



Disentangling the roles of social and individual effects on cadmium tolerance in the ant *Temnothorax nylanderi*

Romain Honorio, Lauren Jacquier, Claudie Doums, Mathieu Molet

► To cite this version:

Romain Honorio, Lauren Jacquier, Claudie Doums, Mathieu Molet. Disentangling the roles of social and individual effects on cadmium tolerance in the ant *Temnothorax nylanderi*. *Biological Journal of the Linnean Society*, 2021, 134 (4), pp.823-834. 10.1093/biolinnean/blab116 . hal-03391659

HAL Id: hal-03391659

<https://hal.science/hal-03391659>

Submitted on 21 Oct 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

Disentangling the roles of social and individual effects on cadmium tolerance in the ant *Temnothorax nylander*.

Romain Honorio^{*1}, Lauren Jacquier^{*1}, Claudie Doums^{2,3} & Mathieu Molet¹

¹ Sorbonne Université, Université Paris Est Créteil, Université Paris Diderot, CNRS, INRAE, IRD, Institute of Ecology and Environmental Sciences-Paris, iEES-Paris, F-75005 Paris, France

² Institut de Systématique Évolution Biodiversité, ISYEB, F-75005 Paris, Sorbonne Université, MNHN, CNRS, EPHE-PSL, Université des Antilles, France

³ EPHE, PSL University, F-75014 Paris, France

* These authors contributed equally to this work, ordered alphabetically.

Corresponding authors :

R. Honorio, Sorbonne Université, Université Paris Est Créteil, Université Paris Diderot, CNRS, INRAE, IRD, Institute of Ecology and Environmental Sciences-Paris, iEES-Paris, F-75005 Paris, France. E-mail address: romainhonorio@yahoo.fr

L. Jacquier, Sorbonne Université, Université Paris Est Créteil, Université Paris Diderot, CNRS, INRAE, IRD, Institute of Ecology and Environmental Sciences-Paris, iEES-Paris, F-75005 Paris, France. E-mail address: lauren.jacquier@gmail.com

Abstract

Urbanization brings new pressures for individuals. Among them, trace elements such as cadmium are important stressors. A recent study highlights a weaker negative effect of cadmium on city colonies relative to their forest counterparts in the ant *Temnothorax nylander*. Here, we aim to test whether this better tolerance of city colonies in this species results from (1) a better ability of workers to rear larvae despite stressful conditions and/or (2) a better ability of larvae to develop properly despite stressful conditions. We performed a cross-fostering experiment of workers and larvae from city and forest colonies, under common garden in the laboratory. Colonies were fed using cadmium-enriched or cadmium-free food for two months, and we measured four life-history traits. As expected, cadmium had a negative impact on all traits. Unexpectedly, we did not observe a better tolerance of city colonies to cadmium contrary to our previous study, preventing us from disentangling the respective contributions of workers and larvae to cadmium tolerance. Interestingly, forest

colonies seem to be of better quality in our laboratory conditions. Finally, colony size increased adult survival, but only in the absence of cadmium, suggesting that social buffering could collapse under strong external disturbances.

Key-words: trace element, body size, colony size, common garden, cross-fostering, urbanization.

Introduction

Cities are natural laboratories for investigating evolutionary changes in an urbanizing world (Alberti, Marzluff, & Hunt, 2017). Urbanization and human-induced environmental changes are forcing organisms to face many new pressures (Grimm *et al.*, 2008), including changes in impervious land cover and increases in temperature (including the ‘urban heat island’ - (Levermore *et al.*, 2018; Diamond & Martin, 2020). Pollution is also a major component of urbanization (be it sound, chemicals, light or others; e.g. Westby and Medley, 2020), especially trace metals which are persistent pollutants that bioaccumulate in the food chain. For example, cadmium is a highly toxic trace metal to various species (e.g. increasing mortality rate, decreasing hatching rate or impeding development, Gomot, 1998 (snails); Järup and Åkesson, 2009 (human); Malakar *et al.*, 2009 (grasshopper)) that can be found in concentrations ten times higher in cities than in rural soils (Foti *et al.*, 2017). These growing pressures lead to changes in various individuals traits (physiology, morphology, phenology, behavior), both plastically (e.g. changing in behavior (Ditchkoff, Saalfeld, & Gibson, 2006) or morphology (Prosser, Hudson, & Thompson, 2006)) and evolutionarily (e.g. changes in phenology (Chick *et al.*, 2019) or physiology (Partecke, Schwabl, & Gwinner, 2006) - see also Johnson and Munshi-South, 2017; reviewed in Alberti *et al.*, 2017). For instance, urban populations exhibit earlier timing of reproduction (plants: Jochner and Menzel, 2015; ants: Chick *et al.*, 2019), lower corticosterone stress response (birds: Partecke *et al.*, 2006), higher thermal tolerance (ants: Diamond *et al.*, 2017), decreased migratory behavior and aggressiveness (birds, respectively: Evans *et al.*, 2012; Hasegawa *et al.*, 2014, and smaller size (Daphnia: Brans *et al.*, 2017; spider: Dhirel *et al.*, 2019).

Although many organisms face these pressures alone, individuals living in groups benefit from a filter provided by their society against external disturbances. The social environment is a key component to be considered when studying the response of social species to external pressures, as it may buffer external disturbances (Straub *et al.*, 2015). For

example in social insects, colony size plays a role in buffering the effects of temperature, with larger colonies usually having less mortality or stronger trait stability (cold: Heinze *et al.*, 1996; warm: Molet *et al.*, 2017). This social buffering is found for various pressures, such as exposure to parasites (Scharf *et al.*, 2012), seasonality (Kaspari & Vargo, 1995), starvation (Heinze *et al.*, 1996; Modlmeier, Foitzik, & Scharf, 2013) and pesticides (Crall *et al.*, 2019). In addition to workers, the presence of larvae can also improve colony resistance to stress in harsh periods, such as under starvation (Rueppell & Kirkman, 2005; Modlmeier *et al.*, 2013). Furthermore, the social environment highly modulates brood development in insect societies such as in ants (Linksvayer, 2006; Wills *et al.*, 2018) or bees (Linksvayer *et al.*, 2011), and therefore can influence the phenotypes produced within colonies. Cross-fostering experiments, mixing worker and larval origins, are particularly useful to partition the effects of the social environment, that is assessing the relative contribution of adults and larvae. For example, crossfostering has been used in some ants species to assess the relative contributions of genetic and social effects on individual and colony traits (Linksvayer, 2006), the influence of social structure on brood development (Purcell & Chapuisat, 2012), the effects of matriline and patriline on immune defense (Armitage *et al.*, 2011) and the consequences of the rearing environment on individual and colonial personalities (Keiser *et al.*, 2015). However, as far as we are aware, it has never been used to investigate the relative effect of the social environment and larvae themselves on the ability to tolerate stressful conditions such as pollution.

In this study, we used a cross-fostering experiment while rearing colonies under common laboratory conditions in the ant *Temnothorax nylanderi* in the presence or absence of a stressor (cadmium). A previous study showed that city colonies were less negatively affected by cadmium than forest ones for two traits: the size of produced workers and the survival of larvae (Jacquier *et al.*, 2021a). In the present study, we used the same laboratory rearing setup and we measured the same traits (i.e. body size at emergence and survival) as well as the development duration. Moreover, in contrast with Jacquier *et al.* (2021a) who only assessed worker traits, we also measured sexual production and size of sexuals. Size of sexuals is an important trait linked with fitness in ants as larger sexuals have greater mating success and colony founding success in some species (Abell *et al.*, 1999; Wagner & Gordon, 1999; Wiernasz *et al.*, 2001; Wiernasz & Cole, 2003; Fjerdingstad & Keller, 2004). By performing a cross-fostering, our study allows us to test the relative contribution of the origin of workers (forest or cities) and of the origin of larvae (forest or cities) on the life history traits measured in presence and absence of cadmium.

Material and methods

Ants rearing

Temnothorax nylanderi is a tiny ant, nesting in acorns or twigs, widespread in western and central Europe and living in small colonies (a few dozen workers), and common in urban and forest habitats (Foitzik, Strätz, & Heinze, 2003). We collected colonies at two sites where a differential response to cadmium has been documented (Jacquier *et al.*, 2021a). We collected 103 city colonies in an urban park in Paris (Parc des Buttes-Chaumont, 48°52'47.1"N; 2°23'45.6"E) and 109 forest colonies in a forest located 50km away from Paris (Chantilly forest, 49°10'59.8"N; 2°28'43.6"E). Colonies were harvested in April 2019 and then installed in the laboratory. The artificial nest consisted of two microscope slides (75 x 26 x 1 mm) separated by a 1-mm auto-adhesive layer of plastic foam with 3 chambers. The top slide was covered with dark paper to protect the colony from the light. The nest was placed in a plastic box (11.5 x 11.5 x 5.5cm) for foraging. We discarded 52 colonies with more than one queen or with no queen (28 colonies in city, 24 in forest), as well as 18 colonies (3 in city, 15 in forest) with cestode-infected workers (that display typical pale-yellow coloration (Scharf *et al.*, 2012)). We finally retained 72 queenright colonies for city habitats and 70 colonies for forest habitats. Workers and larvae in each colony were counted. Colony size ranged from 3 to 256 workers and brood from 8 to 96 larvae. Colonies were acclimated for seven days in the laboratory before the experiment began. Using a microscope, we only kept second instar larvae (i.e. larvae under 2mm long with hairs mainly on the front, Plateaux, 1970), so that all larvae would be exposed to cadmium at the same time and for the majority of their development duration.

Experimental design

Workers and larvae had two different origins (city or forest), resulting in four different crosses: city workers with city larvae from another colony (CC, 36 colonies), forest workers with forest larvae from another colony (FF, 36 colonies), city workers with forest larvae (CF, 35 colonies) and forest workers with city larvae (FC, 35 colonies). Colonies were matched by pairs for brood exchange, depending on the number of workers and larvae. Each colony received all the second-instar brood from its paired colony. Larvae of the two paired colonies were exchanged simultaneously, larvae remained isolated for a maximum of 15 minutes without workers. Such brood exchanges have already been performed in other *Temnothorax*

species (Linksvayer, 2006, 2007), and the brood is accepted by workers without discrimination.

Each colony was then exposed to a feeding treatment with two levels: control or cadmium. Colonies were evenly assigned to control or cadmium treatment depending on their colony size, in order to obtain a similar distribution of colony sizes within treatment levels. Control colonies were fed three times a week with a mixture of diluted honey, soya yogurt and dried crickets. Cadmium colonies were fed with the same mixture containing 100µg/g cadmium (Jacquier *et al.*, 2021a). The concentration used in the food corresponds to the LC50 (the concentration that triggers 50% mortality over a given period of time), a physiologically but not ecologically relevant dose. Indeed, this concentration is about fifty times higher than the typical cadmium concentration measured in urban soils. However, in the field, recurrent exposure to pollutants can lead to bioaccumulation and cadmium concentration can reach 102 mg/kg in the ant *Formica polyctena*, for a soil concentration of 2 mg.kg⁻¹ (Starý & Kubizňáková, 1987). Therefore, the concentration that we used was high but still in the upper range of real field values (Jacquier *et al.*, 2021a). We used 15 “CC”, “FC” and “FF” colonies, and 16 colonies “CF” with control food. We assigned more colonies to cadmium treatment in order to compensate for the potential risk of mortality, so we used 21 “CC” and “FF” colonies, 19 “CF” colonies and 20 “FC” colonies with cadmium food.

Colonies were reared in a climatic chamber (reference CTS TP10/600) at 22-27°C – 12h : 12h cycle with natural photoperiod for 64 days, so that larvae could emerge as adults (Molet *et al.*, 2017). Workers that emerged in the laboratory (called “laboratory-born workers”) were identified based on their pale color and collected once they started to move. They were anesthetized by cooling down and stored in 96% ethanol for morphological measurements (see below). Newly-laid eggs were removed every week to prevent the hatching of additional larvae in order to keep the initial number of larvae constant and not to interfere with the experimental larvae that originate from crosses. To do so, we removed the top microscope slide from the nest, took out the eggs, and then let the colony move to a new nest positioned just in front of the old one. This way, we did not need to put a microscope slide back on the nest, which could have crushed some workers.

Adult workers born in the field were removed from the colonies when they died and counted to compute the worker mortality rate. The emergence rate was computed as the ratio between the number of larvae that developed to adulthood over the number of larvae initially given to the colony. For each produced worker, gyne and male, we also measured the

development duration (computed as the number of days from the beginning of the experiment to emergence) as well as the head width (see below).

Size measurement

For morphological measurements, heads were separated from bodies. Heads were stuck on double-sided tape and photographed under a Discovery V12 Zeiss stereomicroscope connected to a computer via a Zeiss AxioCam ICc 5 video camera and measured with ImageJ 1.8 (Abràmoff, Magalhães, & Ram, 2004). Head width is a standard measurement of body size in ants (Tschinkel, Mikheyev, & Storz, 2003).

Statistical analyses

We investigated the effects of worker origin (forest/city), larval origin (forest/city), treatment (cadmium/control) and the initial colony size on the adult mortality rate, the emergence rate, the development duration of produced individuals (sexuals and workers) and the size of produced individuals. We included a second-order interaction between feeding treatment, larval origin and worker origin, that allowed us to test whether the effect of treatment differed across each crossing combination. We also included the interaction between the feeding treatment and the colony size to investigate whether larger colonies were less affected by cadmium than smaller ones.

All statistical analyses were carried out with R v3.6.1 (www.r-project.org). All plots were generated using ggplot2 (Wickham, 2009). First, we assessed the effect of predictor variables on colony-level dependent variables (adult worker mortality rate, emergence rate, and the head width of produced individuals) using generalized linear models (GLMs). Adult worker mortality and emergence were analyzed using quasibinomial distribution. Finally, we assessed the effect of predictor variables on the development duration and the head width of emerged individuals, two individual-level dependent variables, using linear mixed models (LMMs, package *nlme*, (Pinheiro *et al.*, 2016)), with colony as random factor to take into account pseudo-replication. All models are summarized in Table 1.

For each analysis, the minimum adequate model was selected using a backward stepwise approach where explicative variables were removed one by one from a full model based on a log likelihood ratio test. Fisher or χ^2 tests (according to the type of model used) and associated *P*-values for each predictive variable were obtained to obtain the *P*-values for each predictor variable by comparing the minimum adequate model with a model excluding or including the variable of interest (according to whether the variable was present in or

absent from the minimal adequate model respectively). We first tested the significance level of the second-order interaction, then first-order interactions. Normality of the residuals and homogeneity of variances were checked visually following (Pinheiro & Bates, 2000); no transformation of the data was necessary.

Data availability statement

The dataset and R scripts are available on Zenodo (10.5281/zenodo.5105734).

Results

Six queens died during the experiment (one ‘FF’, ‘FC’ and ‘CC’ colony; three ‘CC’ colonies) and their colonies were excluded from data analyses. Our final dataset consisted of 136 queenright colonies. We collected a total of 1 688 workers, 1 092 males, one intercaste (worker phenotype with two ocelli, Okada et al., 2013) and 134 gynes. As we collected gynes from only 15 colonies (13 control colonies: four ‘FF’ colonies, three ‘CC’, ‘CF’ and ‘FC’ colonies; two cadmium colonies: one ‘CF’ and one ‘FF’ colony), we did not include gyne production in our dataset because of a lack of statistical power, with only five gynes under cadmium treatment.

Whatever the variables measured, we did not find a significant second-order interaction (Treatment : Worker origin : Larva origin, Table 1), i.e. there was no differential sensitivity to cadmium depending on the crossing. Moreover, contrary to our expectation, there was no better cadmium tolerance in city colonies: we did not find any interaction between Treatment and Worker origin (at the exception of male development duration, see below) or Treatment : Larval origin interaction (Table 1). Those interactions were therefore removed from the model and in the following results, we only present the main effects of worker origin, larval origin, colony size and treatment, as well as the interaction between colony size and treatment (detailed statistics are presented in Table 1).

Adult mortality rate

Cadmium increased adult mortality rate by 148% (Treatment effect, $F_{131:133} = 92.43$, $P < 0.001$; figure 1a). The origin of workers (city or forest) significantly affected the mortality rate ($F_{131:132} = 7.72$, $P = 0.0060$) with city workers (‘CC’ and ‘CF’) having a 14% higher mortality rate than forest workers (origin of workers effect, $F_{131:132} = 7.72$, $P = 0.0060$; figure 1a). Not surprisingly, there was no effect of the origin of larvae (Table 1). The mortality rate

marginally differed depending on the treatment in interaction with the colony size (Treatment x Colony size interaction, $F_{131:132} = 3.67$, $P = 0.057$). The mortality rate increased with the initial colony size under cadmium, while it decreased with the initial colony size in the control (figure 2).

Emergence rate and development duration

Cadmium decreased emergence rate by 50% (Treatment effect, $F_{131:132} = 107.93$, $P < 0.001$; figure 1b). The origin of workers had no effect on emergence rate ($F_{129:131} = 0.55$, $P = 0.60$; figure 1b) whereas the origin of larvae significantly affected the emergence rate (origin of larvae effect, $F_{131:132} = 7.54$, $P = 0.0070$, figure 1b) with forest larvae having a 17% higher emergence rate than city larvae. There was no effect of colony size on emergence rate (Table 1).

The development of produced individuals was 6% longer under cadmium (Treatment effect, workers: $\chi^2_1 = 7.14$, $P = 0.007$; figure 1c - males: $\chi^2_1 = 64.69$, $P < 0.001$; figure 1d). The origin of workers had no effect on worker development duration whereas the origin of larvae had a significant effect (origin of larvae effect, $\chi^2_1 = 4.22$, $P = 0.039$; Table 1). Forest larvae took 3.1% less time to develop into adult workers than city larvae (figure 1c). We found no effect of colony size on the development duration of larvae (Table 1). Regarding males, we found a marginally significant Treatment : Worker origin interaction ($\chi^2_1 = 3.82$, $P = 0.051$), with longer development under cadmium for males, but not under control. The origin of larvae had no effect on the development duration of males (Table 1). Development duration increased with the initial colony size (Colony size effect, workers: $\chi^2_1 = 6.22$, $P = 0.012$; males: $\chi^2_1 = 5.07$, $P = 0.024$).

Size (head width) of produced individuals

Cadmium reduced the size of the produced workers and males by 14% and 16% respectively (Treatment effect, workers: $\chi^2_1 = 157.97$, $P < 0.001$; figure 3a - males: $\chi^2_1 = 156.24$, $P < 0.001$; figure 3b). Neither worker origin nor larvae origin had an effect on worker head width (Table 1). Colony size had no effect on worker head width (Table 1). Regarding males, forest workers produced marginally larger males relative to city workers (+2.3%, effect of worker origin, $\chi^2_1 = 3.57$, $P = 0.059$; figure 3b). Larvae origin and colony size had no effect on male head width (Table 1).

Discussion

We aimed at disentangling the implications of workers (the social environment) and larvae in the better cadmium tolerance of larvae previously observed in city colonies by Jacquier *et al.* (2021a). Just like other studies (den Besten *et al.*, 1989; Gomot, 1998; Cervera *et al.*, 2004; Demenesku *et al.*, 2014; Jacquier *et al.*, 2021a), we found that cadmium had a negative impact on all life history traits measured, i.e. it increased worker mortality rate (+148%), reduced emergence rate (-50%), extended development duration (+6%) and reduced the size of individuals, for both workers (-14%) and males (-16%). The intensity of the effect on workers was of the same order of magnitude as in Jacquier *et al.* (2021a). However, in contrast with Jacquier *et al.* (2021a), the effect of cadmium was similar in city and forest colonies regarding emergence rate and worker head width, preventing us from disentangling the effect of workers and larvae on the expected differential cadmium tolerance. This absence of differential response between habitats could result from annual random variation and the fact that a comparison between only two sites did not allow us to detect a difference in our study. However, our sample size in terms of number of colonies was large, excluding a lack of statistical power. In addition, the differential response between the two habitats was observed for five replicated comparisons (Jacquier *et al.*, 2021a,b) making chance an unlikely explanation. Also, as the same persons performed the experiments, with the same laboratory set up (even though in our experiment larvae were manipulated which was not the case in Jacquier *et al.* (2021a) – see below), we think that the observed difference is a true biological phenomenon.

We propose three non-mutually exclusive hypotheses that could explain the unexpected absence of a differential cadmium tolerance between forest and city colonies. A first explanation is a disruptive effect of cross-fostering. This experimental method is commonly used to split the relative contributions of different social aspects on measured traits in ants (Linksvayer, 2006; Armitage *et al.*, 2011; Purcell & Chapuisat, 2012). Therefore, adult workers had to rear larvae from foreign colonies. Because workers are unrelated to these larvae, they may invest less in rearing them. However, Foitzik *et al.*, (2007) showed that in *Temnothorax nylanderi* there was little to no discrimination of non-nestmates, maybe because colonies and populations are weakly differentiated genetically and thus could have similar cuticular hydrocarbons, therefore impeding worker discrimination based on this cue. Besides, as in other successful cross-fostering experiments (Linksvayer, 2006; Armitage *et al.*, 2011; Purcell & Chapuisat, 2012), larvae were largely accepted by workers at the beginning of our

experiment (more than 90% of the brood survived after the larval exchange between colonies), so workers are probably not able to detect that larvae originated from another colony. A second hypothesis is a within year variation of cadmium tolerance. We collected colonies in April, whereas (Jacquier *et al.*, 2021a) collected them in February. In some species, trace metals show a seasonal pattern of accumulation, correlated with seasonal variation in detoxification enzymes or proteins (Fialkowski *et al.*, 2003; Gorbi, Baldini, & Regoli, 2005). Recent data suggest that the better cadmium tolerance of city colonies depends on both hibernation temperature and the period of the year at which the experiment was conducted (Jacquier *et al.*, 2021b; Jacquier *et al.*, *under review*). In this study, we collected colonies later (April) than in (Jacquier *et al.*, 2021a), in order to obtain more sexuals. This may have impeded further differential cadmium tolerance between city and forest colonies, as Jacquier *et al.* (*under review*) showed that colonies collected several weeks after hibernation (and therefore with eggs that were laid after the hibernation period, i.e. early June here) do not have differential emergence rates in response to cadmium. Further studies are needed to investigate the possible effect of within-year variation in environmental factors on cadmium tolerance between the two habitats, since urban and forest ants have a slightly shifted phenology (Chick *et al.*, 2019). A last hypothesis stands in inter-annual variation in cadmium tolerance. In Jacquier *et al.* (2021a), urban colonies tolerated cadmium better than forest colonies, but the level of differential tolerance varied among sites and years. In this study, we only used one study site, therefore we lack replication and the observed results could be due to natural variation in cadmium tolerance at this site for this particular year. Moreover, Jacquier *et al.* (2021b) suggested that cadmium tolerance could also depend on environmental factors such as hibernation temperature. Colonies may have experienced special climatic conditions the year of our experiment, and therefore we found results that contrast with previous findings.

Our study quantified the effect of cadmium on male production, and therefore the potential consequences on the reproductive success of colonies. In *T. nylanderi*, brood production is highly synchronized (Kipyatkov, 1993), but males develop faster than workers (our study and (Jacquier *et al.*, 2021a). As we started our study later (April vs February – March for Jacquier *et al.* 2021a) and winter was milder, colonies were probably more advanced in the process of male production. The relative lack of gyne production is not surprising, given that resource allocation can be redirected towards workers and males after a large worker loss (29% on average for control colonies, 72% for cadmium colonies at the end of the experiment) in this species (Foitzik & Heinze, 2000; Foitzik *et al.*, 2003; Honorio,

Doums, & Molet, 2020). Forest workers reared males that were 1.7% larger than those reared by city workers. Thus, when forest workers reared larvae, they produced larger males whatever the origin of larvae and whatever the treatment. Larger males have greater mating success and sperm production in ants (Abell *et al.*, 1999; Wiernasz *et al.*, 2001; Brown & Keller, 2006), therefore forest colonies may have a higher fitness than city colonies whatever the treatment, suggesting a better general condition. As adult city workers unexpectedly faced a higher mortality (in contrast with Jacquier *et al.* (2021a), a hypothesis is that city workers may have undergone higher stress levels, maybe leading to less brood care.

Different studies documented phenotypic divergence between city and rural insect populations. For example, body size is sometimes higher (Weller & Ganzhorn, 2004; y Gomez & Van Dyck, 2012; Lowe, Wilder, & Hochuli, 2014; Merckx *et al.*, 2018; Theodorou *et al.*, 2021) and sometimes lower (Merckx *et al.*, 2018; Eggenberger *et al.*, 2019) in city populations depending on the species. Interestingly, in *Temnothorax* species, previous studies found no differences in worker size across urban and rural populations (Yilmaz *et al.*, 2019; Jacquier *et al.*, 2021a). Emergence rate (grasshoppers: Gomez and Van Dyck, 2012; damselfly: Tüzün *et al.*, 2017) and development duration in the field also differs between urban and rural populations depending on the species (mosquitoes: Li *et al.*, 2014, damselfly: Tüzün *et al.*, 2017). In this study, we found a limited effect of the origin of larvae, independently of the feeding treatment, and no effect of the social environment, except for male size. Forest larvae had a 17% higher emergence rate and a 3.1% shorter development duration but a 2.3% larger size at emergence when they developed into workers, therefore appearing to be of better quality (even though worker size does not seem to be of primary value for colony fitness in this species (Colin *et al.*, 2017; Honorio *et al.*, 2020, 2021 - but see Modlmeier *et al.*, 2013). Our study does not allow to further substantiate the processes involved, but two potential proximate mechanisms can be proposed. First, city colonies may have had a poor physiological and metabolic state at the beginning of the experiment because they faced multiple stressors related to their urban environment prior to the experiment. Therefore, city queens could lay lower quality eggs, with larvae that develop slower and undergo lower emergence rate. Differences in egg provisioning though maternal or epigenetic effects can generate differences in development and emergence rate between larvae. For instance, juvenile hormone and ecdysteroid are transmitted to brood and influence development and emergence rate in ants (Schwander *et al.*, 2008; Cahan, Graves, & Brent, 2011; Libbrecht *et al.*, 2013) - see also (LeBoeuf *et al.*, 2016, 2018). Second, the differences between city and forest colonies could be genetically determined, as a recent study found a

few discriminating loci (Khimoun *et al.*, 2020). However, this study also highlighted a lack of genetic differentiation between urban and forest populations.

Another important social trait is colony size. Stress tolerance increases with colony size in many eusocial species (Naug, 2009; Scharf *et al.*, 2012; Modlmeier *et al.*, 2013; Crall *et al.*, 2019). Here, we found no buffering effect of colony size with the exception of a contrasting effect on worker mortality rate (positive relationship under control treatment, but negative under cadmium) and a positive relationship with the development duration of workers and males. This is surprising, as other studies on *T. nylander* showed that larger colonies better resist starvation (Modlmeier *et al.*, 2013), parasites (Scharf *et al.*, 2012) or high temperatures (Molet *et al.*, 2017). (Jacquier *et al.*, 2021a) did not find any buffering effect of colony size on worker mortality, whatever the treatment. This suggests that the buffering effect of colony size in *Temnothorax nylander* is quite unstable and may depend on the measured traits and on the time at which the experiment is conducted. The absence of buffering effect of colony size under cadmium in our study could be explained by a higher sensitivity of large colonies to disturbances. Having more individuals could enhance the exposures to various stressors (pathogens, parasites, pollutions, etc.) and their transmission to nestmates. For example, in honey bees, larger colonies could be more sensitive to sublethal effects of pesticides and pathogens and undergo colony collapse disorder (Barron, 2015). When the disturbance is too high, larger colonies could be more destabilized and collapse. Thus, social organisms could be more exposed and vulnerable to environmental stressors than solitary organisms. Finally, the positive relationship between colony size and the development duration of workers and males is consistent with the loss of per-capita productivity as a function of colony size, found in many ant species (reviewed in Kramer *et al.*, 2014).

To conclude, we did not find any difference in cadmium tolerance between forest and city colonies, so we were not able to assess the relative contributions of social environment and larval intrinsic factors. Replicating this study with colonies collected at the end of winter (the period at which a differential response to cadmium was found between forest and city colonies) would be necessary to assess it. However, interestingly, we found phenotypic differences between forest and city colonies, independently of the treatment, with forest larvae having a higher emergence rate and turning into larger workers. These differences were independent of the social environment, i.e. workers, contrasting with previous studies on other social insect species (Pankiw, Tarpy, & Page, 2002; Linksvayer, 2006, 2007; Linksvayer *et al.*, 2011; Armitage *et al.*, 2011). However, regarding males, we found a slight effect of the social environment, suggesting a higher social control on the phenotype of sexuals than on

that of workers. We suggest that more attention should be paid to the role of the social environment in phenotypic divergence. Since our studies only involved two sites and one species, the generalization of our results will require the assessment of additional locations and species.

Acknowledgments

We thank Romain Péronnet for technical assistance and ant collection. We thank the three anonymous reviewers for their helpful comments on the manuscript.

Author's contributions

LJ and RH designed the study, collected and reared ants, performed the experiment and statistical analyses, and wrote the manuscript. CD designed the study, contributed to statistical analyses and wrote the manuscript. MM designed the study, wrote the manuscript and supervised the project. All authors read and approved the final manuscript.

Funding

This work was funded by the Institute of Ecology and Environmental Sciences - Paris and the Institut de la Transition Environnementale de Sorbonne Université.

Compliance with ethical standards

Conflict of interest

The authors declare that they have no conflict of interest.

Ethical approval

All applicable international, national and/or institutional guidelines for the care and use of animals were followed.

References

- Abell AJ, Cole BJ, Reyes R & Wiernasz DC. 1999.** Sexual selection on body size and shape in the western harvester ant, *Pogonomyrmex occidentalis* Cresson. *Evolution* **53**: 535–545.
- Abràmoff MD, Magalhães PJ & Ram SJ. 2004.** Image processing with imageJ. *Biophotonics International* **11**: 36–41.
- Alberti M, Marzluff J & Hunt VM. 2017.** Urban driven phenotypic changes: Empirical observations and theoretical implications for eco-evolutionary feedback. *Philosophical Transactions of the Royal Society B: Biological Sciences* **372**.
- Armitage SAO, Broch JF, Marín HF, Nash DR & Boomsma JJ. 2011.** Immune defense in leaf-cutting ants: a cross-fostering approach. *Evolution; International Journal of Organic Evolution* **65**: 1791–1799.
- Barron AB. 2015.** Death of the bee hive: Understanding the failure of an insect society. *Current Opinion in Insect Science* **10**: 45–50.
- den Besten PJ, Herwig HJ, Zandee DI & Voogt PA. 1989.** Effects of cadmium and PCBs on reproduction of the sea star *Asterias rubens*: Aberrations in the early development. *Ecotoxicology and Environmental Safety* **18**: 173–180.
- Brans KI, Jansen M, Vanoverbeke J, Tüzün N, Stoks R & De Meester L. 2017.** The heat is on: Genetic adaptation to urbanization mediated by thermal tolerance and body size. *Global Change Biology* **23**: 5218–5227.
- Brown WD & Keller L. 2006.** Resource supplements cause a change in colony sex-ratio specialization in the mound-building ant, *Formica exsecta*. *Behavioral Ecology and Sociobiology* **60**: 612–618.
- Cahan SH, Graves CJ & Brent CS. 2011.** Intergenerational effect of juvenile hormone on offspring in *Pogonomyrmex* harvester ants. *Journal of Comparative Physiology B: Biochemical, Systemic, and Environmental Physiology* **181**: 991–999.
- Cervera A, Maymó AC, Sendra M, Martínez-Pardo R & Garcera MD. 2004.** Cadmium effects on development and reproduction of *Oncopeltus fasciatus* (Heteroptera: Lygaeidae). *Journal of Insect Physiology* **50**: 737–749.
- Chick LD, Strickler SA, Perez A, Martin RA & Diamond SE. 2019.** Urban heat islands advance the timing of reproduction in a social insect. *Journal of Thermal Biology* **80**: 119–125.
- Colin T, Doums C, Péronnet R & Molet M. 2017.** Decreasing worker size diversity does not affect colony performance during laboratory challenges in the ant *Temnothorax nylanderi*. *Behavioral Ecology and Sociobiology* **71**: 92.
- Crall JD, de Bivort BL, Dey B & Versypt ANF. 2019.** Social buffering of pesticides in bumblebees: Agent-based modeling of the effects of colony size and neonicotinoid exposure on behavior within nests. *Frontiers in Ecology and Evolution* **7**.

472 **Dahirel M, De Cock M, Vantieghem P & Bonte D. 2019.** Urbanization-driven changes in
473 web building and body size in an orb web spider. *Journal of Animal Ecology* **88**: 79–91.

474 **Demenesku J, Mirkov I, Ninkov M, Popov Aleksandrov A, Zolotarevski L,**
475 **Kataranovski D & Kataranovski M. 2014.** Acute cadmium administration to rats exerts
476 both immunosuppressive and proinflammatory effects in spleen. *Toxicology* **326**: 96–108.

477 **Diamond SE, Chick L, Perez A, Strickler SA & Martin RA. 2017.** Rapid evolution of ant
478 thermal tolerance across an urban-rural temperature cline. *Biological Journal of the Linnean*
479 *Society* **121**: 248–257.

480 **Diamond SE & Martin RA. 2020.** Evolutionary Consequences of the Urban Heat Island.
481 Urban Evolutionary Biology, 91-110. Oxford University Press.

482 **Ditchkoff SS, Saalfeld ST & Gibson CJ. 2006.** Animal behavior in urban ecosystems:
483 Modifications due to human-induced stress. *Urban Ecosystems* **9**: 5–12.

484 **Eggenberger H, Frey D, Pellissier L, Ghazoul J, Fontana S & Moretti M. 2019.** Urban
485 bumblebees are smaller and more phenotypically diverse than their rural counterparts. *Journal*
486 *of Animal Ecology* **88**: 1522–1533.

487 **Evans KL, Newton J, Gaston KJ, Sharp SP, McGowan A & Hatchwell BJ. 2012.**
488 Colonisation of urban environments is associated with reduced migratory behaviour,
489 facilitating divergence from ancestral populations. *Oikos* **121**: 634–640.

490 **Fialkowski W, Rainbow PS, Smith BD & Zmudzinski L. 2003.** Seasonal variation in trace
491 metal concentrations in three talitrid amphipods from the Gulf of Gdansk, Poland. *Journal of*
492 *Experimental Marine Biology and Ecology* **288**: 81–93.

493 **Fjerdingstad EJ & Keller L. 2004.** Relationships between phenotype, mating behavior, and
494 fitness of queens in the ant *Lasius niger*. *Evolution* **58**: 1056–1063.

495 **Foitzik S, Sturm H, Pusch K, D’Ettorre P & Heinze J. 2007.** Nestmate recognition and
496 intraspecific chemical and genetic variation in *Temnothorax* ants. *Animal Behaviour* **73**: 999–
497 1007.

498 **Foitzik S & Heinze J. 2000.** Intraspecific parasitism and split sex ratios in a monogynous and
499 monandrous ant (*Leptothorax nylanderi*). *Behavioral Ecology and Sociobiology* **47**: 424–431.

500 **Foitzik S, Strätz M & Heinze J. 2003.** Ecology, life history and resource allocation in the
501 ant, *Leptothorax nylanderi*. *Journal of Evolutionary Biology* **16**: 670–680.

502 **Foti L, Dubs F, Gignoux J, Lata JC, Lerch TZ, Mathieu J, Nold F, Nunan N, Raynaud**
503 **X, Abbadie L & Barot S. 2017.** Trace element concentrations along a gradient of urban
504 pressure in forest and lawn soils of the Paris region (France). *Science of the Total*
505 *Environment* **598**: 938–948.

506 **y Gomez GSM & Van Dyck H. 2012.** Ecotypic differentiation between urban and rural
507 populations of the grasshopper *Chorthippus brunneus* relative to climate and habitat
508 fragmentation. *Oecologia* **169**: 125–133.

509 **Gomot A. 1998.** Toxic effects of cadmium on reproduction, development, and hatching in the
510 freshwater snail *Lymnaea stagnalis* for water quality monitoring. *Ecotoxicology and*
511 *Environmental Safety* **41**: 288–297.

512 **Gorbi S, Baldini C & Regoli F. 2005.** Seasonal Variability of Metallothioneins, Cytochrome
513 P450, Bile Metabolites and Oxyradical Metabolism in the European Eel *Anguilla anguilla* L.
514 (*Anguillidae*) and Striped Mullet *Mugil cephalus* L. (*Mugilidae*). *Archives of Environmental*
515 *Contamination and Toxicology* **49**: 62–70.

516 **Grimm NB, Faeth SH, Golubiewski NE, Redman CL, Wu J, Bai X & Briggs JM. 2008.**
517 Global change and the ecology of cities. *Science* **319**: 756–760.

518 **Hasegawa M, Ligon RA, Giraudeau M, Watanabe M & McGraw KJ. 2014.** Urban and
519 colorful male house finches are less aggressive. *Behavioral Ecology* **25**: 641–649.

520 **Heinze J, Stahl M & Hölldobler B. 1996.** Ecophysiology of hibernation in boreal
521 *Leptothorax* ants (Hymenoptera: Formicidae). *Ecoscience* **3**: 429–435.

522 **Honorio R, Doums C & Molet M. 2020.** Manipulation of worker size diversity does not
523 affect colony fitness under natural conditions in the ant *Temnothorax nylander*. *Behavioral*
524 *Ecology and Sociobiology* **74**: 104.

525 **Honorio R, Doums C & Molet M. 2021.** Worker Size Diversity Has No Effect on
526 Overwintering Success under Natural Conditions in the Ant *Temnothorax nylander*. *Insects*
527 **12**: 379.

528 **Jacquier L, Doums C, Four-Chaboussant A, Peronnet R, Tirard C & Molet M. 2021a.**
529 Urban colonies are more resistant to a trace metal than their forest counterparts in the ant
530 *Temnothorax nylander*. *Urban Ecosystems* **24**: 561–570.

531 **Jacquier L, Molet M, Bocquet C & Doums C. 2021b.** Hibernation Conditions Contribute to
532 the Differential Resistance to Cadmium between Urban and Forest Ant Colonies. *Animals* **11**:
533 1050.

534 **Järup L & Åkesson A. 2009.** Current status of cadmium as an environmental health
535 problem. *Toxicology and Applied Pharmacology* **238**: 201–208.

536 **Jochner S & Menzel A. 2015.** Urban phenological studies - Past, present, future.
537 *Environmental Pollution* **203**: 250–261.

538 **Johnson MTJ & Munshi-South J. 2017.** Evolution of life in urban environments. *Science*
539 **358**.

540 **Kaspari M & Vargo EL. 1995.** Colony size as a buffer against seasonality: Bergmann's rule
541 in social insects. *American Naturalist* **145**: 610–632.

542 **Keiser CN, Wright CM, Singh N, DeShane JA, Modlmeier AP & Pruitt JN. 2015.** Cross-
543 fostering by foreign conspecific queens and slave-making workers influences individual- and
544 colony-level personality. *Behavioral Ecology and Sociobiology* **69**: 395–405.

545 **Khimoun A, Doums C, Molet M, Kaufmann B, Peronnet R, Eyer PA & Mona S. 2020.**
546 Urbanization without isolation: The absence of genetic structure among cities and forests in
547 the tiny acorn ant *Temnothorax nylanderii*. *Biology Letters* **16**: 20190741.

548 **Kipyatkov VE. 1993.** Annual cycles of development in ants: diversity, evolution, regulation.
549 In: *Proceedings of the colloquia on social insects*. Russian-Speaking Section of the IUSI,
550 25–48.

551 **Kramer BH, Scharf I & Foitzik S. 2014.** The role of per-capita productivity in the evolution
552 of small colony sizes in ants. *Behavioral Ecology and Sociobiology* **68**: 41–53.

553 **LeBoeuf AC, Waridel P, Brent CS, Gonçalves AN, Menin L, Ortiz D, Riba-Grognuz O,**
554 **Koto A, Soares ZG, Privman E, Miska EA, Benton R & Keller L. 2016.** Oral transfer of
555 chemical cues, growth proteins and hormones in social insects. *eLife* **5**: 1–28.

556 **LeBoeuf AC, Cohan AB, Stoffel C, Brent CS, Waridel P, Privman E, Keller L &**
557 **Benton R. 2018.** Molecular evolution of juvenile hormone esterase-like proteins in a socially
558 exchanged fluid. *Scientific Reports* **8**: 1–10.

559 **Levermore G, Parkinson J, Lee K, Laycock P & Lindley S. 2018.** The increasing trend of
560 the urban heat island intensity. *Urban Climate* **24**: 360–368.

561 **Li Y, Kamara F, Zhou G, Puthiyakunnon S, Li C, Liu Y, Zhou Y, Yao L, Yan G & Chen**
562 **XG. 2014.** Urbanization Increases *Aedes albopictus* Larval Habitats and Accelerates
563 Mosquito Development and Survivorship. *PLOS Neglected Tropical Diseases* **8**: e3301.

564 **Libbrecht R, Corona M, Wende F, Azevedo DO, Serrão JE & Keller L. 2013.** Interplay
565 between insulin signaling, juvenile hormone, and vitellogenin regulates maternal effects on
566 polyphenism in ants. *Proceedings of the National Academy of Sciences of the United States of*
567 *America* **110**: 11050–11055.

568 **Linksvayer TA. 2006.** Direct, maternal, and subsocial genetic effects on individual and
569 colony traits in an ant. *Evolution* **60**: 2552.

570 **Linksvayer TA. 2007.** Ant species differences determined by epistasis between brood and
571 worker genomes. *PLoS ONE* **2**.

572 **Linksvayer TA, Kaftanoglu O, Akyol E, Blatch S, Amdam G V. & Page RE. 2011.** Larval
573 and nurse worker control of developmental plasticity and the evolution of honey bee queen–
574 worker dimorphism. *Journal of Evolutionary Biology* **24**.

575 **Lowe EC, Wilder SM & Hochuli DF. 2014.** Urbanisation at Multiple Scales Is Associated
576 with Larger Size and Higher Fecundity of an Orb-Weaving Spider. *PLOS ONE* **9**: e105480.

577 **Malakar C, Ganguly A & Haldar P. 2009.** Influence of cadmium on growth, survival and
578 clutch size of a common Indian short horned grasshopper, *Oxya fuscovittata*. *Am. J. Toxicol.*
579 *Sci* **1**: 32–36.

580 **Merckx T, Souffreau C, Kaiser A, Baardsen LF, Backeljau T, Bonte D, Brans KI, Cours**
581 **M, Dahirel M, Debortoli N, Wolf KD, Engelen JMT, Fontaneto D, Gianuca AT, Govaert**
582 **L, Hendrickx F, Higuti J, Lens L, Martens K, Matheve H, Matthysen E, Piano E, Sablon**

583 **R, Schön I, Doninck KV, Meester LD & Dyck HV. 2018.** Body-size shifts in aquatic and
584 terrestrial urban communities. *Nature* **558**: 113.

585 **Modlmeier AP, Foitzik S & Scharf I. 2013.** Starvation endurance in the ant *Temnothorax*
586 *nylanderi* depends on group size, body size and access to larvae. *Physiological Entomology*
587 **38**: 89–94.

588 **Molet M, Péronnet R, Couette S, Canovas C & Doums C. 2017.** Effect of temperature and
589 social environment on worker size in the ant *Temnothorax nylanderi*. *Journal of Thermal*
590 *Biology* **67**: 22–29.

591 **Naug D. 2009.** Structure and resilience of the social network in an insect colony as a function
592 of colony size. *Behavioral Ecology and Sociobiology* **63**: 1023–1028.

593 **Okada Y, Plateaux L & Peeters C. 2013.** Morphological variability of intercastes in the ant
594 *Temnothorax nylanderi*: pattern of trait expression and modularity. *Insectes sociaux* **60**: 319–
595 328.

596 **Pankiw T, Tarpy DR & Page RE. 2002.** Genotype and rearing environment affect honeybee
597 perception and foraging behaviour. *Animal Behaviour* **64**: 663–672.

598 **Partecke J, Schwabl I & Gwinner E. 2006.** Stress and the city: Urbanization and its effects
599 on the stress physiology in European Blackbirds. *Ecology* **87**: 1945–1952.

600 **Pinheiro JC, Bates DM, DebRoy S & Sarkar D. 2016.** R Core Team (2016) nlme: linear
601 and nonlinear mixed effects models. R package version 3.1-128.

602 **Pinheiro JC & Bates DM. 2000.** Linear mixed-effects models: basic concepts and examples.
603 *Mixed-effects models in S and S-Plus*: 3–56.

604 **Plateaux L. 1970.** Sur le polymorphisme social de la fourmi *Leptothorax nylanderi* (Förster).
605 I. Morphologie et biologie comparées des castes. *Annales des Sciences Naturelles, Zoologie et*
606 *Biologie Animale* **12**: 373–478.

607 **Prosser C, Hudson S & Thompson MB. 2006.** Effects of Urbanization on Behavior,
608 Performance, and Morphology of the Garden Skink, *Lampropholis guichenoti*. *Journal of*
609 *Herpetology* **40**: 151–159.

610 **Purcell J & Chapuisat M. 2012.** The influence of social structure on brood survival and
611 development in a socially polymorphic ant: insights from a cross-fostering experiment.
612 *Journal of Evolutionary Biology* **25**: 2288–2297.

613 **Rueppell O & Kirkman RW. 2005.** Extraordinary starvation resistance in *Temnothorax*
614 *rugatulus* (Hymenoptera, Formicidae) colonies: Demography and adaptive behavior. *Insectes*
615 *Sociaux* **52**: 282–290.

616 **Scharf I, Modlmeier AP, Beros S & Foitzik S. 2012.** Ant societies buffer individual-level
617 effects of parasite infections. *American Naturalist* **180**: 671–683.

618 **Schwander T, Humbert JY, Brent CS, Cahan SH, Chapuis L, Renai E & Keller L. 2008.**
619 Maternal Effect on Female Caste Determination in a Social Insect. *Current Biology* **18**: 265–
620 269.

621 **Starý P & Kubizňáková J. 1987.** Content and transfer of heavy metal air pollutants in
622 populations of *Formica* spp. wood ants (Hym., Formicidae). *Journal of Applied Entomology*
623 **104**: 1–10.

624 **Straub L, Williams GR, Pettis J, Fries I & Neumann P. 2015.** Superorganism resilience:
625 Eusociality and susceptibility of ecosystem service providing insects to stressors. *Current*
626 *Opinion in Insect Science* **12**: 109–112.

627 **Theodorou P, Baltz LM, Paxton RJ & Soro A. 2021.** Urbanization is associated with shifts
628 in bumblebee body size, with cascading effects on pollination. *Evolutionary Applications* **14**:
629 53–68.

630 **Tschinkel WR, Mikheyev AS & Storz SR. 2003.** Allometry of Workers of the Fire Ant,
631 *Solenopsis invicta*. *Journal of Insect Science* **3**: 1–11.

632 **Tüzün N, Beeck LO de, Brans KI, Janssens L & Stoks R. 2017.** Microgeographic
633 differentiation in thermal performance curves between rural and urban populations of an
634 aquatic insect. *Evolutionary Applications* **10**: 1067–1075.

635 **Wagner D & Gordon DM. 1999.** Colony age, neighborhood density and reproductive
636 potential in harvester ants. *Oecologia* **119**: 175–182.

637 **Weller B & Ganzhorn JU. 2004.** Carabid beetle community composition, body size, and
638 fluctuating asymmetry along an urban-rural gradient. *Basic and Applied Ecology* **5**: 193–201.

639 **Westby KM & Medley KA. 2020.** Cold nights, city lights: Artificial light at night reduces
640 photoperiodically induced diapause in urban and rural populations of *aedes albopictus*
641 (Diptera: Culicidae). *Journal of Medical Entomology* **57**: 1694–1699.

642 **Wickham H. 2009.** Ggplot2: Elegant graphics for data analysis. *Springer-Verlag, New-York*.

643 **Wiernasz DC, Sater AK, Abell AJ & Cole BJ. 2001.** Male Size, Sperm Transfer, and
644 Colony Fitness in the Western Harvester Ant, *Pogonomyrmex Occidentalis*. *Evolution* **55**:
645 324.

646 **Wiernasz DC & Cole BJ. 2003.** Queen size mediates queen survival and colony fitness in
647 harvester ants. *Evolution* **57**: 2179–2183.

648 **Wills BD, Powell S, Rivera MD & Suarez A V. 2018.** Correlates and Consequences of
649 Worker Polymorphism in Ants. *Annual Review of Entomology* **63**: 575–598.

650 **Yilmaz AR, Chick LD, Perez A, Strickler SA, Vaughn S, Martin RA & Diamond SE.**
651 **2019.** Remarkable insensitivity of acorn ant morphology to temperature decouples the
652 evolution of physiological tolerance from body size under urban heat islands. *Journal of*
653 *Thermal Biology* **85**: 102426.

654

Figures & table

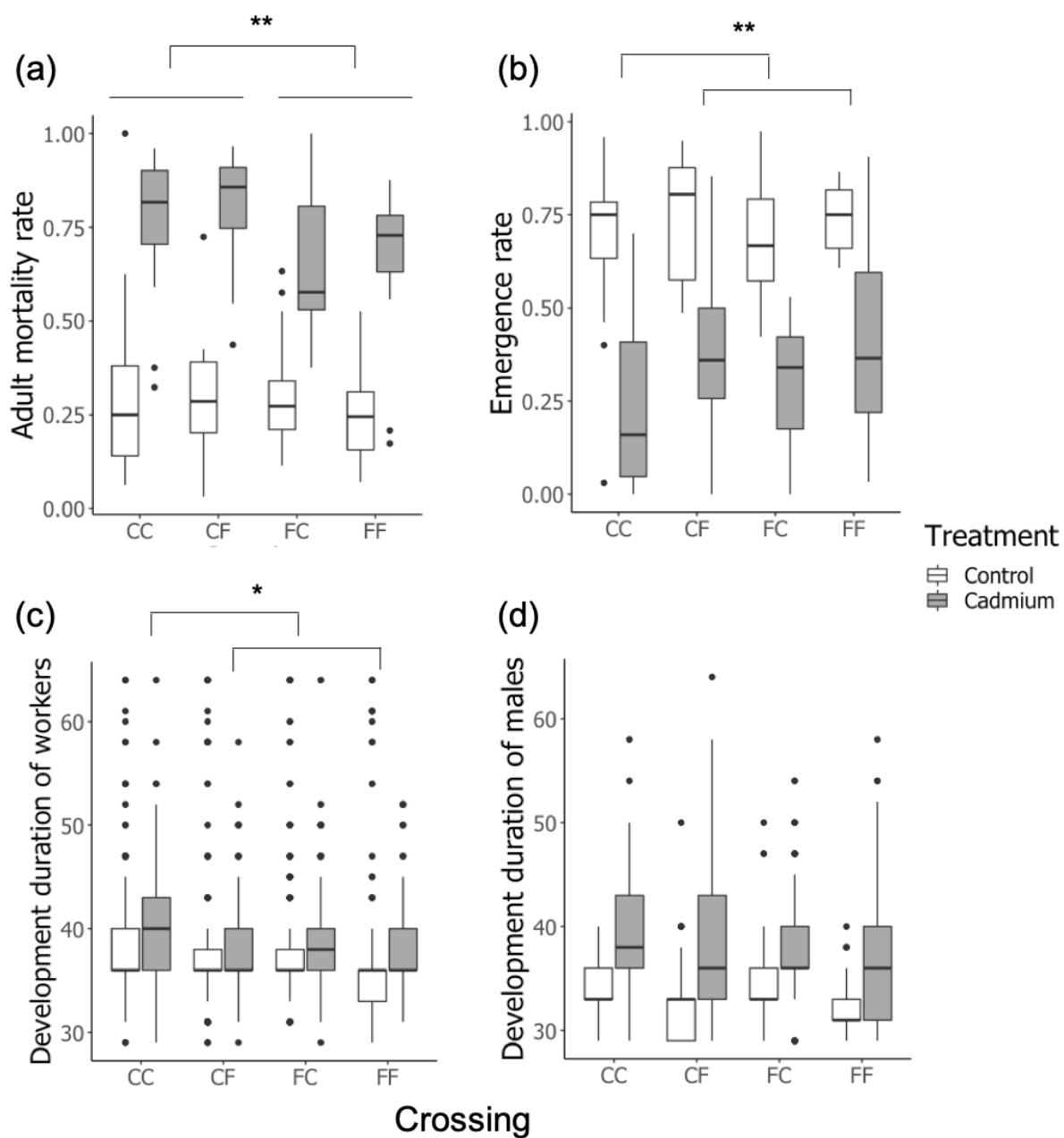


Figure 1: Boxplots comparing the effects of the worker and larval origins (crossing) and feeding treatment on (a) the adult worker mortality rate, (b) the larval emergence rate, the development duration of laboratory-born (c) workers and (d) males. Boxes show median, quartiles and extremes. Significance is indicated by ** $p < 0.01$ and * $p < 0.05$. Feeding had a significant effect on all variables (not shown in the figure for clarity). Statistics are presented in Table 1.

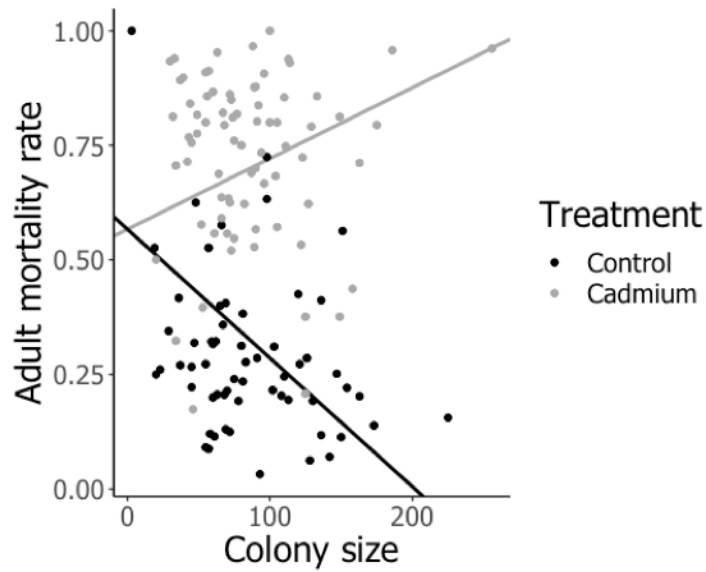


Figure 2: Correlation between the initial colony size and the adult worker mortality rate. When the initial colony size increased, the adult mortality increased for cadmium treatment and decreased for control treatment (black line: control, estimate = $-2.81e^{-3} \pm 0.43e^{-3}$, grey line: cadmium, estimate = $1.54e^{-3} \pm 0.42e^{-3}$).

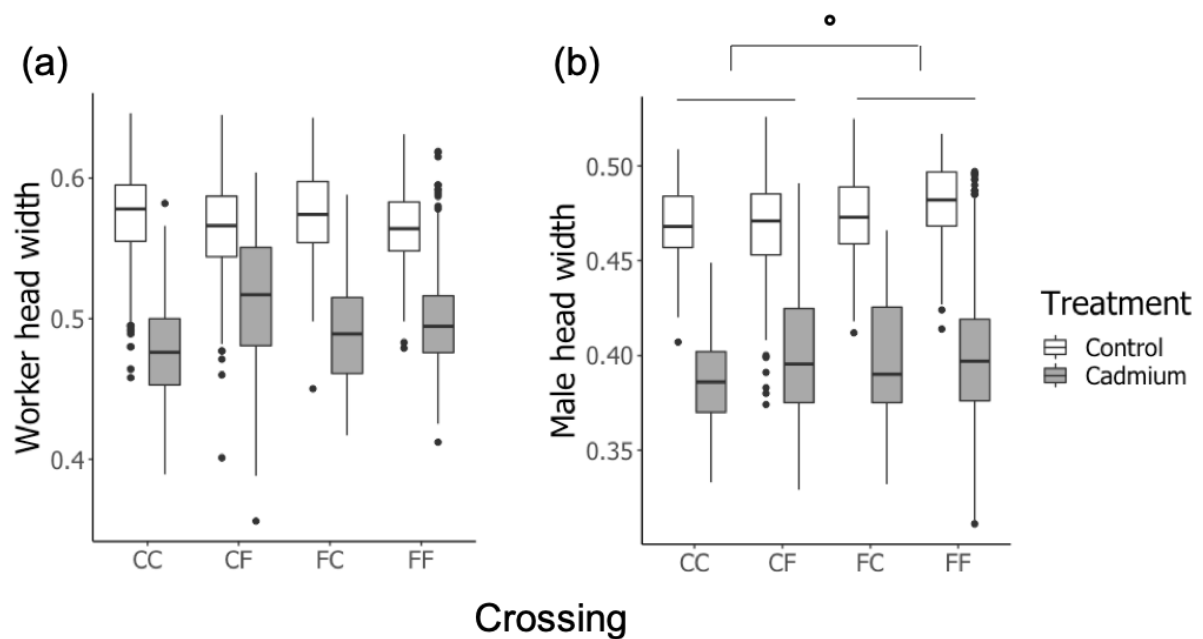


Figure 3: Boxplots comparing the effects of the worker and larval origins (crossing) and treatment on the head width (mm) of laboratory-born (a) workers and (b) males. Boxes show median, quartiles and extremes. Trend ($0.05 < p < 0.06$) is indicated by °. Treatment had a significant effect on all variables (not shown in the figure for clarity). Statistics are presented in Table 1.

Table 1: Models and statistics for the different response variables. “Minimum model” means that the predictor was retained in the minimum model selected. A mixed effect model with colony as random factor was used for the development duration and head width of produced individuals, in order to control for the origin of workers. Significant effects are in bold.

Predictors	Response variables					
	Adult mortality rate	Emergence rate	Worker development duration	Male development duration	Worker head width	Male head width
Treatment	Minimum model $F_{131:133} = 92.43$, $P < 0.001$	Minimum model $F_{131:132} = 107.93$, $P < 0.001$	Minimum model $\chi^2_1 = 7.14$, $P = 0.0075$	Minimum model $\chi^2_1 = 64.69$, $P < 0.001$	Minimum model $\chi^2_1 = 157.97$, $P < 0.001$	Minimum model $\chi^2_1 = 156.24$, $P < 0.001$
Worker origin	Minimum model $F_{131:132} = 7.72$, $P = 0.0060$	$F_{130:131} = 0.81$, $P = 0.37$	$\chi^2_1 = 0.46$, $P = 0.50$	$\chi^2_1 = 0.56$, $P = 0.45$	$\chi^2_1 = 0.57$, $P = 0.45$	Minimum model $\chi^2_1 = 3.57$, $P = 0.059$
Worker origin - Treatment	$F_{130:131} = 2.96$, $p = 0.087$	$F_{129:131} = 0.60$, $P = 0.55$	$\chi^2_2 = 2.00$, $P = 0.37$	Minimum model $\chi^2_1 = 3.82$, $P = 0.05$	$\chi^2_2 = 0.62$, $P = 0.73$	$\chi^2_1 = 0.027$, $P = 0.87$
Colony size	Minimum model $F_{131:133} = 2.77$, $P = 0.066$	$F_{130:131} = 1.65$, $P = 0.20$	Minimum model $\chi^2_1 = 6.22$, $P = 0.012$	Minimum model $\chi^2_1 = 5.61$, $P = 0.017$	$\chi^2_1 = 1.01$, $P = 0.31$	$\chi^2_1 = 0.22$, $P = 0.64$
Colony size - Treatment	Minimum model , $F_{131:132} = 3.67$, $P = 0.057$	$F_{129:131} = 2.28$, $P = 0.10$	$\chi^2_1 = 0.01$, $P = 0.91$	$\chi^2_1 = 0.75$, $P = 0.38$	$\chi^2_2 = 3.59$, $P = 0.16$	$\chi^2_2 = 0.24$, $P = 0.89$
Larva origin	$F_{131,130} = 0.0054$ $P = 0.94$	Minimum model $F_{131:132} = 7.54$, $P = 0.007$	Minimum model $\chi^2_1 = 4.22$, $P = 0.039$	$\chi^2_1 = 2.53$, $P = 0.11$	$\chi^2_1 = 0.41$, $P = 0.52$	$\chi^2_1 = 0.07$, $P = 0.80$
Larva origin - Treatment	$F_{131,129} = 0.14$ $P = 0.86$	$F_{130:131} = 0.91$, $P = 0.34$	$\chi^2_1 = 1.21$, $P = 0.27$	$\chi^2_2 = 2.9$, $P = 0.23$	$\chi^2_2 = 2.57$, $P = 0.28$	$\chi^2_2 = 0.30$, $P = 0.86$
Second-order interaction worker origin - larva origin - Treatment	$F_{131,126} = 0.82$ $P = 0.53$	$F_{126:131} = 0.45$, $P = 0.81$	$\chi^2_5 = 3.70$, $P = 0.59$	$\chi^2_4 = 4.38$, $P = 0.36$	$\chi^2_6 = 6.01$, $P = 0.42$	$\chi^2_5 = 2.78$, $P = 0.73$