

Information on population trends and biological constraints from bat counts in roost cavities: a twenty-two-year case study of a Pipistrelle bats (Pipistrellus pipistrellus, Schreber) hibernaculum.

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1	Title: Information on population trends and biological constraints from bat counts in roost
2	cavities: a twenty-two-year case study of a Pipistrelle bats (Pipistrellus pipistrellus, Schreber)
3	hibernaculum.
4	
5	Running head: Pragmatic approach to analyse count data of bats in hibernaculum
6	
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25 **Abstract** (Words count: 349)

Context: According to the current trend of biodiversity loss, information on population trends at
 large temporal and spatial scales is necessary. However, well-documented animal population
 dynamics are generally based on intensive protocols requiring animal manipulation, which can
 be impossible to conduct in species for which conservation is a concern.

Aims: For many bat species, an alternative approach entails performing an appropriate analysis
 of counts in roost cavities. Because of managers' perception of chaotic variations through time,
 relatively few count monitoring surveys are regularly analysed. Here, we present the analysis of
 a twenty-two-year survey of a large hibernaculum of pipistrelle bats (*Pipistrellus pipistrellus*)
 located in a railway tunnel in Paris, France.

35 *Methods*: We propose that using combinations of population dynamics modelling using 36 demographic parameters from the literature and statistical analyses helps with identifying the 37 biological and methodological effects underlying the dynamics observed in census analyses.

Key results: we determined that some of the observed year-to-year variations of population size 38 39 cannot be explained only by the intrinsic dynamics of the population. In particular, in 1993-40 1994, the population size increased by >40%, which should have implied a massive immigration. 41 This change coincided with the end of the operation of the railway line. After consideration of a 42 drastic trend of population decline (7%/year) we were able to detect this event and several environmental effects. Specifically, the winter conditions and the temperature in July affected 43 the colony size, presumably because of aggregative behavior and reproduction success, 44 45 respectively.

Conclusions: Emigration-immigration processes might have preponderant effects on population
 dynamics. In addition, we our analysis demonstrates that (1) the study population suffered a
 large decline (2) a combination of human disturbance and meteorological variation explains

these dynamics; (3) Emigration-immigration processes have preponderant effects on the
population dynamics,

Implications: To conduct a meaningful analysis of non standard time series and provide a source of data for implementing biodiversity indicators, it is necessary to include (1) the local knowledge of the people involved in the field surveys in these analyses (the existence of disturbances and site protections) and (2) meteorological information for the appropriate seasons of the year.

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61 Key words

62 *Pipistrellus pipistrellus*; Population dynamics; count monitoring; deterministic matrix model;

63 Stochastic-individual-based population model; disturbance

64

66 Introduction

In the current context of the drastic erosion of biodiversity, "unprecedented additional efforts 67 would be needed to achieve a significant reduction in the rate of biodiversity loss at all levels" 68 (Biodiversity Synthesis of the Millennium Ecosystem Assessment (MEA 2005). To reach this 69 target, the efforts undertaken by each country should be assessed. In Europe, the 70 71 implementation of the Convention on Biological Diversity has been outlined by the Streamlining European Biodiversity Indicators (SEBI) initiative that proposes a first set of indicators to 72 monitor the progress in Europe (European Environment Agency, 2009). The first set of this 73 74 indicator can be considered "direct" core biodiversity indicators based on population trends 75 and the diversity of a group of species (Balmford et al. 2005) or the Red List Index (Butchart et al. 2005). 76 Well-documented animal population dynamics studies are generally based on quantitative 77 methods allowing unbiased and independent estimations of survival, fecundity and dispersal 78 rates and subsequent projection modelling using these specific rates (Beissinger and 79 80 McCullough 2002). However, the acquisition of such data requires intensive protocols (e.g., 81 capture-mark-recapture protocols to estimate survival and/or dispersal rates). These methods, 82 often involving animal handling, such as capture-recapture, could be controversial approaches when the species are endangered and/or sensitive to disturbance and could also affect the 83 results (Saraux et al. 2011), particularly with regard to bats (Punt 1970; Dietz et al. 2006). In 84 85 addition, such time-consuming methods require highly qualified personnel and can hardly be 86 extended to a large number of sites and species.

87

Most of the European bat populations have a poor conservation status because of various
pressures, such as the loss of suitable foraging habitats (Walsh and Harris 1996; Kunz and

90 Fenton 2003), agricultural intensification (Swanepoel et al. 1999; Wickramasinghe et al. 2003), urbanisation (Kurta and Teramino 1992; Loeb et al. 2009) and forest management (O'Donnell 91 2000). Human activities strongly reduce the availability of underground sites used by bats for 92 93 breeding or/and hibernation. Whereas some underground sites are filled in, blocked up or 94 transformed (Mitchell-Jones et al., 2007), others are opened for tourism activities, resulting in 95 recurrent disturbances. Such activities may have major consequences on the population 96 dynamics of bat species at large scales since (1) several studies have stressed the high 97 sensitivity of bats to human disturbance (Mc Craken 1989; Speakman et al. 1991; Gore et al. 2012); (2) in some cases, the loss of a single site can affect the dynamics of bat populations over 98 areas of several thousands of square kilometers (Mitchell-Jones et al. 2007). Within this 99 100 context, an increasing number of bat populations occurring in roost cavities are protected and 101 regularly counted (see the Conservation Trust and Eurobat annuals reports). However, because of the managers' perception that hibernacula counts exhibit chaotic variations over time, only 102 103 relatively few hibernacula counts are regularly analysed, which, in turn may lead to the 104 abandonment of such monitoring. Nevertheless, efforts have been made to analyse count data 105 provided by volunteers in the context of breeding roost surveys and to provide sound 106 assessment of population trends (Warren and Witter 2002). Less effort has been devoted so far to hibernaculatum monitoring, despite the availability of hibernaculatum count data and the 107 108 need for conservationists to detect population declines when formulating management decisions (Trombulak et al. 2001; Frick et al. 2010). O'Shea et al. (2003) advocated that new 109 110 techniques must be explored and modern statistical designs applied in order to improve the 111 scientific basis for conclusions about future bat population trends. Some recent studies and 112 reports have considered hibernaculatum count data to assess population variations (see 113 Prendergast *et al.* 2010; European Environment Agency 2013)

A better understanding of the biological information included in these time series evaluations
could allow (1) to design and propose of more efficient count protocols (e.g., which
environmental variable to record at monitoring sites), (2) to detect population trends and then
provide efficient biodiversity indicators, and (3) to assess the impact of management actions
(such as fitting a grille across the hibernaculatum entrance permitting the free passage of bats
but not people, Provision of additional roosting places, etc.)

120 The purpose of this study is to investigate how simple, specific count-based datasets can be 121 used in combination with standard ecological methods and generic demographic data to 122 understand the processes underlying the observed population dynamics. If count data from 123 roosts in winter can be considered as a reliable proxy of a local population's size (for sedentary species), we hypothesize that we will be able to detect the effects of meteorological conditions 124 125 on local population dynamics. Weather and food availability are known to be highly correlated 126 with each other and with parturition in insectivorous bats (Arlettaz et al. 2011), which is in turn 127 related with reproductive success and survival of newly volant young (Tuttle 1976). Therefore, during the reproductive period, local bad weather conditions such as low temperature or 128 129 important rainfall should affect the reproductive success and then impact negatively population 130 size in winter, and, subsequently, count data. However, if population dynamics are only assessed through roost count data (and not with e.g., variation in demographic parameters), 131 132 meteorological effects may be masked by (1) a population trend (such as a large decline), (2) an aggregative behavior linked with winter condition (Mumford 1958; Sandel et al. 2001) and (3) 133 human disturbance. 134

Local population dynamics do not only rely on intrinsic processes (mortality and fecundity) but may also be more or less strongly influenced by immigration or emigration processes. Although disantangling the roles of intrinsic dynamics and immigration/emigration is not possible with 138 count data, however based on a good knowledge of the species' demography it is possible to detect changes that cannot be explained only by intrinsic dynamics and involve necessary 139 140 events of immigration or emigration. We first used population dynamics models with a set of 141 generic demographic parameter values available from the literature to construct a realistic 142 distribution of expected annual growth rates in the absence of emigration/immigration. In a 143 second step, we compared this expected distribution (which only depends on the life history 144 traits of the species and its capacity of intrinsic population growth) with the observed annual growth rates to detect important immigration-emigration events or local catastrophic events, 145 146 such as epizootics (which is not an exclusive hypothesis) (Blehert et al. 2009). 147 After these exploratory steps, we conducted a statistical modelling to test the following assumptions: 148 149 (1) Favorable meteorological conditions such as higher temperatures during the reproductive 150 period should positively influence counts from roost in following winter (larger number of 151 young produced), whereas unfavorable meteorological conditions such as strong precipitation should have a negative impact. (2) Cold temperatures in winter may lead to increased counts 152 153 from roosts (in the same winter), due to aggregative behavior. (3) human disturbance in the 154 roost should adversely affect counts. We believe that applying this simple approach to bat count data will allow a better 155 156 understanding of the biological information contained in available time series. We focused on a large population of common pipistrelle bats (*Pipistrellus pipistrellus*, Schreber) 157 that was monitored for 22 years in an urban area. 158

159

160 MATERIALS AND METHODS

161 Study area

162 From 1991 to 2012, we monitored the population size of a large hibernaculum of pipistrelle bats (Pipistrellus pipistrellus) located in a railway tunnel in Paris, France. This railway tunnel was 163 164 in service until 1993. The bat counts were conducted in the winter by the same observers 165 during the entire period. The counting method is a direct visual count of individuals who are 166 located in shallow notches. These notches are expansion joint vertical depth of 10 cm between 167 concrete slabs (see supplementary material S1). The count procedure leads thus to very thin 168 measurements errors, because all individuals are observable. This hibernaculum is one of the 169 most important for the species in France and the only one currently known wintering site inside 170 the city of Paris (Lustrat and Julien 1993). Although it is possible that other unknown 171 hibernaculums exist in Paris or in its immediate surroundings, common pipistrelles are generally found singly or in small numbers in crevices of buildings and trees in winter (Dietz et 172 173 al. 2007). Thus, if existing, other hibernaculums are probably scarce and small.

174 Population dynamics modelling

To discriminate between the roles of intrinsic dynamics (fecundity and mortality) and yearly emigration-immigration processes on the population dynamics, we used a combination of deterministic and stochastic models for population dynamics.

178

179 Deterministic matrix model

180 In a first step, we used a one-sex, deterministic, age-classified Leslie-matrix model (Caswell,

181 2001) to infer the equilibrium intrinsic population properties (in particular, the asymptotic

time-invariant population growth rate). This model was implemented using the ULM software

183 (Legendre and Clobert 1995; Ferrière *et al.* 1996). We considered a two-age class model (Fig. 1.;

184 Sendor and Simon 2003) with the following parameter values obtained from the literature: the

185 first reproduction occurring in the first year (Schober and Grimmberger 1998; Dietz *et al.* 2007);

186	the first year survival probability, s ₀ , was set to 0.53 (Sendor and Simon 2003); the adult survival
187	probability, s_{ad} , varied from 0.37 to 0.80 (Thompson 1987; Gerell and Lundberg, 1990; Sendor
188	and Simon 2003,) and the average female productivity, F, varied from 0.58 to 0.988 (Webb <i>et</i>
189	al. 1996; Barclay et al. 2004). In many respects, bats have relatively slow life histories (Barclay
190	et al. 2004) thus one can expect small growth rate variations and this modelling approach aims
191	to assess a range of realistic values for annual growth rates.
192	
193	Approximate position of figure 1
194	
195	Stochastic, individual-based population model
196	In addition, to compare the actual observed annual growth rates with their theoretical
197	distribution under demographic stochasticity in the absence of immigration and emigration, we
198	developed a stochastic two-sex individual-based population model. This model allowed the
199	complete description of all of the individuals in terms of sex and age. In each time interval
200	(year), adult individuals were randomly paired according to a polygynous mating system with
201	an unrestricted harem size, i.e., we assumed that all mature females present in the population
202	can reproduce provided that there was at least one male in the population (Legendre et al.
203	1999). This hypothesis is realistic, as recent results suggest that the proportion of breeding
204	females among adult females is usually very high (about 95%, see Barclay et al., 2004). The
205	reproductive success of each pair was determined according to a Bernoulli drawing of
206	parameter F (in cases where F<1.0), or a Poisson trial of parameter F (in cases where F≥1). The
207	sex of each newborn individual was randomly determined according to a 1:1 mean sex ratio.
208	Reproduction was followed by the differential mortality of individuals according to age-specific
209	annual survival rates (S $_0$ and S $_{ad}$). Each individual survival event was drawn from a Bernoulli

210 function. The parameters used were the same as for the deterministic matrix models. Three scenarios were considered (pessimistic, median and optimistic) using either the minimal, 211 212 median or maximal values of S_{ad} and F obtained from the literature (Thompson 1987; Gerell and 213 Lunddberg 1990; Webb et al. 1996; Sendor and Simon 2003; Barclay et al. 2004). For each 214 scenario, the changes in the population size were investigated for 1000 independent random 215 population trajectories drawn over a fixed time horizon (22 years). The distributions of the 216 ratios of population size at a time t (N_t) / population size at a time t-1 (N_{t-1}) were then 217 compared to the observed ratios. At time zero, the population trajectories were assumed at the 218 demographic equilibrium (stable age-class distribution and balanced sex ratio), with a total 219 population size equal to the observed population size in 1991. In each model trajectory, years with Nt <20 individuals were excluded to avoid an inflation of the variance in the growth rates 220 221 due to a strong sampling effect (see e.g., Lande et al. 2003), which would have precluded the 222 comparison with the actual data (the observed population sizes were in the range of 350-1500). 223

224 Analysis of year-to-year variation and trend in population size. Effect of weather

conditions and human disturbance.

We examined temporal variations in the number of bats counted in the hibernaculatum by (1) 226 227 testing for a (decreasing or increasing) trend in numbers over the study period (1991-2012); (2) testing for an effect of the railway traffic on bat numbers. The railway tunnel was in service 228 until 1993; since traffic can substantially affect roost attractiveness, we included this factor in 229 230 our statistical analysis as a binary variable (presence or absence of railway traffic). (3) Finally, to 231 assess the effect of weather conditions on the winter population size, we tested the existence of correlations between the count data of bat and several weather variables. Some 232 meteorological variables, such as the winter conditions, are expected to influence hibernating 233

bat aggregation in certain favourable hibernacula, such as tunnels, which become more
attractive when the conditions are extreme. Other meteorological variables, such as the
precipitation or temperature levels, recorded the year before the winter bat counts were
expected to influence the reproductive success and survival of young (Grindal *et al.* 1992; Zahn
1999) and, thus, affect the population size the following winter. The weather data were
obtained from two websites: Infoclimat and European Climate Assessment and Dataset.

240

241 Statistical analysis

As we expected that some weather variables might be correlated with each other or might show trends over the period considered, we used Spearman's rank test to assess the correlation between paired samples of variables (see supplementary material S1), with the aim of avoiding the inclusion of correlated explanatory variables in statistical analyses. Weather variables in winter indeed showed strong correlations between them, as did the group of variables recorded in the spring before the winter counts (supplementary material S2). Only the variable *average temperature in June* showed a yearly significant increasing trend

249 (supplementary material S2).

250 In a first step, we assessed the influence of each variable (yearly trend, human disturbance

251 (railway traffic), weather conditions) on count data. In a second step, in order to take into

account possible population trend that could hide the effect of human disturbance or weather

253 conditions on bat numbers, we adjusted these variables to yearly trend using a sequential

ANOVA (type I) in which the effects are tested sequentially.

255 When the variables exhibited a significant influence, we choose the variable that explained

256 more of the deviance using the Akaike Information Criterion (AIC) for these two groups (winter

variable and spring variable) because this criterion has the advantage of generality and can be

258 applied further than the normal linear models (Faraway 2006). Finally, we constructed a full 259 model including yearly trend, railway traffic and meteorological variables and we performed an 260 AIC-based model selection by removing variables (see supplementary material S4 for details) to 261 test if the full model was the most parsimonious choice. We used Generalised Linear Models 262 (GLMs) with Poisson error distribution due to the nature of the data (count) (Crawley 2009) and 263 we account for over-dispersion following Faraway's (2006) recommendations (note that similar 264 results were obtained with GLM with negative binomial error distribution). 265 We also carried out an alternative approach (see supplementary material S3), in which we took 266 into account simultaneously all variables (trend, railway traffic, weathers conditions), using Hierarchical Partitioning (HP, R package hier.part). This approach allowed us to identify the 267 most likely causal factors of variation in bat numbers while alleviating multicollinearity 268 269 problems (Mac Nally 2000).

270

271 **RESULTS**

272 Comparison of observed growth rates and those assessed with modelling.

273 Considering the range of demographic parameter values available in the literature, the

asymptotic time-invariant population growth rate (λ) given by our matrix model was between

275 0.523 and 1.065. Moreover, increasing simultaneously demographic parameters, such as

276 fecundity or survival, rarely led to λ >1.4 (Fig. 2).

277

278

Approximate position of figure 2

279

280 The stochastic model allowed us to obtain a distribution of annual growth rates that accounted

for the random fluctuations in the population size due to demographic stochasticity under the

282	three demographically reasonable scenarios for the Pipistrelle (Fig. 3) These three contrasted
283	scenarios are: a pessimistic scenario, S_0 =0.53, S_{ad} =0.37, F=0.58; a median scenario, S_0 =0.53,
284	S_{ad} =0.59, F=0.78 and an optimistic scenario, S_0 =0.53, S_{ad} =0.8, F=0.99. The 95% confidence
285	intervals of the annual growth rates were (0.38 - 0.66), (0.72 - 0.86), (1.01 - 1.11), respectively,
286	for the pessimistic, median and optimistic scenarios.
287	
288	Approximate position of figure 3
289	
290	The comparison between these expectations and the observed annual growth rates indicated
291	that some of the strong year-to-year increases ($N_t/N_{t-1}>1.15$) in the population size observed in
292	the time series cannot be explained only by the intrinsic dynamics of the population without
293	emigration-immigration process (Fig. 3) and may have implied massive immigration events. In
294	particular, there were two years (1993-1994 and 2008-2009, Fig. 4) in which the population size
295	increased by >40%, presumably because of a massive immigration (see the Discussion). The
296	increase observed between 1993 and 1994 (+384 individuals) corresponds to the abandonment
297	of the railway line crossing the tunnel occupied by the study population. As this change is very
298	likely related to an arrival of individuals following the increased attractiveness of the site, we
299	only considered the 1995-2012 period in our subsequent analysis.
300	
301	Approximate position of figure 4
302	
303	We ran an additional scenario (hereafter, the adjusted scenario) by slightly modifying the set of
304	demographic parameters of the median scenario to yield a growth rate similar to the one
305	observed over the 1995-2012 period (details of protocol and parameters are provided in

supplementary material S3). This adjusted scenario (S₀=0.57, S_{ad}=0.69, F=0.84) predicts that the
95% confidence interval of the year-to-year growth rate should be between 0.854 and 1.013
(Fig. 3). The comparison with the observed values of the growth rate during the 1995-2012
period (grey circles in Fig. 3) indicates that only three of the observed growth rates are included
in the confidence interval of this adjusted scenario, whereas eight are lower, and ten are
higher.

312

Population trend analyses and the importance of weather on between year

314 variations.

Over the 1991-2012 period, we did not detect any temporal autocorrelation between the yearto-year growth rates in year *t* and *t+1* (Spearman's rank correlation test, P=0.18), this indicates the absence of a strong bias in the bat count. Indeed, when an observer overestimates (or underestimates) bat numbers in a roost in a given year, one could expect that the estimation might be lowered (or inflated) in the next year, leading to an artefact of the alternation between the decreases and increases.

321 When considering each variable separately, only the year appeared significant, showing a 322 negative trend (Table 1, this model explained 53% of the deviance). When the weather 323 variables were adjusted to the year trend, most of their p-values drastically decreased (sequential ANOVA, Table 1). According to AIC values (Table 1) and the percentage of 324 independent variance explained provided by the Hierarchical Partitioning (see supplementary 325 326 material S5), the most important variable appeared to be the yearly trend, followed by presence of railway traffic, the group of winter weather variables (particularly frost day in 327 January) and secondarily the group of summer variables. We then retained the number of frost 328 329 day in January (Frost day) for the group of winter weather variables and the temperature in July

330 (T._{July}) for the group of summer weather variables. We, thus, built a complete model in which the count data (period of 1991-2012) were explained by the Year + railway traffic + Frost day + 331 T.July. AIC-based model selection (see supplementary material S4) indicated that this full model 332 333 is the most parsimonious choice. This model explained 93% of the variance in the population 334 size (however, this high value should be considered with caution because the model is at the 335 limit of the over-parameterization); 68% of this explained variance was explained by the year 336 trend, 28% by the railway traffic presence or absence, 3% by number of frost day in January and 337 less than 1% by the temperature in July of the previous year (these results are consistent with the Hierarchical Partitioning results, see supplementary material S3). The number of common 338 339 Pipistrelle bats in the hibernaculum significantly declined over the period considered -1991 to 2012-(Table 1, *F*_{1,19}=1416.85; *P*<0.001; *θ* =-0.078 [SE] 0.006), which corresponds to drastic trend 340 341 of population decline: 7%/year. In addition, we detected a significant negative effect of the railway traffic (*F*_{1,18}=985.84; *P*<0.001; *θ*=-0.781 [SE] 0.100), a positive effect of number of frost 342 days in January (F_{1,17}=116.82; P<0.001; B=0. 0.021 [SE] 0.007 Fig. 5, and positive effect of 343 temperature recorded in July the year prior to the winter bat counts, $F_{1,16}$ =34.02; P<0.001; β =0. 344 345 0.027 [SE] 0.018, Fig. 5).

346

347

Approximate position of figure 5

348

349 **DISCUSSION**

350 Despite the good availability of multi-site, long-term count data for several bat species

351 (Mitchell-Jones *et al.* 2007), the dynamics of most monitored populations remain largely

352 unexplored because of the conceptual and technical difficulties in discriminating the effects of

353 multifactorial environmental disturbances, protocol artefacts and dispersal.

We propose that using combinations of population dynamics modelling even with demographic parameters from the literature and statistical analyses helps with identifying the biological and methodological effects in census analyses. Despite the managers' perception that population sizes show chaotic variations over time, we show that counts made by volunteers in hibernaculum can provide useful information on local population dynamics and on the environmental factors underlying these dynamics.

360

361 Demographic processes underlying the observed local dynamics

Using the demographic data from the literature (i.e., generic data) in combination with count 362 and local environmental data (i.e., specific data) allowed us to define the realistic boundaries of 363 364 local population variations due to intrinsic dynamics. Despite an early age at sexual maturity, the low annual fecundities reported in the literature lead to a relatively long generation length 365 366 (3.9 years for the median scenario) and a relatively high sensitivity of the population growth rate to adult survival (elasticity values are respectively 0.73 and 0.27 for adult survival and 367 fecundity, for the median scenario). Even when considering an optimistic demographic scenario 368 369 that assumes high survival and fertility rates and including variations in the annual growth due 370 to demographic stochasticity, our results suggest that some of the strong year-to-year increases in the population size could not have resulted merely from intrinsic dynamics (see Figs. 2 and 3) 371 372 and might imply immigration events from unknown hibernacula. In particular, we suggest that the most rapid year-to-year increase in size (i.e., the 1993-1994 period) is linked to a change in 373 the attractiveness of this hibernaculum as a result of the end of the exploitation of the railway 374 375 line. This observation underlines the impact of disturbance on the population sizes of 376 hibernating bats. Many studies have shown the importance of disturbance on colony dynamics and sustainability (Mc Craken 1989; Speakman et al. 1991; Simon et al. 2004; Mitchell-Jones et 377

378 al. 2007). The very short time delay between the end of the disturbance and its consequence on the population dynamics underlines the very active prospecting behavior of bats in choosing 379 a favourable hibernaculum each winter (see Simon et al. 2004). More generally, our analysis of 380 381 an adjusted scenario (which uses median demographic data from the literature that was 382 modified to fit the growth rate observed during the 1995-2012 period) suggests that 383 emigration-immigration processes might have preponderant effects on population dynamics. 384 With such an approach we were able to detect that some strong year-to-year increases could not have resulted only from intrinsic dynamics and implied immigration events. However, for 385 smaller increases that were consistent with the maximum intrinsic population growth, we were 386 not able to distinguish the roles of intrinsic growth and immigration. Moreover, strong year-to-387 year reduction in population size can be related to either strong mortality or massive 388 389 emigration, with little opportunity to distinguish the source of this variation.

390

391 Population trends and environmental effects

After removing the drastic trend of population decline (7%/year) over the 1991-2012 period, 392 we detected a negative effect of the railway traffic on number of common pipistrelles and we 393 394 further revealed some effects of the winter and summer temperatures. However the incidence of meteorological variable is very thin compare to yearly trend and presence of railway traffic. 395 396 The winter effect is presumably associated with aggregation in favourable hibernacula when 397 the winter conditions are severe, which suggests that common pipistrelles might be able to shift their winter site within the hibernating season. Several counts within the winter could, 398 399 therefore, bring useful information on the extent of movements between cavities. Furthermore, correlations between the winter weather conditions and population variations 400 401 within the cavity network could inform us about the specific role of each type of cavity. The

402 slightly positive influence of the temperature and slightly negative effect of rainfall during the month of July (note that T_{July} and P_{July} are not correlated) is more likely linked with reproductive 403 parameters (i.e., fledging success or/and the survival of young). This observation fit well with 404 405 the Pipistrelle's cycle of life: July is a crucial period because it corresponds to fledging of young 406 Pipistrelle and the dispersal of the breeding colony (Arthur and Lemaire 2009). 407 Importantly, these two seasonal environmental effects were not significantly correlated with 408 the yearly trend, which emphasises that the observed decline of the pipistrelle population is 409 not only explained by yearly climate variations.

410

411 Limitations and perspectives

412 Tuttle (2003) underlined that complete enumeration of hibernating bats is possible in situations 413 where numbers of bats are not extremely large and cavities lack great surface irregularities. In 414 larger hibernacula with bats in roosting dense clusters, the only feasible techniques involve estimating individual densities (Thomas and Laval 1988). However, the reliability of such 415 techniques is still unknown (Tuttle 2003). In addition to the differences in habitat 416 417 characteristics, different species of bat exhibit different behaviors in winter. Some species are easily detectable, such as *Rhinolophus ssp* (which hibernates suspended to ceiling), whereas 418 other species such as Myotis daubentonii, M. bechsteinii and M. nattereri hibernate in deep 419 420 cracks and are less detected. The hibernaculum studied here is an ideal case since available 421 slots are standardized (joints between concrete slabs), which allows detecting almost all individuals. 422

We did not consider immigration-emigration in our population dynamics models, since we had
no means to add realistic constraints to immigration and emigration rates. Unlike the intrinsic
dynamics, which are limited by biological, species-specific constraints (i.e., maximum

reproductive output of the species), immigration-emigration processes mainly depend on the
context of the focal population (e.g., distance to other populations). Such wider context could
be examined through a meta-population approach in which surrounding populations
interacting with the focal one are explicitly modelled (in such case, emigration would be
constrained by the available pool of migrants in the meta-population). Although our dataset
does not allow us such larger scale modelling for the moment, current national hibernation
surveillance scheme could provide such information in the future.

433 **Recommendations**

434 To conduct a meaningful analysis of time-series count data of bat in hiberculatum, it is

435 necessary

436 (i) to include the local knowledge of the people involved in the field surveys in the analyses, for

437 example, information on the existence of disturbances (the type, magnitude and frequency) or

the implementation of site protection measures should be recorded and integrated with the

439 count data;

440 (ii) to include the winter weather conditions, which are very likely associated with aggregation

441 behavior (e.g., the number of frost day in January, the average temperature of January),

although they do not necessarily imply demographic consequences;

(iii) to include the spring-summer weather information collected during the season before the

444 winter counts (e.g., the July rainfall), as such data are expected to affect reproductive

445 parameters, which might, in turn, influence population dynamics and

(iv) our analysis also strongly suggests that a multi-site approach using count data might allow

447 strong inferences to be drawn on the intrinsic and environmental processes underlying

448 dispersal and population dynamics.

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454

455

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592 Table 1: Effects of disturbance and meteorological variables on the estimates of number of bat. T.day is the local ambient temperature at the time of the count; T.January is the average 593 temperature in January; T.winter is the average temperature in winter (December-January-594 595 February); T.Anomalies is the temperature anomaly observed in January relative to the average 596 temperatures over the 1951-1989 period; Frost day is the number of frost days in January; P.June 597 is the rainfall recorded in June of the previous year; P.July is the rainfall recorded in July of the previous year; T._{June} is the average temperature in June of the previous year; T._{July} is the average 598 temperature in July of the previous year; Slope of the effect (β) are given for significant test. 599

600

					Winter va	riables			Summer	variables	
	Year	Traffic	T. _{day}	T.January	T.winter	T.Anomalies	Frost	P.June	P.July	T. June	T.July
		railway					day				
Effect of each	<i>Z</i> =-4.753	Z=-0.085	<i>Z</i> =0.34	Z=-0.68	Z=0.73	Z=0.06	Z=0.66	Z=0.41	<i>Z</i> =-0.04	Z=-1.39	Z=0.09
variable	<i>P<0.0001</i> β=-0.046	P=0.93	P=0.74	P=0.50	P=0.47	P=0.54	P=0.51	P=0.69	P=0.97	P=0.16	P=0.92
	AIC=1574	AIC=3154	AIC=3083	AIC=1574	AIC=2907	AIC=2467	AIC=2922	AIC=3130	AIC=3155	AIC=1574	AIC=3153
Effect of each v	ariable after	F _{1,19} =955.6	F _{1,18} =36.11	F _{1,19} =106.0	F _{1,19} =35.18	F _{1,16} =5.95	F _{1,18} =113.81	F _{1,19} =16.01	<i>F_{1,19}</i> =16.01	<i>F_{1,19}</i> =144.3	F _{1,19} =12.19
adjustment to th	he year effect	3	P<0.0001	1	P<0.0001	P=0.015	P<0.0001	P<0.0001	P<0.0001	2	P=0.001
		P<0.001	<i>6=-0.009</i>	P<0.0001	<i>β=-0.037</i>	<i>β=-0.037</i>	<i>6=0.021</i>	<i>β=</i> -0.001	<i>β</i> =-0.001	P<0.0001	<i>β=</i> 0.017
		β=-0.754	AIC=1532	в=-0.043	AIC=1533	AIC=1529	AIC=1454	AIC=1558	AIC=1560	6=0.088	AIC=1564
		AIC=621		AIC=1470						AIC=1432	
Effect of each va	riable after adju	ustment to	F _{1,17} =15.65	F _{1,18} =83.35	<i>F_{1,17}</i> =31.96	F _{1,15} =1.482	F _{1,17} =116.82	F _{1,18} =0.05	F _{1,18} =16.56	F _{1,18} =0.07	F _{1,18} =19.63
the year effect a	nd railway traff	ic	P<0.0001	P<0.0001	P<0.0001	P=0.22	P<0.0001	P=0.82	P<0.0001	P=0.79	P<0.0001
			в=-0.006	<i>β=</i> -0.038	<i>β=-0.037</i>		<i>в=0.020</i>		<i>β</i> =-0.001		<i>β=</i> 0.019
			AIC=597	AIC=539	AIC=552	AIC=553	AIC=467	AIC=622	AIC=606	AIC=622	AIC=603

Figure 1: Life cycle of the common pipistrelle (*Pipistrellus pipistrellus*) modelled with Leslie's matrix, where N_i is the number of individuals of age, i, σ is the sex ratio, F is the female fecundity (number of young emancipated per female), S₀ is the first year survival probability and S_{adt.} is the survival probability of adults.





611

Figure 2: Values of the deterministic growth rate (λ) as a function of the adult survival and fecundity rates. Dark grey shading indicates high value of λ. Other demographic parameters are σ =0.5; S₀=0.53.

615



Population Growth rate

618

Figure 3: Comparison between the expected and observed annual growth. Expected distributions (lines) are based on 1000 stochastic trajectories from the stochastic individualbased model (see the Materials and Methods for details). Observed values (circles) correspond to the N_t/N_{t-1} ratios observed over the 1991-1995 (black circles) and 1995-2012 (grey circles) periods. Pessimistic scenario, S₀=0.53, S_{ad}=0.37, F=0.58; median scenario, S₀=0.53, S_{ad}=0.59, F=0.78; Optimistic scenario, S₀=0.53, S_{ad}=0.8, F=0.99; Adjusted scenario, S₀=0.57, S_{ad}=0.69, F=0.84.

626



627



Figure 4: Change in bat abundance (as estimated from winter counts) between 1991 and 2012

Figure 5. Effect of number of frost day in January (upper panel) and July temperature (lower panel) on changes in the observed population size. Values of July temperature were recorded in the year prior to the winter bat counts. Climate variables are adjusted to the year effect and railway traffic effect, so changes in the observed population size are expressed in residuals from the modelling.



639 640 641

642 **Supplementary material - Appendix S1**

Characteristics of the hibernaculum of pipistrelle bats (Pipistrellus pipistrellus) located in a 643 railway tunnel (600 meter long, Picture 1). The counting method is a direct visual count of 644 individuals who are located in shallow notches. These notches are expansion joint vertical 645 depth of 10 cm between concrete slabs (Picture 2), This is close to some natural cracks (Picture 646 3) The count procedure leads thus to very thin measurements errors, because all individuals are 647 648 observable.

649



Picture railway tunnel where 1: the hibernaculum of pipistrelle bats is located

650



Picture 2: pipistrelle in a notche between Picture 3: pipistrelle in a natural cracks concrete slabs.



Supplementary material - Appendix S2: Table of correlations (Spearman's rank test) between the pairs of weather variables used in the modelling. T. day is the local ambient temperature at the time of the count; T.January is the average temperature in January; T.winter is the average temperature in the winter (December-January-February); T.Anomalies is the temperature anomaly observed in January relative to the average temperatures over the 1951-1989 period; Frost day is the number of frost days in January; *P. June* is the rainfall recorded in June of the previous year; P.July is the rainfall recorded in July of the previous year; T.June is the average temperature in June of the previous year; T.July is the average temperature in July of the previous year; Significance tests are indicated in bold characters (α =0.05).

			Winter vai	Spring variables					
	T. _{day}	T.January	T.winter	T. Anomalies	Frost day	P.June	P.July	T.June	T.July
Year	p=0.63	p=-0.83	p=-0.87	p=-0.63	p=-0.45	p=0.54	p=0.45	P=0.004	P=0.79
	rho=-0.12	rho =-0.05	rho=0.04	rho=0.12	rho=-0.18	rho=-0.14	rho=0.18	rho=0.61	rho=0.06
T day		p=0.11	P=0.52	p=-0.03	P=0.34	p=0.94	p=0.15	p=0.56	p=0.55
uuy		rho=0.39	rho=0.16	rho=0.50	rho=-0.24	rho=-0.02	rho=0.35	rho=-0.14	rho=-0.14
T January			P<0.001	p=0.02	P=0.002	p=0.37	p=0.66	p=0.85	p=0.73
. isunuury			rho=0.76	rho=0.53	rho=-0.67	rho=0.22	rho=0.11	rho=-0.05	rho=-0.08
T winter				P=0.01	P<0.001	p=0.31	p=0.23	p=0.45	p=0.78
··winter				rho=0.57	rho=-0.72	rho=0.25	rho=0.28	rho=0.18	rho=-0.06
T. Anomalies					P=0.28	p=0.39	p=0.10	p=0.80	p=0.62
/ montanes					rho=-0.26	rho=0.21	rho=0.38	rho=-0.06	rho=0.12
Frost day						p=0.93	p=0.20	p=0.15	p=0.47
						rho=-0.02	rho=-0.31	rho=-0.34	rho=0.18
Plune							p=0.04	p=0.40	p=0.84
, sunc							rho=-0.45	rho=-0.19	rho=-0.05
P. luly								p=0.82	p=0.95
- isuly								rho=0.06	rho=-0.01
T. lune									p=0.77
···									rho=-0.07

^{*} Spearman's rho showed no important correlations between variables (correlation coefficient
|rho| < 0.5; Freckleton 2002).

666 Reference

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 regression. Journal of Animal Ecology 71: 542-545.

676 Supplementary Material. - Appendix S3 of The adjusted scenario.

We aimed to examine an additional scenario (hereafter, adjusted scenario) as follows: (i) with the same demographic structure and properties as our median scenario (i.e., a scenario based on the most reliable demographic parameters provided by the literature) and (ii) yielding a growth rate similar to the one observed over the 1995-2010 period (0.93).

For this purpose, we kept the same transition matrix structure as the median scenario (see Figure S2.1 below, this model yields a deterministic growth $\lambda_{median}=0.797$), and we slightly modified the demographic parameters to obtain $\lambda_{adjusted}=0.93$.

684 This can be achieved by multiplying the transition matrix by a constant term, α = 685 $\lambda_{adjusted}/\lambda_{median}$ =1.17, or by multiplying the demographic parameters such that all of the non-686 zero terms of the transition matrix are multiplied by α (Caswell 2001).

687

688 **Figure S2.1.** Transition matrix corresponding to the median scenario. σ =0.5; s₀=0.53; s_{ad}=0.59; 689 F=0.78.



 $\begin{pmatrix} \sigma s_0 F & \sigma s_0 F \\ s_{ad} & s_{ad} \end{pmatrix}$

699 We obtained the following parameters for the adjusted scenario:

700 s0'= √α×0.53=0.57

701 sad'= α×0.59=0.69

702 $F' = \sqrt{\alpha \times 0.78} = 0.84$

703

This new set of parameters leads to a population with similar demographic properties as the median scenario (i.e., the same generation time and proportions of age classes at demographic

equilibrium) but a different deterministic growth rate ($\lambda_{adjusted}$ =0.93).

709	Supplementary material - Appendix S4: Model selection table

Model	AIC
	1574
	3154
Count ~ T.day	3083
	1574
Count ~ T.winter	2907
Count ~ T.Anomalies	2467
Count ~ Frost day	2922
Count ~ P. _{lune}	3130
Count ~ P. _{July}	3155
Count ~ T. June	1574
Count ~ T. _{July}	3153
Count ~ Year + Traffic	621
Count ~ Year + T. _{day}	1532
Count ~ Year + T. _{January}	1470
Count ~ Year + T. _{winter}	1533
Count ~ Year + T. _{Anomalies}	1529
Count ~ Year + Frost day	1454
Count ~ Year + P. _{June}	1558
Count ~ Year + P. _{July}	1560
Count ~ Year + T. _{June}	1432
Count ~ Year + T. _{July}	1564
Count ~ Year + Traffic	3090
Count ~ Traffic + T. _{day}	3079
Count ~ Traffic + T. _{January}	3090
Count ~ Traffic + T. _{winter}	2909
Count ~ Traffic + T. _{Anomalies}	2443
Count ~ Iraffic + Frost day	2921
Count ~ Iraffic + P. _{June}	3128
Count ~ Traffic + P.July	3155
Count ~ Traffic + T. June	2088
Count \sim Voor + Traffic +T	5154
Count \sim Voor + Traffic +T	520
Count ~ Vear + Traffic +T \therefore	552
	552
Count ~ Year + Traffic +Frost day	467
Count \sim Year + Traffic +P. lung	622
Count ~ Year + Traffic +P. http://www.	606
Count ~ Year + Traffic +T. June	622
Count ~ Year + Traffic +T. July	603
Count ~ Year + Traffic +T. _{day} + P. _{lune}	598
Count ~ Year + Traffic +T. _{day} + P. _{July}	593
Count ~ Year + Traffic +T. _{day} + T. _{June}	597
Count ~ Year + Traffic +T. _{day} + T. _{July}	577
Count ~ Year + Traffic + T. January + P. June	536
Count ~ Year + Traffic + T. _{January} + P. _{July}	533
Count ~ Year + Traffic + T. _{January} + T. June	541
Count ~ Year + Traffic + T. _{January} + T. _{July}	516
Count ~ Year + Traffic + T. _{winter} + P. _{June}	553
Count ~ Year + Traffic + T. _{winter} + P. _{July}	539
Count ~ Year + Traffic + T. _{winter} + T. June	553
Count ~ Year + Traffic + T. _{winter} + T. _{July}	517
Count ~ Year + Traffic + T. _{Anomalies} + P. _{June}	553
Count ~ Year + Traffic + T. _{Anomalies} + P. _{July}	537
Count ~ Year + Traffic + T. Anomalies + T. June	555
Count ~ Year + Traffic + T. _{Anomalies} + T. _{July}	524
Count ~ Year + Traffic + Frost day + P. June	469
Count ~ Year + Irattic + Frost day + P. July	465
Count ~ Year + Irattic + Frost day + T. June	467
Count ~ Year + Traffic + Frost day + T. _{July}	435

711 Supplementary material - Appendix S5:

Figure S3.1. Identification of the most likely causal factors of temporal variations in recorded
 bat abundance, using hierarchical partitioning (R package *hier.part*), including (A) All variables;
 (B) variables adjusted to the yearly trend and (C) variables adjusted to both yearly trend and

- 715 railway traffic effects.
- 716

