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Ecological niche distribution along soil toxicity gradients: Bridging theoretical expectations and metallophyte conservation

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ABSTRACT

Ecological niche modelling helps us to understand the spatial assembly of species in heterogeneous environments. Three patterns have been widely reported in the research literature regarding the relationship between realised niches and macronutrient concentration gradients: (1) species' optima are unevenly distributed, with a higher frequency in mesic conditions; (2) species' response curves are narrower when optima density is higher; and (3) species with optima at the extremes of the gradients have skewed response curves with a longer tail toward mesic conditions. This study aims to test the existence of these patterns on a vegetation model occurring in metalliferous soils comprising copper and cobalt along a toxicity gradient in south-eastern D.R. Congo. Realised niches of 80 taxa were modelled using generalised additive models. The niche optima and the niche widths were determined for each taxon. Results highlighted three groups which differ according to the niche optima location along the soil metal concentration gradients. The patterns found along macronutrient concentration gradients were, to some extent, transposable along micronutrient concentration gradients. Our findings on the diversity and assembly of realised niches has consequences for plant conservation strategies.

1. Introduction

The field of vegetation ecology has long been concerned with describing the distributions of species along edaphic gradients (Austin, 2007, 1987, 1985). Ecological niche theory and modelling help us to understand the spatial assembly of species in heterogeneous environments (Chase and Leibold, 2003; Higgins et al., 2012). Since Hutchinson's definition (1957), ecological niches of species distributed along one or several environmental factor gradients has been addressed using statistical and mathematical tools (Austin, 2007; Guisan et al., 2002; Yee and Mitchell, 1991). A fundamental ecological niche is related to the physiological tolerance of species, whereas a realised ecological niche is the range of conditions and resources where the species is actually found given the presence of biotic interactions (Hutchinson, 1957). At a species level, the response curve along one environmental factor gradient, obtained from field occurrence or abundance data, is the classical representation of the one-dimensional realised niche (Austin, 2007; Austin and Meyers, 1996; Coudun and Gégout, 2006; Rydgren et al., 2003).

Along edaphic macronutrient soil concentration gradients, the combination of species tolerance and interspecific competition leads to three characteristic patterns in species response curves: (1) species' optima are unevenly distributed, with a higher frequency in mesic conditions (e.g. Minchin, 1989; Lawesson and Oksanen, 2002); (2) a species response curve tends to be narrower when optima density is higher (e.g., Minchin, 1989; Lawesson and Oksanen, 2002); and (3) species with optima at the extremes of the gradient present skewed response curves with a longer tail toward mesic conditions (e.g., Austin and Gaywood, 1994; Chase, 2011). These patterns were observed for distinct vegetation types with components such as calcium (Hájková et al., 2008), magnesium (Wright et al., 2006), phosphorus (Austin and Meyers, 1996) or carbon-nitrogen ratio (Heikkinen and Mäkipää, 2010). However, the existence of such distribution patterns along micronutrient soil concentration gradients has not been addressed so far, although soils in some areas of the world do present unbalanced micronutrients concentrations that constrain vegetation type and species distributions (Baker et al., 2010; Bizoux et al., 2004; Brady et al., 2005; Duvi-gneaud and Denaeyer-De Smet, 1963; Kruckeberg, 1992; Whiting et al., 2002).

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As micronutrients and beneficial elements (i.e. elements that can enhance physiological mechanisms but are not considered nutrients) are needed in small to very small quantities by plants, the range of their available concentrations in soil may considerably differ from that of macronutrient concentrations. The first of these differences is that a deficiency in micronutrients, which is uncommon and generally observed on a very small range of concentrations (Pilon-Smits et al., 2009), could lead to the absence of skewed species response curves at the lowest micronutrient concentrations. Secondly, many micronutrients are toxic at a lower concentration threshold than macronutrients (Ernst, 2006; Pilon-Smits et al., 2009) and can profoundly affect the distribution of species, implying the presence of truncated curves at the highest values of micronutrient gradients. Additionally, at high micronutrient concentration levels species' capacity to tolerate or accumulate those micronutrients can be a determinant for the occurrence of a particular species, which may in turn lead to very low density of realised niche optima.

Metalliferous soils and related vegetation offer favourable conditions as a model to test the three identified patterns of realised species response curves along a micronutrient soil concentration gradient. These ecosystems are characterised by variety of taxa with distinct tolerance thresholds to metals; these plants are called metallophytes. Vegetation hosting metallophytes is associated with the presence of particular metals in soil worldwide (Antonovics et al., 1971; Boyd, 2004; Ernst, 2006; Hall, 2002; Macnair, 1987; Pollard, 2000). Metallophyte taxa can be defined as 'absolute' if all individuals occur on metalliferous soils and 'facultative' if more than 25% of individuals occur on normal soils. Vegetation composed of both absolute and facultative metallophyte taxa, accompanied with other non-tolerant taxa, is expected to present a high diversity of species response curves (in terms of distribution shape) and a heterogeneous density of optima along metal gradients.

Copper (Cu) and Cobalt (Co) are trace metal elements in soils, and their presence depends on their parent materials (Hough, 2010; Ma and Hooda, 2010). Average total content in soil is around 30mg kg^{-1} dry weight for Cu and 8mg kg^{-1} for Co with slight differences between Geographical Zones (Alloway, 1995; Aubert and Pinta, 1980; Hough, 2010; Lopes, 1980; McLean and Bledsoe, 1992). Depending on the soil properties, a proportion of total metal concentration is available for plants and microbes (Degryse et al., 2009). Cu is known as an essential redox-active micronutrient for plants; the role of Co is still being identified, but it could be considered as a beneficial element (Pilon-Smits et al., 2009; Yruea, 2005). They are both found at less than 10mg kg^{-1} dry weight in plant tissues (Palit et al., 1994; Yruea, 2005). Toxicity or deficiency thresholds of metal are species- or population-dependent. Elevated concentrations of these metals induce toxic effects such as inhibition of various processes in plant metabolism (Prasad and Hagemeyer, 1999) while deficiency is uncommon (Foy et al., 1978; Pilon-Smits et al., 2009; Yruea, 2005).

The metalliferous grasslands of the south-eastern D.R. Congo represent highly unusual vegetation where community types are dependent on copper and cobalt concentrations in soils (Duvigneaud and Denaeey-De Smet, 1963; Malaisse et al., 2016). Available metal concentrations here reach approximately 300 times those in the predomi-

nant soils of the region, creating unique edaphic conditions and, thus, a succession of plant communities along metal soil concentration gradients (Delhay et al., 2016; Saad et al., 2012; Séleck et al., 2013). These plant communities host more than 550 species including 56 endemic taxa which are absolute or facultative metallophytes (Faucon et al., 2010; Leteinturier, 2002; Malaisse et al., 2016). The economic value of Cu and Co encourages mining activities in the region, and this has consequences on the vegetation. More than 75% of endemic taxa are threatened (VU, EN, CR) according to IUCN criteria (Faucon et al., 2010). Restoration efforts are currently being implemented, but their success is highly dependent on a good understanding of vegetation-soil patterns. Although some recent studies have been implemented at a community (Delhay et al., 2016; Ilunga wa Ilunga et al., 2015, 2013) or species scale (Boisson et al., 2017; Boisson et al., 2016, 2016b; Chipeng et al., 2010; Lange et al., 2014), no one has addressed a quantitative overview of flora distribution along the gradients of selective edaphic factors: i.e. copper and cobalt soil concentrations.

Considering this as an opportunity to make a theoretical and practical contribution to science, the present study aims to characterise the patterns of one-dimensional realised niche distribution of plant species along two micronutrient soil concentration gradients. We tested the existence of three patterns along copper and cobalt soil concentration gradients: (1) the uneven distribution of optima along the gradients; (2) negative correlation between the niche optima density and the niche width; and (3) negative correlation between the skewness coefficient and metal soil concentration. Finally, we discussed how patterns of niche distribution can inform ecological restoration strategies.

2. Method

2.1. Study area and vegetation

This study was performed in the Katangan Copperbelt (D.R. Congo). The annual average precipitation of about 1174 mm is concentrated in the wet season (November – March). The average daily temperature during the wet season is 20.2°C , and the temperature ranges between $15\text{--}17^{\circ}\text{C}$ at the beginning of the dry season (Kazadi and Kaoru, 1996).

2.2. Vegetation sampling and soil analysis

In two separate trips undertaken in March 2008 and 2009, 172 square meter plots (1m^2) were established on three natural sites in the area between Tenke and Fungurume towns, on the Fungurume V, Kavifwafwaulu [1] and Kazinyanga outcrops (Table 1). For each site, sampling was performed according to a systematic grid following six to eight parallel transects stretched from the bottom to the top of the hill. Plots were located at nodes between transects and level curbs, every 5 m in altitude (Fig. 1b).

In each plot, all present plant taxa were identified at the botanic species or variety level. Identification of specimens was based on the African Plant Database (Conservatoire et Jardin Botanique de la ville de Genève and South African National Biodiversity Institute – Pretoria 2013). Due to the lack of recent taxonomic revision, some individuals

Table 1

Site, hill, geographical coordinates (utm), sampled area (ha), number of sampling plots and minimum and maximum copper and cobalt concentrations (mg kg^{-1}) of three hills of Katangan copper belt (Fungurume V, Kavifwafwaulu and Kazinyanga) measured by flame atomic absorption spectroscopy after EDTA extraction. Each concentration is an average of measured concentration of 4 quadrat soil samples.

Site	Geographical coordinates (utm)		Sampled area (ha)	Number of plots	[Cu] (mg kg^{-1})		[Co] (mg kg^{-1})	
	X	Y			Min	Max	Min	Max
Fungurume V	422276	8826409	25.9	83	92	6737	6	927
Kavifwafwaulu [1]	407436	8830457	11.9	37	29	7559	2	65
Kazinyanga	418560	8825047	7.36	52	60	10136	2	126

Fig. 1. Predicted value of copper and cobalt concentrations on one copper hill, Fungurume V (a) and the location of the performed sampling plots (b).

could not be identified at the species level; these were instead individualised at putative species level.

One composite soil sample was made from four soil subsamples collected at the four corners of each plot at a depth of 15 cm. Each composite sample soil was air-dried and sieved to 2 mm prior to physio-chemical analyses. Soil extractable concentrations of Cu and Co were measured using 0.5 N ammonium acetate-EDTA ($\text{CH}_3\text{COONH}_4\text{-EDTA}$; (Lakanen and Erviö, 1971). The extraction soil/solution ratio was fixed to 1:5 using a buffer at pH 4.65 (Kucak and Blanuša, 1998). Finally, metal concentrations were measured in the filtrate (the filter used was S&S-595-1/2) via flame atomic absorption spectroscopy (VARIAN 220), after which the solution was stirred for 30 min.

Taxa present in a minimum of 7 plots were selected, and their species response curves modelled. Many statistical methods provide models for one or more environmental variables (Austin, 2007; Austin and Meyers, 1996; Guisan et al., 2002; Guisan and Zimmermann, 2000; Meynard and Quinn, 2007; Oksanen and Minchin, 2002) but the generalised additive model (GAM, Hastie and Tibshirani, 1990; Yee and Mitchell, 1991) does not require any assumption regarding the shape of the response curves to environmental factors (Oksanen and Minchin, 2002). For this reason, we selected this model with binomial likelihood to model the probability of occurrence for each selected species along the two gradients: available Cu and Co soil concentrations. For each taxon, the best fitted model from three degrees of smoothing (3, 4 or 5) was selected with the Akaike Information Criterion (AIC), choosing the model with the lowest AIC value.

The niche optimum for each species along the two soil concentration gradients was estimated using the projected value of the highest probability of occurrence on the x-axis. The borders of niche width (amplitude) were determined by projecting the lower and upper limits (LL and UL respectively) of 80% of the area under the curve centred on the optimum on the x-axis (Gégout and Pierrat, 1998). Niche width was calculated via the difference between UL and LL. Statistical modelling and values extraction were performed via a function GAM.summary (Appendix E) operating with 'mgcv' library (Wood, 2006) in the R statistical software to model response curves with GAMs. This function is able to generate response curves using the parameters of the model and append the curve of each factor included in the dataset. The existence of the three above-mentioned patterns was experimentally

demonstrated the following ways: (1) to test an uneven distribution of niche optima along the Cu-Co gradient, we used a kernel density function (density function uses the Fourier transformation and applies a linear approximation to model a curve of optima density (y-axis) along gradients (x-axis)); (2) to test the relationship between the optima density and the niche width a Pearson's correlation was performed; and (3) to test the relationship between the skewness and the optima location, the skewness of species response curves was determined using Fisher's coefficient (Fisher, 1930):

$$\gamma_1 = \rho_3 / \sigma^3$$

where ρ_3 is the third central moment and σ is the standard deviation. Each response curve can be assimilated to probability density function with an area under the curve (A) higher than 1. Mathematical expectations were thus defined as

$$E(x^n) = \int x^n \mu(x) dx \xrightarrow{\text{defined domain}} \frac{\sum_{j=1}^m x_j^n \mu_s(x) \Delta x}{A}$$

where m is the number of rectangles, Δx is $x_{\max}/m$, $\mu_s(x)$ is the probability of occurrence of species s , and A is the sum of m rectangles areas. For the calculation: $m=1000$ and $x_{\max}$ = maximum measured value of copper or cobalt. Mathematical expectations were used to estimate $V(x)$, the variance (second central moment) of each response curve and the n^{th} central moment. Finally, linear regressions were respectively used to quantify the relationship of the niche optima of taxa between Cu and Co gradients and to determinate the relationship between skewness coefficients and optima values along gradients. All statistical analyses were executed under R Statistical Software (R Development Core Team, 2010).

3. Results

3.1. Data ranges and metal concentration gradients

Across the three studied sites, Cu concentrations ranged from 29 mg kg^{-1} to $10\,136 \text{ mg kg}^{-1}$ and Co from 2 mg kg^{-1} to 927 mg kg^{-1} (Table 1). The Spearman's rank correlation coefficient between copper and cobalt concentrations was 0.61 ($p\text{-value} < 0.05$).

3.2. Species distribution overview

Out of 184 taxa in the dataset, 80 occurred in at least seven plots (Appendix A, Taxa list: Appendix B, models: Appendix C). Along Cu and Co concentration gradients, three types of distributions showed similar patterns based on the niche optima position and the response curves. More than half of taxa yielded truncated response curves with tails toward the highest concentrations or, inversely, with tails toward the lowest concentrations. The remaining taxa displayed clearly defined niche optima (i.e. being unequal to LL or UL) along the gradients.

Along the copper soil concentration gradient, 40 taxa presented niche optima at the lowest value (29 mg kg^{-1} Cu, Fig. 2), 31 taxa had optima ranging from 453 to $7\,467\text{ mg kg}^{-1}$ Cu (Fig. 2), and nine taxa had their niche optima at the highest value of the gradient ($10,136\text{ mg kg}^{-1}$ Cu, Fig. 2, Appendix C). Along the cobalt soil concentration gradient, 45 taxa positioned their niche optima at the lowest value (i.e. 2 mg kg^{-1} Co, Fig. 3, Appendix D), 27 taxa with niche optima ranging from 26 to 664 mg kg^{-1} Co were identified, and seven taxa had niche optima at the highest value of the concentration gradient (927 mg kg^{-1} Co, Fig. 3). The Pearson correlation coefficient between the copper and the cobalt optimum values was 0.38 (p -value < 0.001). 61% of total taxa (49 taxa) displayed similar niche optima positions along both the copper and cobalt concentration gradients.

3.3. Pattern 1: Uneven distribution of optima along the gradients

The distribution of optima was uneven along the concentration gradients of both metals, with a high variation recorded for cobalt. In general, the density of optima was the highest in low metal concentrations (43 taxa between $0\text{--}1\,000\text{ mg kg}^{-1}$ Cu and 49 taxa between $0\text{--}100\text{ mg kg}^{-1}$ Co) and decreased with increasing metals concentrations (Fig. 4). For the cobalt concentration gradient only, density of niche optima approached zero (lower than 5.10^{-5}) at concentrations from 530 to 780 mg kg^{-1} (Fig. 4b).

3.4. Pattern 2: Negative correlation between niche optima density and niche width

The niche width of taxa was negatively correlated with the density of their niche optima along the copper and the cobalt concentration gradients. For the copper gradient, the niche widths of taxa decreased with increasing density of niche optima for all taxa ($r = -0.75$; p -value < 0.001) and also for Group 2 ($r = -0.68$; p -value < 0.001). For the cobalt gradient, a similar relationship was observed for all taxa ($r = -0.61$; p -value < 0.001) and also for taxa with clearly defined niche optima ($r = -0.69$; p -value < 0.001). The niche widths of taxa were the narrowest in the lowest concentrations along both metal soil concentration gradients.

3.5. Pattern 3: Negative correlation between the skewness coefficient and niche optima

The skewness coefficient of niches was significantly negatively correlated with copper optima locations for all taxa ($r = -0.81$; p -value < 0.001) and for taxa with clearly defined niche optima ($r = -0.72$; p -value < 0.001). One-dimensional realised niches were more symmetric when optima ranged between $4\,000$ and $6\,000\text{ mg kg}^{-1}$ (Fig. 5a). Along the cobalt gradient, skewness coefficients had similar relationships with the niche widths for all taxa ($r = -0.69$; p -value < 0.001) and for taxa with clearly defined niche optima ($r = -0.54$; p -value < 0.001). Symmetry of species response curves was encountered at 150 mg kg^{-1} for taxa with clearly defined niche optima and at 500 mg kg^{-1} for all taxa (Fig. 5b). Along these two metal soil concentration gradients, the taxa whose niche optima were located at the lowest metal concentrations had response curves with tails towards the

highest concentrations, while the taxa whose niche optima were located in the highest metal concentrations displayed curves with tails towards the lowest concentrations.

4. Discussion

4.1. General patterns of realised niche distribution along the metal concentration gradients

The niche concept is complex and species response curves along concentration gradients of available soil metals depend on several edaphic factors (Austin, 2007; Soberón and Nakamura, 2009; Whittaker et al., 1973). However, this study has highlighted that the pattern of realised niche distributions identified along macronutrient soil concentration gradients is in some ways similar to those along the micronutrient gradients of copper and cobalt. Most taxa displayed realised niche optima in metal concentrations lower than $2\,500\text{ mg kg}^{-1}$ Cu and 100 mg kg^{-1} Co, and presented the narrowest realised niche width. These results suggest that mesic conditions exist at the lowest concentrations of the copper and cobalt gradients up to values of respectively $2\,500\text{ mg kg}^{-1}$ Cu and 100 mg kg^{-1} Co. Close to the limits of these soil concentration gradients, the response curves were skewed with a longer tail into the centre of the gradients.

Although the correlation between copper and cobalt realised niche optima was significant, the correlation coefficient was relatively low. This clearly indicates that a particular taxon does not necessarily occur in the same range of concentrations along each gradient (defined by Groups 1, 2 and 3). For example, the taxon *Crotalaria cobalticola* (Fabaceae) was located in Group 3 for copper, signifying that the species optimum occurred in the highest concentrations, while its optimum (realised niche) occurred in the lowest concentrations in Group 1 for cobalt.

4.2. Origins and consequences of observed patterns

The highest optima density being identified at the lowest metal concentrations suggests that competition should be a determinant factor leading to narrow niche width by the competitive exclusion principle (Grime, 1973; Hardin, 1960). This has consequences for the plant-soil relationship as highlighted by previous studies on metalliferous vegetation (Saad et al., 2012; Séleck et al., 2013), which has been shown to be more common in the steppic savannah ($0\text{--}3\,500\text{ mg kg}^{-1}$ Cu; $0\text{--}100\text{ mg kg}^{-1}$ Co) than in the steppe ($3\,500\text{--}10\,000\text{ mg kg}^{-1}$ Cu; $100\text{--}1\,000\text{ mg kg}^{-1}$ Co). Séleck et al. (2013) highlighted 7 vegetation parts in those plant communities in which 6 are structured as a continuum along the copper and the cobalt gradient. Competition has recently been identified by Delhaye et al. (2016) as being higher in the lowest copper-cobalt concentration, confirming the uneven distribution of optima found in this study.

Variation of niche width has previously been highlighted along the copper soil concentration gradient for taxa occurring on another site near Lubumbashi (Ilunga wa Ilunga et al., 2013) and for endemic species (Boisson et al., 2016c). However, some exceptions could be observed for some taxa such as the grass *Loudetia simplex*, which displayed a niche optimum equal to the lowest copper-cobalt gradient limit (LL) but nevertheless was shown to have a broad niche width. This observation corroborated previous occurrence observations for the same species in two plant communities, steppes and steppic savannah (Boisson et al., 2016a; Duvigneaud and Denaeyer-De Smet, 1963), and suggests that the species has a wide range of tolerance to soil concentrations of metals compared to taxa with narrow realised niches in the same group, for example, *Aeschynomene pygmaea*.

The broad realised niches of some taxa could result from the high population variations between different realised niches, considering that soil metal concentrations exert a strong selective pressure on plants (Antonovics et al., 1971; Baker et al., 2010; Shaw, 1990; Whitting et al., 2002), and that the metal concentrations varied be-

Fig. 2. Modelled realised niches sorted by niche optima along available copper gradient (29 to 10 136 mg kg⁻¹) by generalised additive models (GAMs) for 80 metallophyte taxa in the Katanga Copperbelt. Plain dots, squares and diamonds are the niche optima of taxa determined by the gradient value of the highest probability of occurrence for Group 1, 2 and 3, respectively. Straight lines are the niche widths calculated by the difference between the borders of area of 80% under GAM-curve.

Fig. 3. Modelled realised niches sorted by niche optima along available cobalt gradient (2–927 mg kg⁻¹) by generalised additive models (GAMs) for 80 metallophyte taxa in the Katanga Copperbelt. Empty dots, squares and diamonds are the niche optima of taxa determined by the gradient value of the highest probability of occurrence for Group 1, 2 and 3, respectively. Straight lines are the niche widths calculated by the difference between the borders of area of 80% under GAM-curve.

Fig. 4. Kernel density estimations of taxa optima distribution along copper and cobalt gradient (mg kg^{-1}).

Fig. 5. Skewness coefficients relationship with the copper and cobalt optima positions for the 80 identified taxa from the three groups. Plain or empty dots, squares and diamonds are the niche optima for respectively Group 1, 2 and 3 along copper/cobalt gradient. (a) Linear regression of skewness coefficients along copper niche optima. (b) Linear regression of the skewness coefficients along cobalt niche optima.

tween and across sites. Adaptation and differentiation are also promoted by extreme dispersal limitations (Ackerly, 2003), which can occur in the plant taxa of the copper hills. The taxa could also show differences in realised niches between individuals; individuals could be specialised to distinct ranges of concentrations along the soil metal gradients (Bolnick et al., 2007). Determination of the fundamental niche of distinct individuals from a single population of a taxon could us help to understand the origin of the broad realised niches (Boisson et al., 2017). Another study on one tolerant species from metalliferous soils suggested that niches may evolve relatively quickly in serpentine soils between ecotypes of *Collinsia sparsiflora* (Wright et al., 2006).

The skewed curves of the realised niches could be due to a progressive decrease in physiological advantages for an extreme-adapting species until it is outperformed by more productive species (Austin and Smith, 1989; Austin and Gaywood, 1994), in accordance with the species coexistence theory. The relationship between the skewness and the optima location was not identical for both metals; this was as expected because copper and cobalt do not have the same physiological role in plants. Copper is known as an essential redox-active micronutrient while cobalt is identified as a beneficial element (Pilon-Smits et al., 2009; Yruea, 2005). Cobalt is expected to induce toxicity stress at lower concentrations than copper. In the context of our study, based on metalliferous outcrops, mesic conditions were expected to be $0\text{--}65 \text{ mg kg}^{-1}$ and $0\text{--}6 \text{ mg kg}^{-1}$ of available copper and cobalt, respectively (Anderson et al., 1973; Colinet, 2013; Faucon et al., 2011; Ma and Hooda, 2010; McLean and Bledsoe, 1992).

Many factors are constantly in interaction and have an effect on metal bioavailability, including the total metal present in the soil, pH, clay and hydrous oxide content, organic matter and redox conditions (Reichman, 2002). Also, the bioavailability of nutrients and the met-

als in the rhizosphere could change over an area according to local variations in pH and the acidification ability of the plants (Marschner et al., 1986). By changing the availability of metal, these factors and related interaction are subject to change the realised niche parameters (optimum, width, skewness) of the species by having distinct toxicity or resource effect.

4.3. Implications for ecological restoration

The three patterns identified in the realised niches of plant taxa along large copper and cobalt soil concentration gradients have consequences for ecological restoration strategies (Boisson et al., 2017; Young et al., 2005). Strategies for soil restoration and revegetation traditionally focus on the transfer of plant communities, with attempts to maintain populations of the main species (Society for Ecological Restoration, 2004). For metallophyte communities, reproduction of the natural realised niche patterns is required; this can be achieved by recreating the initial soil conditions. Changes in edaphic conditions compared with those found in the natural site in terms of Cu or Co soil concentration gradients could drastically change plant-soil relationships and consequently plant-plant interaction. The non-simultaneous existence of any of these three patterns in a restored plant community pool could create vacant niche space and promote the colonisation of undesirable plants with broad niche widths (i.e. weeds).

The skewed curves of realised niches in combination with the niche optima-width relationship are complex to convert into practical restorative actions. This primarily suggests that the assemblage of taxa gradually changes along the metal concentration gradient, and thus communities of multiple taxa are associated with interdependent dominance in plant communities for any given metal concentration. When the taxa

are transferred to a restoration site, the assemblage should be reproduced to respect the original pattern of the natural ecosystem and thereby ensure long term stability of the plant communities.

Species with narrow realised niche widths in the highest metal concentrations are likely to be highly dependent on metals in the soil. For example, the absolute metallophyte *Haumaniastrum robertii*, with a narrow realised niche in high metal concentrations, is recognised as the 'copper flower' and is not able to grow in low Cu and Co soil concentrations (Ilunga Kabeya et al., 2018). The taxa with broad realised niches in low Cu and Co soil concentrations are associated with a community of taxa producing a large biomass (Delhayé et al., 2016) and high competition for light and nutrients.

To better ensure the success of restoration strategies at taxa scale, study of the fundamental niches (as suggested to understand population variations) could help us to understand the relationship between these taxa and the presence of metals in the soil. The species with niche optima in the highest metal concentrations and a broad niche width provide an opportunity to test these plants' efficacy in rehabilitation strategies, such as *Crepidiorhodon perennis* and *Andropogon shirensis* (metal tolerant grass), in the Katangan Copperbelt (Boisson et al., 2016a; Faucon et al., 2011).

Identifying the distribution pattern of realised niches along micronutrient soil concentration gradients helps significantly in disentangling ecosystem interactions. While the concept of the ecological niche has been somewhat neglected over the last decade, this multidisciplinary approach is useful in linking together a range of theoretical concepts in plant ecology. In this context, the further study of the distribution of realised niches along other soil micronutrient or element concentration gradients appears to be scientifically relevant and could provide practical solutions for restoration purposes.

5. Conclusion

The model-based studies about the realised niche of metallophyte are rare. In this study, the usage of model considerably helps to identify the realised niche of metallophyte along copper and cobalt gradient. The results in term of optimum and width (i.e. positive correlation) match with field observation and other scientific papers about the same flora. Even if the skewness is a parameter that is difficult to interpret, we find that the pattern skewness-optimum corresponds with other studies about realised niche. Expected patterns along macronutrient are confirmed along micronutrient excepted for the skewness-optimum relationship with slight differences between copper and cobalt.

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Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:<https://doi.org/10.1016/j.jecolmodel.2019.108861>.

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