Consequences." *Journal of Vegetation Science* 17 (2):
- 617 255–60.
- 618 Grime, J. P. 2001. Plant strategies, vegetation processes, and ecosystem properties. 2nd. ed.

- 619 Wiley, Chichester, UK.
- 620 Gross, N., T. M. Robson, S. Lavorel, C. Albert, Y. Le Bagousse-Pinguet, and R. Guillemin.
- 621 2008. "Plant Response Traits Mediate the Effects of Subalpine Grasslands on Soil
- 622 Moisture." *New Phytologist* 180: 652–62.
- 623 Guerrero, I., C.P. Carmona, M. B. Morales, J. J. Oñate, and B. Peco. 2014. "Non-Linear
- 624 Responses of Functional Diversity and Redundancy to Agricultural Intensification at the
- Field Scale in Mediterranean Arable Plant Communities." *Agriculture, Ecosystems & Environment* 195: 36–43.
- 627 Guilpart, N., A.Metay, and C.Gary. 2014. "Grapevine Bud Fertility and Number of Berries
- 628 per Bunch Are Determined by Water and Nitrogen Stress around Flowering in the
- 629 Previous Year." *European Journal of Agronomy* 54: 9–20.
- Gunton, R.M., S. Petit, and S. Gaba. 2011. "Functional Traits Relating Arable Weed
- 631 Communities to Crop Characteristics." *Journal of Vegetation Science* 22 (3): 541–50.
- 632 Ingels, C. A., K. M. Scow, D.A. Whisson, and R.E. Drenovsky. 2005. "Effects of Cover
- 633 Crops on Grapevines, Yield, Juice Composition, Soil Microbial Ecology, and Gopher
 634 Activity." *American Journal of Enology and Viticulture* 56 (1): 19–29.
- 635 King, A.P., and A.M. Berry. 2005. "Vineyard 15N, Nitrogen and Water Status in Perennial
- 636 Clover and Bunch Grass Cover Crop Systems of California's Central Valley."
- 637 *Agriculture, Ecosystems and Environment* 109 (3-4): 262–72.
- 638 Kleyer, M., R. M. Bekker, I. C. Knevel, J. P. Bakker, K. Thompson, M. Sonnenschein, P.
- 639 Poschlod, et al. 2008. "The LEDA Traitbase: A Database of Life-History Traits of the
 640 Northwest European Flora." *Journal of Ecology*. 96 (6): 1266–1274.
- 641 Laliberté, E., and J.M. Tylianakis. 2012. "Cascading Effects of Long-Term Land-Use
- 642 Changes on Plant Traits and Ecosystem Functioning." *Ecology* 93 (1): 145–55.

- 643 Lavorel, S, and E Garnier. 2002. "Predicting Changes in Community Composition and
- Ecosystem Functioning from Plant Traits" 16(5): 545–56.
- 645 Lavorel, S. 2013. "Plant Functional Effects on Ecosystem Services." *Journal of Ecology* 101:
 646 4–8.
- Lavorel, S. and K. Grigulis. 2012. "How Fundamental Plant Functional Trait Relationships
 Scale-up to Trade-Offs and Synergies in Ecosystem Services." *Journal of Ecology* 100
 (1): 128–40.
- Lavorel, S., J. Storkey, R.D. Bardgett, F.de Bello, M.P. Berg, X.Le Roux, M. Moretti, et al.
- 651 2013. "A Novel Framework for Linking Functional Diversity of Plants with Other
- 652 Trophic Levels for the Quantification of Ecosystem Services." *Journal of Vegetation*653 *Science* 24 (5): 942–48.
- Legere, A., F.C. Stevenson, and A. Vanasse. 2011. "Short Communication: A Corn Test Crop
 Confirms Beneficial Effects of Crop Rotation in Three Tillage." *Can. J. Plant Sci.*, 91:
 943–46.
- Litrico, I., and C. Violle. 2015. "Diversity in Plant Breeding : A New Conceptual
 Framework." *Trends in Plant Science* 20 (10): 604–13. d
- Lososova, Z, J Danihelka, and M Chytry. 2003. "Seasonal Dynamics and Diversity of Weed
 Vegetation in Tilled and Mulched Vineyards." *Biologia* 58 (1): 49–57.
- Mackey, R. L., and D. J. Currie. 2001. "The Diversity- Disturbance Relationship: Is It
 Generally Strong and Peaked?" *Ecology* 82: 3479–92.
- 662 Generally Strong and Peaked?" *Ecology* 82: 3479–92.
- Maillet, J. 1980. "Evolution de La Flore Adventice Des Vignobles Du Montpellierais Sous La
 Pression Des Herbicides." In *IVème Coll. Biol. Ecol. Syst. Mauvaises Herbes*, 359–66.
- Mason, N.W. H., D. Mouillot, W. G. Lee, and J.B. Wilson. 2005. "Functional Richness,
- 666 Functional Evenness and Functional Divergence: The Primary Components of

667 Functional Diversity." *Oikos* 111 (1): 112–18.

- 668 Mézière, D., Gary, C., Barbier, J.M., Bernos, L., Clément, C., Constant, N., Delière, L.,
- 669 Forget, D., Grosman, J., Molot, B., Pio, P., Sauvage, D., Sentenac, G. 2009. "Vers des
- 670 systèmes de cultures économes en produits phytosanitaires. Volet 1 : Analyse
- 671 comparative de différents systèmes en viticulture. "*ECOPHYTO R&D*. (Tome III).
- 672 INRA.
- Minden, V., and M. Kleyer. 2011. "Testing the Effect–response Framework: Key Response
- and Effect Traits Determining above-Ground Biomass of Salt Marshes." *Journal of*
- 675 *Vegetation Science* 22: 387–401.
- Monteiro, A., and I. Moreira. 2004. "Reduced Rates of Residual and Post-Emergence
 Herbicides for Weed Control in Vineyards." *Weed Research* 44: 117–28.
- Monteiro, A., and C.M. Lopes. 2007. "Influence of Cover Crop on Water Use and
- 679 Performance of Vineyard in Mediterranean Portugal." Agriculture, Ecosystems &
- 680 *Environment* 121 (4): 336–42. Moreira, I. 1994. "Entretien Du Sol et Évolution de La
- 681 Flore." *Phytoma-La Défense Des Végétaux* 462: 28–30.
- 682 Naeem, S., and J.P. Wright. 2003. "Disentangling Biodiversity Effects on Ecosystem
- Functioning: Deriving Solutions to a Seemingly Insurmountable Problem." *Ecology Letters* 6: 567–79.
- Navas, M.L. 2012. "Trait-Based Approaches to Unravelling the Assembly of Weed
- 686 Communities and Their Impacts on Agro-Ecosystem Functionning." *Weed Research* 52
 687 (6): 479–88.
- Petchey, O.L., and K.J. Gaston. 2006. "Functional Diversity: Back to Basics and Looking
 Forward." *Ecology Letters* 9 (6): 741–58.
- 690 Poorter, H, and A van der Werf. 1998. "Is Inherent Variation in RGR Determined by LAR at

- 691 Low Light and by NAR at High Light?" In *Inherent Variation in Plant Growth:*
- 692 Physiological Mechanisms and Ecological Consequences., (eds H. Lambers, H. Poorter
- 693 & M. M. I. Van Vuuren), pp. 309–336. Backhuys Publishers, Leiden.
- 694 Poorter, H., and R.D. Jong. 1999. "A Comparison of Specific Leaf Area, Chemical
- 695 Composition and Leaf Construction Costs of Field Plants from 15 Habitats Differing in
 696 Productivity." *New Phytol.* 143: 163-176.
- 697 Renting, H., W.A.H. Rossing, J.C.J. Groot, J.D. Van der Ploeg, C. Laurent, D. Perraud, D.J.
- 698 Stobbelaar and M.K. Van Ittersum 2009. "Exploring multifunctional agriculture. A
- 699 review of conceptual approaches and prospects for an integrative transitional
- framework." J. Environ. Manage. 90: 112–123.
- Ripoche, A., A. Metay, F. Celette, and C. Gary. 2011. "Changing the soil surface
- management in vineyards: Immediate and delayed effects on the growth and yield of
 grapevine." *Plant Soil* 339: 259-271.
- Smart, R.E., Dick, J.K., Gravett, I.M. and Fisher, B.M. 1991. "Canopy management to
- 705 improve grape yield and wine quality: Principles and practices. "*South african journal*706 *of Enology and Viticulture*: 11(1): 3-17.
- 707 Steenwerth, K.L., A. Calderón-Orellana, R.C. Hanifin, C. Storm, and A.J. Mcelrone. 2016.
- 708 "Effects of Various Vineyard Floor Management Techniques on Weed Community
- 709 Shifts and Grapevine Water Relations." 2: 153–62.
- 710 Storkey, J. 2004. "Modelling Seedling Growth Rates of 18 Temperate Arable Weed Species
- as a Function of the Environment and Plant Traits." *Annals of Botany* 93 (6): 681–89.
- 712 Storkey, J. 2006. "A Functional Group Approach to the Management of UK Arable Weeds to
- 713Support Biological Diversity." Weed Research 46: 513–22.
- 714 Storkey, J., and D.B. Westbury. 2007. "Managing Arable Weeds for Biodiversity." Pest

- 715 *Management Science* 63 (December 2006): 517–23.
- 716 Suding, K.N., S. Lavorel, F. S. Chapin, J.H.C. Cornellissen, S. Diaz, E. Garnier, D.E.
- Goldberg, D U Hooper, S.T. Jackson, and M.L. Navas. 2008. "Scaling Environmental
- 718 Change through the Community-Level: A Trait-Based Response-Effect- Framework for
- 719 Plants." *Global Change Biology* 14: 1125–40.
- 720 Tesic, D., M. Keller, and R.J. Hutton. 2007a. "Influence of Vineyard Floor Management
- Practices on Grapevine Vegetative Growth, Yield, and Fruit Composition." *American Journal of Enology and Viticulture* 58 (1): 1–11.
- 723 Tilman, D., J. Knops, D. Wedin, P. Reich, M. Ritchie, and E. Siemann. 1997. "The Influence
- of Functional Diversity and Composition on Ecosystem Processes." *Science* 277: 1300–
 1303.
- Tilman, D., D. Wedin, and J. Knops. 1996. "Productivity and Sustainability Influenced by
 Biodiversity in Grassland Ecosystems." *Nature* 379: 718–20.
- 728 Villéger, S., N.W.H. Mason, and D. Mouillot. 2008. "New Multidimensional Functional
- Diversity Indices for a Multifaceted Framework in Functional Ecology." *Ecology* 89 (8):
 2290–2301.
- Violle, C., E. Garnier, J. Lecoeur, C. Roumet, C.Podeur, A. Blanchard, and M.L.Navas. 2009.
 "Competition, Traits and Resource Depletion in Plant Communities." *Oecologia* 160 (4):
 733 747–55.
- Violle, C., and L. Jiang. 2009. "Towards a Trait-Based Quantification of Species Niche." *Journal of Plant Ecology* 2 (2): 87–93.
- 736 Wardle, D.A., Bonner, K.I. & Barker, G.M. 2002). "Linkages between plant litter
- decomposition, litter quality, and vegetation responses to herbivores. "*Functional*
- *Ecology* 16: 585–595.

739	Weiher, E, A	Van Der Werf, K	Thompson,	M Roderick, E	Garnier, and C) Eriksson. 1999

740 "Challenging Theophrastus: A Common Core List of Plant Traits for Functional

741 Ecology." *Journal of Vegetation Science* 10 (1968): 609–20.

- 742 Westoby, M. D.S. Falster, A.T. Moles, P.A. Vesk, and I.J. Wright. 2002. "Plant ecological
- strategies : Some Leading Dimensions of Variation Between Species." 91(6): 1005-1006.
- 744 Wilson, J. B. 1994. "The Intermediate Disturbance Hypothesis of Species Coexistence Is
- 745 Based on Patch Dynamics. New Zealand Journal of Ecology" 18: 176-181.
- 746 White, P.S., and S.T.A. Pickett. 1985. "Natural Distrubance and Patch Dynamics: An
- 747 Introduction." In *The Ecology of Natural Disturbance and Patch Dynamics*, edited by
- 748 S.T.A. Picket and P.S. White, Academic Press, 3–13. New York.
- 749 Wood, Stephen A, Daniel S Karp, Fabrice Declerck, Claire Kremen, Shahid Naeem, and
- 750 Cheryl A Palm. 2015. "Functional Traits in Agriculture : Agrobiodiversity and
- 751 Ecosystem Services,." 30 (9): 531–539.
- 752 Wright, I.J., P.B. Reich, M.Westoby, D.D. Ackerly, Z. Baruch, F.Bongers, J.Cavender-Bares,
- et al. 2004. "The Worldwide Leaf Economics Spectrum." *Nature* 428 (6985): 821–27.

754