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## Gene flow does not prevent personality and morphological differentiation between two blue tit populations

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**Gene flow does not prevent personality and morphological differentiation between two blue tit populations.**

Journal:	<i>Journal of Evolutionary Biology</i>
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Keywords:	Cyanistes caeruleus, genetic divergence, local adaptation, personality, Phenotypic plasticity, Qst - Fst

Montreal, 8 april 2018

Dear editor,

Please find enclosed a manuscript entitled “Gene flow does not prevent personality and morphological differentiation between two blue tit populations” that we wish to submit for publication in your journal.

We believe our work will be of interest to the broad readership of your journal because it challenges the general idea that gene flow limits populations genetic divergence and the importance of plasticity versus genetic effects in shaping population divergence for personality traits. We report on a common garden experiment exploring the genetic basis of phenotypic differences observed in the wild for personality, physiological and morphological traits between two blue tit (*Cyanistes caeruleus*) populations inhabiting contrasting habitats separated by a small spatial scale and connected by gene flow. We raised nestlings originating from the two habitats in aviaries for up to five years and then compared their adult phenotypes. Our results revealed differences similar to those found in the wild, suggesting a genetic divergence for all traits. In addition,  $Q_{st} - F_{st}$  comparisons revealed that the observed quantitative genetic divergence is likely the result of contrasting selection pressures rather than of neutral processes. Our study is one of the first to report  $Q_{st} - F_{st}$  comparisons for personality traits and suggests that genetic divergence is possible at a small spatial scale for behavioural and physiological traits. Such small scale evolution of animal personality and physiology has rarely been reported and shows that population genetic divergence is possible at a small spatial scale for traits generally considered less prone to genetic divergence.

We would like to thank you and the reviewers for providing constructive comments on the first version of the manuscript that we think have greatly improved our work. Please find attached a revised version of our manuscript and a detailed answer to all comments.

We agree for the dataset to be shared on Dryad after the paper is published. The manuscript is not under consideration for publication in another journal. All persons entitled to authorship have been named and have approved the submission of this version of the manuscript. The manuscript is 5978 words.

We hope you will consider for publication in your journal this revised version of our manuscript and we look forward to your assessment.

Best regards,

The authors

**Title**

Gene flow does not prevent personality and morphological differentiation between two blue tit populations.

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**Running head**

Divergence between blue tit populations

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(NWO-VENI grant 863.09.011) to SPC. GDM received PhD fellowships from the Fonds de Recherche Québec Nature et Technologies and the Natural Science and Engineering Research Council of Canada. We declare no conflict of interests.

Dear Dr Klingenberg,

Thank you for your feedback regarding our manuscript # JEB-2017-00630. We are very thankful for the new comments you provided as well as those from the two reviewers. We outline below how all these comments have been taken into account to provide a revised version of our manuscript that will now, we hope, be accepted for publication in Journal of Evolutionary Biology.

### Comments from Editor

In my own reading of your paper, I have come across the act that you used the method of random skewers for comparing P matrices. This method has been criticised for lack of power and other statistical properties. Very recently, such a critique has been published by Jim Rohlf. Please have a look at this paper and think how the criticism applies to your analysis. I am not prescribing a particular course of action to you, but I would like you think about it as part of the further revisions.

*>> Since the Qst, Pst and Fst comparison is more robust and provides stand-alone results that do not necessitate any complementary analysis, we have decided to remove the matrix comparison done with the random skewer method that has been recently criticized. This does not change any conclusion from the manuscript, but it should remove any doubt regarding random skewers.*

+++++

### Comments from reviewer 1

I have only two comments:

1. - The first one refer to the use of the term phenotypic plasticity (or plastic response) as synonym of environmental factors affecting phenotypes. I think it may confound readers and, whenever possible, I suggest the used of environmental factors throughout the manuscript. This is because phenotypic plastic response to environmental condition may have a genetic component (see for instance Charmantier et al. (2008), Science) and, thus, phenotypic plastic responses may be due not only to environmental factors, but to genetic factors. Remember that the reaction norms may have strong genetic components, but you use plasticity as solely reflecting environmental effects. Traditional terms used in quantitative genetics are genetic, environmental and maternal effects and, thus, would strongly recommend the used of "environmental factors" instead of plasticity.

*>> We partly disagree here with the reviewer's opinion that our use of the term "plasticity" brings confusion in all its use. We now clarify that by plasticity we refer to "the adjustment of individual phenotypes in response to environmental factors" (definition provided in line 30). Note that plasticity does not result in a genetic change, however variance in plasticity can have a genetic origin when plasticity is heritable and lead to evolution of different plasticity across populations. Although this is a side issue to our study, we now explain this in lines 65 to 68. Finally, when addressing the issue of the Qst/Pst comparison, we have followed the reviewer's advice and changed plasticity to environmental factors (lines 119 and 454).*

2. - The second comment refers to the scarce information provided for used mixed model. You should explain how mixed models accounted for the structure of the data set (random blocks) (see for instance Schielzeth, H. & Forstmeier, W. (2009). Conclusions beyond support: overconfident estimates in mixed models. Behav Ecol 20: 416-420). Do you used random

intercept, random slope, or random intercept-and-slope models? Was identity of rearing brood nested within identity brood of origin? In the case of bird identity, this random factor should be nested within the interaction between brood of origin and rearing brood (or, in the case of a completely hierarchized structure of random factors, at the lower level). Remember that a correct definition of hierarchized random blocks is essential for proper estimations of genetic and environmental factor explaining phenotypic variation.

>> *Thank you for pointing out this imprecision. We used random intercepts for bird identity, brood of rearing and brood of origin. We now clarified this (L249 to 253).*

## Comments from reviewer 2

Comments on your written responses:

1. Re. my comment 12: I meant that care takers may not have fed blind with respect to origin. While I don't really think there are box/nest/location, date or care taker effects, some readers may disagree, so perhaps it is best to mention these possibilities and state why you discard it. I think it is always a strength of a paper if it presents (and if space permits) discusses its potential weaknesses, as long as it doesn't distract.

>> *The populations that we study here differ in many traits, one of them being phenology (i.e. timing of reproduction). Because of this timing difference, raising the chick at the same time is impossible. We have made every effort possible for keeping the birds of the two populations into the same conditions, but for sure there are many little, sometimes unidentified, differences, that were impossible to control. Temperature was certainly a bit different (later in the season for the Fango birds), as well as humidity, photoperiod, they travelled in a different boat, etc. among which are the rearing conditions (caretakers, etc). Thus, it is impossible to pinpoint and inventory the exact differences that might have been of relevance to the birds. We have therefore clarified this timing difference between the populations (see L151 to 157), but we believe it is not possible to discuss it much further since we do not know if there were any relevant difference (and which ones) that the birds could have cued on despite our efforts to homogenize their environments and experiences.*

2. Re. comment 41: you don't address the concern of multiple testing, so no mention of the need for correction for a large number of tests, which inflate the probability to find significant results. So Bonferroni correction or something similar might be called for.

>> *Bonferroni corrections have been criticized for being too conservative (Moran 2003). However, applying a Bonferroni correction to our results would lead to the same conclusions as we found a  $p$ -value  $< 0.01$  for all study traits except body mass. If the editor deems it useful, we can add this information in the results section but prefer not to.*

*Moran, M. D. (2003). Arguments for rejecting the sequential Bonferroni in ecological studies. *Oikos*, 100(2), 403-405.*

And instead of Wright, I meant Fisher 's method ([https://en.wikipedia.org/wiki/Fisher%27s\\_method](https://en.wikipedia.org/wiki/Fisher%27s_method)), which to my opinion can also be used as a within-study meta-analysis. You repeatedly test the same hypothesis, just changing the response variable. So you could apply Fisher's method to the  $p$ -values obtained for the habitat of origin tests.

>> *We now use this test. We provide this information and the results of this test in lines 262 to 264 and 306 to 308 ( $p$ -value =  $3.675 \times 10^{-8}$ ).*

3. Re. comment 46: with evergreen habitat being atypical, I mean not the habitat itself but in the evolutionary history of the Blue tit, assuming that the species is largely adapted to deciduous trees and that colonisation of evergreen forest is relatively recent, or that gene flow will be mostly from deciduous to evergreen populations (given the relative population sizes of Blue tits in each of the habitats at a large, European?, scale). Then, even if Blue tits in evergreen habitats are exposed to specific selection pressures and have actually evolved in that direction, that doesn't mean that they are now adapted – there may still be an evolutionary lag. In fact, if you found directional selection on traits \*within\* a habitat, this would confirm local maladaptation. Whether or not size-corrected body mass is a good measure of overall condition (whatever that is) in ad libitum fed, captive birds, is debatable, but I'm thinking of effects that are more general for maladaptation, such as maternal stress hormones or lack of micronutrients deposited in eggs affecting overall development (body, physiology, brain, behaviour). Again, I don't know if this is likely, but if you think it is not, then explaining this in the paper might be the best way forward.

>> *We agree that ongoing selection can be a sign of maladaptation. We also agree that we cannot completely exclude that early environmental effects such as maternal effects occurring before and soon after the chicks were sampled from their nest were at least partly responsible for the observed patterns. We discussed this possibility in length in lines 358 to 377.*

Comments on the new version:

4. L 36: Edelaar & Bolnick 2012 TREE is also a useful citation in this section

>> *We added this citation (L36)*

5. L 53: and/or (you don't need both)

>> *We modified the sentence (L51)*

6. L 54: also Richardson et al. 2014 TREE microgeographic divergence

>> *We added this citation (L52).*

7. L 67: given that you (correctly) scale you question to the dispersal distance, whether a species is highly mobile or not is now unimportant, so remove that part after the comma.

>> *We changed the sentence (L65).*

8. L 110: I like the layered approach to the issue of divergence in this paper, with three different analyses/data sets.

>> *Thank you for the positive comment.*

9. L 110: in this section you need to introduce the issue of Pst and Qst, explain what Pst is (not so commonly known), and why/that you will assume that the Pst of lab-reared birds should approximate Qst to a greater degree than the Pst of wild birds. (see L 106-107, where you do mention that P approximates G), if there is plasticity in the wild.

*>> We added some details and explanations on this topic in L103 to L122.*

10. L 144: do you actually know this? (I assume the birds were ringed)

*>> We meant they were assigned randomly to a cage (L148).*

11. L 145: fed ad libitum (saturation)?

*>> ad libitum means that there was always food and water available for the birds. We never force-fed any bird, neither were they food-restricted. This term is commonly used in the literature.*

12. L 153: tarsus at which age?

*>> Adult (> 1 year of age), we now provide this information (L164).*

13. L 188: blind with respect to what?

*>> We now provide this information (L199).*

14. L 211: what percentage of SNPs did you remove for being potentially under selection? I suppose this must be well under 1%. Nonetheless, many of these outliers will be false positives, i.e. highly divergent SNPs that actually are neutral and therefore should not be filtered out. So you are caught between a rock and a hard place. Perhaps, to obtain a conservative (upper) estimate for Fst, you could also report what value you obtain without this filtering step, so just overall genomic divergence.

*>> We removed 0.7% of the total number of SNPs. We now provide this information (L223). Including or excluding SNPs putatively under divergent selection (Fst of 0.004 and 0.006 respectively) provided high Qst/Fst ratio in both cases. Excluding loci putatively under selection therefore had little effect on the general conclusion, (which is expected in a context of putatively highly polygenic traits and very low genetic differentiation among populations, resulting in few outliers SNPs of large effects and high differentiation).*

15. L 245: check use of singular/plural

*>> Done (L257)*

16. L 265-266: so move this introduction up, to line 110

*>> Done (L114 to 117).*

17. L 278: habitat of

*>> Done (L279-280)*

18. L 279: and individual in the case of body mass? So basically, all random effects variances plus residual variance?

>> *Yes, we now provide this information (L280-281)*

19. L 280: this is the Pst for wild birds? Why not include the random effect of brood?

>> *These birds were measured once adult, for many of them we did not know the brood of origin.*

And I guess observer is random here, because many people were involved?

>> *Yes, L288*

But did you then include this observer variance component into the within-population variance? (For Pst you don't specify which variances were used to calculate the within-pop variance component).

>> *Yes, we now provide this information (L287 to 289)*

This to me would not seem correct, since observer effects are not within-population genetic variation (as mentioned before in my comment 34 of the previous revision). The same might actually be true for rearing brood effect: this could be indirect genetic effects, but could also be purely or mostly environmental. And the same issue for the residual variance: to what extent is this genetic variance, or just unexplained environmental noise around the genotypic value? I think you now assume it is fully genetic. (Note that I'm partly disagreeing with myself, re. comment 34 last review).

What I'm missing in this section is an explanation/justification of which variables are or are not included in the between- and within-population variance components, and methodological/theoretical citations to back this up. And then the effects of the assumptions and decisions on results made should be discussed later on.

>> *We rewrote partly this paragraph and hope that it is now clearer (L275 to 289). For Qst calculation, following the reviewer's comments, we decided to calculate  $\sigma_B$  as the variance attributable to habitat of origin and  $\sigma_W$  as the residual variance (or for body mass as the sum of the variance attributable to the residual and to the individual identity). We did not include any broods effects in these models anymore because the variance attributable to the brood is also attributable to the population of origin in our case. These changes did not affect our conclusions. We changed the results in Table 2 and S5.*

*In our opinion, the observer effect should not be included in the between population variance for Pst calculation but could be considered as residual variance in the context of our analysis. This is why we calculated  $\sigma_B$  as the variance attributable to habitat of origin and  $\sigma_W$  as the sum of the variance attributable to the observer and to the residual variance and the individual effect (L287 to 289).*

20. L 285: I think you mean slightly informative priors? Slightly UNinformative means very informative.

*>> Yes, we changed the sentence (L292)*

21. L 286: I'm surprised you stay with a burn-in of 500 iterations, when running a model of 10 million iterations. Normally the burn-in is more like a third or half of the total iterations. With a thinning of 200, you now discard only the first 2 effective samples, and keep the remaining 50,000 samples. So basically, you have no burn-in period. I propose a burn-in of at least 1 million iterations, unless you know that your chains converge very quickly (in which case you don't need 10 million iterations, which looks like overkill anyway).

*>> We compared models with different burn-in period, iterations and nu and found no important difference between these models and really small autocorrelation (L291 to 297).*

22. L 295: maybe this information goes better with line 278?

*>> Yes, we agree, we changed the location of the sentence to L283-284.*

23. L 331: I still don't get this result: you state that the two matrices are more similar to each other than two random matrices. This is very unsurprising. And irrelevant. What we want to know is if they are dissimilar, as you claim you will test in L 108. But you don't do that. I mentioned this before. I think this analysis is not relevant at all – look for an analysis that will tell us if the matrices are identical or not (e.g. if your correlation of 0.9 is different from 1.0, not if it is different from 0.0, but probably random skewers is not the thing to do).

*>> This comment is in line with the Editor concern regarding the random skewer method. As explained in our response to the Editor, we have removed this comparison.*

24. L 338: extent

*>> We removed this part of the sentence*

25. L 343: suggests

*>> Modified (L343)*

26. L 349: replace cannot be for is not

*>> Modified (L349)*

27. L 351: indicates

*>> Modified (L351)*

28. L 353: remove genetically

*>> Removed*

29. L 354: replace cannot for does not

>> *Modified (L355)*

30. L 355: replace could be for are

>> *Modified (L356)*

31. L 360: remove genetic

>> *Removed*

32. L 367: mention here also the significant effects you found for rearing brood, suggesting the acting of such early effects

>> *We discuss the rearing brood effects in lines 433 to 443.*

33. L 387: or that you removed high  $F_{st}$  neutral ones

>> *We now mention this possibility (L387) but we choose to keep it in parenthesis since this issue is not important and not discussed further.*

34. L 388: but you should not use microsatellites for  $Q_{st}$ - $F_{st}$  comparisons, see several papers on this by Jost, Edelaar and other authors, in *Molecular Ecology* (partly cited in Leinonen et al. 2013).

>> *Yes, this is why we did not use microsatellites in this study, the microsatellite study is mentioned to discuss the  $F_{st}$  level.*

35. L 390: this statement needs to be re-evaluated after checking what SNP filtering does, and whether the  $Q_{st}$  calculations are changing based on my comments above.

>> *This statement still holds after re-evaluation of the consequences of the filtering.*

36. L 397: paper by Dingemanse et al. on predator-presence related population divergence in stickleback personalities also comes to mind (*J Anim Ecol*?)

>> *We added this reference (L398)*

37. L 400: studies

>> *Modified (L419)*

38. L 404: better refer to the  $Q_{st}$  values in Table 2?

>> *Modified (L404)*

39. L 413: I agree, but also mention/discuss if you would expect the traits you used to be correlated in your variance covariance matrix (which they are hardly)

>> *As explained in our response to the Editor, we have removed the comparison between the covariance matrices from the manuscript.*

40. L 428: again, you tested if they are similar, but you should test if they are different

>> *This section was removed.*

41. L 435: associated with

>> *Modified (L431)*

42. L 445: remove (2 to 12 days) since this is irrelevant – statistical power is the issue, not the time frame.

>> *We agree, we changed the location of the parenthesis in the sentence (L441).*

43. L 447: this is the same as your first argument – effectively not sensitive relative to other environmental effects

>> *We removed this sentence*

44. L 459: again, this may need to be revised depending on any  $Q_{st}$  and  $P_{st}$  recalculations. It is kind of strange that divergence is not reduced under a common environment, assuming that any plasticity in the wild would tend to operate in the direction of the divergent selection between habitats. I think this is also the common observation (often even no remaining divergence in a common environment).

>> *We meant that the environmental effects might not be very high in the wild otherwise we would have found a  $P_{st}$  significantly different from the  $Q_{st}$ . See our response to previous comments about the filtering.*

45. L 463: remove phenotypic

>> *Done.*

46. L 464: replace L 465-466 by: but that genomically are diverged much less.

>> *We prefer the original sentence.*

47. L 466: replace past study on by past results for

>> *Done (L461).*

48. L 470: scales (relative to dispersal ability) for

>> *Done (L464).*

Table S6: for HR and tarsus length, the mean is higher than the credible interval, so this needs to be fixed

>> *We corrected this information (Table S5).*

Table 1: as far as I'm concerned (and you, as you don't test its significance), you can remove the info for the intercepts).

Time of day: based on d.f. you fitted a linear effect – have you checked if the effect is actually non-linear? As you have enough data, perhaps fit time as categorical (by hour for example). Same for age. This might change (improve?) your Qst estimates, see L 338.

>> *We have only fitted here a linear effects.*

+++++

## 1 **Abstract**

2 Understanding the causes and consequences of population phenotypic divergence is a central goal  
3 in ecology and evolution. Phenotypic divergence among populations can result from genetic  
4 divergence, phenotypic plasticity or a combination of the two. However, few studies have  
5 deciphered these mechanisms for populations geographically close and connected by gene flow,  
6 especially in the case of personality traits. In this study, we used a common garden experiment to  
7 explore the genetic basis of the phenotypic divergence observed between two blue tit (*Cyanistes*  
8 *caeruleus*) populations inhabiting contrasting habitats separated by 25 km, for two personality  
9 traits (exploration speed and handling aggression), one physiological trait (heart rate during  
10 restraint) and two morphological traits (tarsus length and body mass). Blue tit nestlings were  
11 removed from their population and raised in a common garden for up to five years. We then  
12 compared adult phenotypes between the two populations, as well as trait-specific  $Q_{st}$  and  $F_{st}$ . Our  
13 results revealed differences between populations similar to those found in the wild, suggesting a  
14 genetic divergence for all traits.  $Q_{st} - F_{st}$  comparisons revealed that the traits divergences likely  
15 result from dissimilar selection patterns rather than from genetic drift. Our study is one of the  
16 first to report a  $Q_{st} - F_{st}$  comparison for personality traits and adds to the growing body of  
17 evidence that population genetic divergence is possible at a small scale for a variety of traits  
18 including behavioural traits.

19

## 20 **Keywords**

21 *Cyanistes caeruleus*, genetic divergence, local adaptation, personality, plasticity,  $Q_{st} - F_{st}$

22

## 23 **Introduction**

24 Understanding the evolutionary causes of phenotypic divergence among populations is an  
25 important aspect of the study of diversity. Environmental heterogeneity can have a major role in  
26 generating phenotypic divergence among populations (Wang & Bradburd 2014). Spatial variation  
27 in selection pressures resulting from such environmental heterogeneity can lead to genotype by  
28 environment interactions for fitness and produce phenotypic and genetic divergence between  
29 populations that can lead to local adaptations (Kawecki & Ebert 2004; Wang & Bradburd 2014).  
30 Spatial heterogeneity in ecological conditions can also favour the evolution of phenotypic  
31 plasticity, *i.e.* the adjustment of individual phenotypes in response to environmental factors  
32 (Pigliucci 2005) and cause phenotypic divergence of populations in the absence of genetic  
33 divergence or local adaptation (Sultan & Spencer 2002; Réale *et al.* 2003; Pigliucci 2005).  
34 Phenotypic divergence of populations can also be produced by non-random dispersal of  
35 individuals between habitat types (Wang & Bradburd 2014). Importantly, phenotypic divergence  
36 of populations does not necessarily involve an adaptive process since phenotypic plasticity and  
37 non-random dispersal can be non-adaptive (Edelaar & Bolnick 2012; Fitzpatrick 2012; Wang &  
38 Bradburd 2014) and can occur in the same or in the opposite direction to genetic divergence  
39 (Fitzpatrick 2012). In addition, strong founder effects or genetic drift can also lead to phenotypic  
40 and genetic divergence of populations (Slatkin 1987). Establishing the relative importance of  
41 environmental versus genetic effects involved in the phenotypic divergence of populations  
42 provides fundamental information about the origin of intra-specific diversity in the wild. In  
43 addition, determining if this divergence is adaptive or the result of neutral processes is essential  
44 because it gives important indications about the eco-evolutionary dynamics of traits and their  
45 evolutionary trajectories.

46  
47 Traditionally, it has been considered that the homogenizing effect of gene flow prevents genetic  
48 divergence of populations (Sultan & Spencer 2002; Lenormand 2002). Thus, most research on  
49 genetic divergence focused on populations separated by large spatial scales or by important  
50 landscape barriers to dispersal (Slatkin 1987; Lenormand 2002). Nevertheless, recent theoretical  
51 and empirical studies revealed that even in the presence of gene flow, phenotypic divergence can  
52 have a genetic origin when there is strong divergent selection and/or non-random dispersal  
53 (Richardson *et al.* 2014; Wang & Bradburd 2014). Despite growing interest for such isolation by  
54 environment, there is little empirical data on the mechanisms underlying the phenotypic  
55 divergence of populations separated by small geographic distances and connected by gene flow.

56  
57 Behavioural traits have often been considered as highly plastic and thus less prone to genetic  
58 divergence. However, several studies are now showing that among-individual differences in  
59 behaviour can be repeatable (*personality*; Réale *et al.* 2007), heritable (van Oers & Sinn 2011),  
60 and related to fitness (Smith & Blumstein 2008) and could thus evolve in response to local  
61 conditions. In this context, an increasing number of studies have compared the personality  
62 phenotypes of individuals inhabiting contrasted ecological conditions (Bell 2005; Quinn *et al.*  
63 2009; Atwell *et al.* 2012; Herczeg *et al.* 2013; Miranda *et al.* 2013; Karlsson *et al.* 2016; Jacquin  
64 *et al.* 2016). However, fewer studies have disentangled the role of plasticity from that of genetic  
65 effects in shaping phenotypic divergence between populations separated by distances that are  
66 within the dispersal ability of a species (Atwell *et al.* 2012; Miranda *et al.* 2013). Note that the  
67 plastic response to environmental factors can itself have a genetic basis, hence plasticity levels  
68 can differ across populations because plasticity can be heritable and evolve differently across  
69 populations (e.g. Laurila *et al.* 2002).

70

71 Previously, we have revealed a phenotypic divergence for personality and morphological traits  
72 between two wild populations of blue tits (*Cyanistes caeruleus*) living in contrasting habitats in a  
73 Mediterranean landscape (Charmantier *et al.* 2016; Dubuc-Messier *et al.* 2017). These  
74 populations occupy habitats and valleys dominated by either evergreen (holm oak, *Quercus ilex*)  
75 or deciduous oaks (downy oak, *Quercus pubescens*) yet are separated only by 25 km, which is  
76 within the typical dispersal range of the species (Tufto *et al.* 2005; Winkel & Frantzen 1991). The  
77 dominant tree species in each habitat and valley is suspected to have an important influence on  
78 blue tits' ecological context that translates into phenotypic divergence between populations for  
79 numerous types of traits despite a spatial proximity and gene flow among them (Charmantier *et*  
80 *al.* 2016). For example, blue tits from the evergreen habitat have a higher adult survival  
81 probability, a lower body mass, a smaller tarsus length, a higher docility (lower handling  
82 aggression), and a slower exploration in a novel environment, compared to birds from the  
83 deciduous habitat (Table S1; Grosbois *et al.* 2006; Charmantier *et al.* 2016; Dubuc-Messier *et al.*  
84 2017). In addition, past studies in this system revealed that small birds (mass and tarsus length)  
85 have a selective advantage in the evergreen habitat (Blondel *et al.* 2002; Teplitsky *et al.* 2014),  
86 suggesting that at least some of the observed phenotypic divergence between habitats could be  
87 adaptive.

88

89 In this study, we used a common garden experiment to assess whether the personality and  
90 morphological divergence between these two blue tit populations could have a genetic basis. We  
91 collected blue tit nestlings from the evergreen and deciduous habitats and raised them for up to  
92 five years in aviaries, subsequently comparing their personality, physiological and morphological  
93 phenotypes once adults. Previous experiments in aviaries on this system have found a genetic

94 divergence between these habitats for life-history traits (Lambrechts *et al.* 1997). Based on these  
95 results, we hypothesized that the phenotypic divergence found previously in the wild for  
96 personality and morphological traits would also reflect a genetic divergence. Therefore, we  
97 predicted that, following the common garden experiment, individuals originating from the  
98 evergreen habitat would show a slower exploration in the novel environment, a higher docility  
99 (lower handling aggression), a smaller tarsus and a lower body mass than individuals originating  
100 from the deciduous habitat. We also compared heart rate during manual restraint of birds  
101 originating from the two habitats, a physiological measure of stress reaction often used in  
102 personality studies (Koolhaas *et al.* 1999).

103  
104 Second, we investigated whether the potential genetic divergence between these habitats could be  
105 attributed to different selection pressures or to genetic drift using a  $Q_{st}$  -  $F_{st}$  comparison approach  
106 (Leinonen *et al.* 2013). A trait  $Q_{st}$  measures the amount of additive genetic variance among  
107 populations relative to the total genetic variance in the trait (Leinonen *et al.* 2013).  $F_{st}$  is the  
108 equivalent measure for neutral molecular variance (Weir and Cockerham 1984) and can be used  
109 as a null expectation for the degree of population divergence due to genetic drift and gene flow. If  
110  $Q_{st} > F_{st}$ , the trait divergence is higher than the neutral expectation and is likely the result of  
111 directional selection favouring local adaptation (Leinonen *et al.* 2013) rather than the result of  
112 drift. The two blue tit populations have very large effective population sizes (roughly estimated  
113 around 10,000 in each valley, Perrier *et al.*, genomic work in progress) and have been found  
114 weakly genetically differentiated (Szulkin *et al.* 2016). Consequently, it is unlikely that any  
115 genetic divergence for these traits would be produced by genetic drift. We considered that,  
116 because birds were raised in a common garden, a phenotypic difference among individuals was a  
117 realistic approximation of an additive genetic effect. We thus used the phenotype of the common

118 garden birds to calculate the  $Q_{st}$  and predicted that the  $Q_{st}$  of each trait would significantly exceed  
119 the  $F_{st}$  estimated between both populations. In addition, in order to better understand the  
120 importance of environmental factors in shaping the observed phenotypic differentiation in the  
121 wild, we compared the genetic differentiation ( $Q_{st}$ ) of birds from the common garden experiment  
122 with the phenotypic differentiation of wild birds for the same traits ( $P_{st}$ ; the amount of  
123 phenotypic variance among wild populations relative to the total phenotypic variance in the trait).

124

## 125 **Materials and Methods**

126 The population located in the evergreen habitat (Evergreen-Pirio) is in the Corsican Fango valley  
127 (42°34'N, 08°44'E; 200m elevation) and contains 205 nest-boxes distributed across two study  
128 plots. The population located in the deciduous habitat (Deciduous-Muro) is in the Corsican  
129 Regino valley (42°32'N, 08°55'E, 350 m elevation) and contains 110 nest-boxes distributed  
130 across three study plots. A weekly to daily monitoring over the course of the breeding season  
131 (from early April to the end of June) allowed the recording of exact laying dates and hatching  
132 dates for all broods established in nest boxes.

133

134 Nestlings were collected for the common garden experiment at 7 to 12 days of age and were  
135 brought to the Netherlands Institute of Ecology (NIOO-KNAW, Wageningen, Netherlands)  
136 where they were hand raised under standardized conditions. We used 169 blue tits that were  
137 collected in 2010 and 2011 in the deciduous habitat (2010: 42 birds, 7 broods; 2011: 39 birds, 6  
138 broods) and in the evergreen habitat (2010: 44 birds, 10 broods; 2011: 44 birds, 8 broods). In  
139 2010, before collecting chicks, broods were cross-fostered between nests for another experiment.

140 For this experiment, at 2 to 4 days old, half of the chicks from a given brood were exchanged  
141 with half of the chicks of another brood from the same population.

142  
143 Once collected, all birds were transported by car and hand-fed from Corsica to the Netherlands,  
144 and were hand reared until independence as described in Reparaz *et al.* (2014). Briefly, all the  
145 chicks from a given habitat and year were kept in boxes divided into multiple compartments that  
146 were not isolated from one another, each compartment containing one nest of 3 to 5 nestlings,  
147 until fledgling. Chicks from adjacent nests could easily change compartment, meaning that chicks  
148 from different nests were quickly mixed. After fledgling, birds were housed in cages in groups of  
149 2 to 4 birds, irrespective of their sex and nest of origin (assigned randomly). Up to that period,  
150 chicks were fed every half-hour, 14 hours per day (7:00 am - 9:00 pm), with a diet consisting of a  
151 mixture of curd cheese, ground beef heart, baby cereal, multivitamin solution and calcium  
152 carbonate, supplemented with wax moth larvae and bee larvae, until independence. Raising  
153 chicks from the different habitats at exactly the same time would have been ideal but was  
154 impossible because chicks in the Regino and the Fango valleys hatch one month apart. However,  
155 chicks from different nests and habitats could easily see and hear each other, as they were raised  
156 in the same rooms, and fledglings from the Regino valley were still present in the cages when the  
157 younger chicks from the Fango valley arrived in the laboratory. Caretakers were the same for  
158 birds of different origins.

159  
160 At independence, about 35 days after hatching, birds were relocated to larger individual cages or  
161 aviaries. Food and water were provided *ad libitum*. In 2012 and 2015, birds were moved to the  
162 Centre d'Écologie Fonctionnelle et Évolutive (CEFE-CNRS; Montpellier, France), where they  
163 were kept in outdoor aviaries before being released back into their natal habitat in Corsica.

164 Morphological measurements were taken during the period at the NIOO-KNAW. Tarsus length  
165 was measured once (at > 1 year of age) but body mass was measured several times, always by the  
166 same person. We were interested in testing for a difference in adult body mass and thus kept in  
167 the analysis only the measures made at one year of age and older.

168

#### 169 Behavioural and physiological trials

170 In total, 169 birds were tested for their exploration behaviour and, among these birds, 137 were  
171 tested for handling aggression and 57 for heart rate. All behavioural and physiological traits were  
172 measured once for each bird, which prevented us from reporting their repeatability. However,  
173 these behavioural and physiological traits have been shown to be repeatable in these two  
174 populations in the wild, with repeatability estimates ranging from 0.26 to 0.75 depending on the  
175 trait (see Dubuc-Messier *et al.* 2017 for details). In the present study, exploration behaviour was  
176 measured using a different protocol (see below) than the one used in the wild (Dubuc-Messier *et*  
177 *al.* 2017). Nevertheless, we are confident that the exploration behaviour measured here represents  
178 repeatable characteristics of the individuals because this measure has been shown to be repeatable  
179 in blue tits in several studies using different protocols (Klueen & Brommer 2013; Mutzel *et al.*  
180 2013; Dubuc-Messier *et al.* 2017). For details regarding the phenotyping of wild birds used in the  
181  $P_{st}$  calculations, please refer to Dubuc-Messier *et al.* (2017).

182

#### 183 Exploration behaviour

184 Exploration behaviour trials were done in fall 2011 in the Netherlands Institute of Ecology as  
185 described by Reparaz *et al.* (2014) and using a novel environment chamber slightly modified  
186 from Drent *et al.* (2003). The novel environment chamber consisted of a 4.0 x 2.4 x 2.5m room  
187 with five artificial trees. Individuals were placed in cages adjacent to the main chamber 30 to 120

188 minutes before the trials and introduced in the main chamber through a sliding door. For two  
189 minutes, the observer counted the number of movements between trees and the number of small  
190 jumps on a given tree / branch. Exploration scores was the sum of both and varied from 10 (a  
191 very slow exploration pattern) to 92 (a very fast exploration pattern; Reparaz *et al.* 2014).

192

### 193 Handling aggression

194 Handling aggression was measured assessing the bird's aggression towards a manipulator  
195 (Dubuc-Messier *et al.* 2017). We used a score ranging from 0 to 3. A score of 0 was the lowest  
196 aggression score (no reaction; high docility) and 3 the highest (see Table S2 for detailed protocol).

197 Handling aggression was recorded in 2012 and 2015 at the CEFÉ-CNRS (France). Birds from the  
198 2010 cohort were tested for handling aggression in 2012 or 2015 (at 2 or 5 years of age), while  
199 the entire cohort from 2011 was tested for handling aggression in 2015 (at 4 years of age).

200 Handling aggression score was assessed blindly with respect to habitat of origin in 2015 and was  
201 assessed by two different observers, one in 2012 and one in 2015.

202

### 203 Heart rate during manual restraint

204 Heart rate was recorded in 2012 at the CEFÉ-CNRS (for the 2010 cohort only), as described by  
205 Dubuc-Messier *et al.* (2017). Within a few minutes after capture, we recorded heart rate for 30  
206 seconds using a digital recorder. We used the software Avisoft SASLab Pro version 5.1 to extract  
207 the mean time interval (sec) between two heartbeats using approximately 100 consecutive  
208 heartbeats per individual.

209

## 210 Molecular markers and $F_{st}$ calculation

211 For logistical reasons, we were not able to perform a molecular analysis on the birds used in the  
212 common garden experiment. As an alternative, we used a dataset, published by Szulkin *et al.*  
213 (2016) of wild birds from these two populations (i.e. deciduous,  $n = 49$ ; evergreen,  $n = 83$   
214 individuals) and genotyped at several thousand SNP using RAD-sequencing. We retained loci  
215 genotyped over at least 75% of the individuals. To avoid bias during filtering and in the  $F_{st}$   
216 estimates, we pruned highly related individuals from the dataset to keep only individuals linked  
217 with values of kinship lower than 0.05 (coefficient of Loiselle; Loiselle *et al.* 1995; Cheverud  
218 1996) computed in Genodive 2.27 (Meirmans & Van Tienderen 2004). In order to retain loci  
219 more likely to be informative, we applied a 5% MAF threshold (*Minor Allele Frequency*, using  
220 vcfTools 0.1.11; Danecek *et al.* 2011). We pruned the dataset for SNPs that deviated from Hardy-  
221 Weinberg-Equilibrium in at least one of the two populations ( $p$ -value  $< 0.05$ ) using vcfTools  
222 0.1.11. We retained only the first SNP of each 100 bp locus. To obtain a set of SNPs more likely  
223 to be neutral, we filtered out SNPs potentially under divergent selection between the two habitats  
224 ( $p$ -value  $< 0.015$ ; 0.7 % of total SNPs removed). This was done with a Bayescan 2.0 test (Foll &  
225 Gaggiotti 2008; 5 000 pilot iterations, 50 000 burnin, prior odds of 100). Average  $F_{st}$  and 95%  
226 confidence intervals were estimated using the R-package hierfstat 0.04-22 (Goudet 2005). The  
227 final dataset contained 69 individuals (32 and 37 individuals in the deciduous and evergreen  
228 habitats, respectively) genotyped at 5407 SNPs.

229

## 230 Statistical analysis

### 231 Genetic divergence between habitats of origin

232 We tested for a genetic difference between the two habitats for each trait with univariate linear  
233 mixed-models using the phenotype of each bird as a response variable and habitat of origin, sex,

234 and their interaction as fixed effects. When we found a significant interaction between habitat of  
235 origin and sex, we ran a separate model for each sex. Specific confounding variables were added  
236 as fixed effects for each particular trait. For exploration score, we added a cohort term as fixed  
237 effect to test for any environmental effect early in life or during the hand-rearing period in  
238 captivity. Novel environment tests were done on the two cohorts at the same time (in autumn  
239 2011). Thus, at the time of the test, individuals born in 2010 were almost 1½ years old, while  
240 individuals born in 2011 were 5 months old. Hence, in this model, the cohort term controlled for  
241 the combined effect of cohort and age. For handling aggression score, we added cohort, time of  
242 day (hour), and year of test (2012 or 2015) as fixed effects. For heart rate models, we added as  
243 fixed effect mean individual adult body mass because heart rate is related to the metabolic rate  
244 and both are positively related to body mass (Green *et al.* 2011). Heart rate recordings were done  
245 in 2012 on the 2010 cohort only. We therefore did not add a fixed effect for bird age, cohort or  
246 year to avoid redundancy. We also added in heart rate models the time of day (hour) as a fixed  
247 effect. For body mass, we added age as a continuous variable, cohort, and time of day (hour). For  
248 tarsus length, we added cohort only as fixed effect (*i.e.* 2010 and 2011).

249  
250 In all models, we used random intercepts for the brood of origin and rearing brood to account for  
251 the non-independence of birds coming from the same brood or / and the effect of foster parents  
252 for nestlings that have been cross-fostered prior to the captivity period. Because body mass was  
253 measured several times for each bird, we also added a random intercept for bird identity for this  
254 trait.

255  
256 All response variables were Z-transformed prior to analyses. We tested the significance of the  
257 fixed effects and selected a minimal models by LRT (log likelihood ratio test) in a stepwise

258 elimination procedure starting with a model that included all variables (Bates *et al.* 2014). We  
259 kept all the random effects in final minimal models. We present in Table S3 the L-ratios and p-  
260 values associated with all variables in initial models. Analyses were done with *R* (R Core Team  
261 2017) using the function *lmer* of the package *lme4* (Bates *et al.* 2015). Confidence intervals  
262 (95%) were generated with the function *confint.merMod* (*lme4*). We assumed a Gaussian  
263 distribution for all traits, which was confirmed after visual inspection of the residuals. We also  
264 evaluated the population of origin effect across all five traits using Fisher's combined probability  
265 test run with the *sumlog* function of the R package *metap* (Dewey 2017).

266  
267  $Q_{st}$ ,  $P_{st}$  and  $F_{st}$  comparison  
268 Because birds were raised in a common garden, we considered that a phenotypic difference  
269 among individuals was a realistic approximation of an additive genetic effect. For each trait, we  
270 thus calculated the  $Q_{st}$  between the two habitats based on the phenotypes of birds from the  
271 common garden using a procedure similar to Bertrand *et al.* (2016) with univariate mixed models  
272 in a Bayesian framework. We calculated  $Q_{st}$  as:

273  
274 
$$Q_{st} = \sigma_B / (\sigma_B + 2 * \sigma_W)$$

275  
276 Where  $\sigma_B$  is the between-habitat phenotypic variance and  $\sigma_W$  the within-habitat variance (or  
277 residual; Wright 1949). The two variance components were extracted from a univariate linear  
278 mixed model including habitat of origin (and identity of the bird for body mass) as random  
279 intercepts. We also included the fixed effects structure selected previously (minimal model)  
280 excluding the term habitat of origin. We calculated  $\sigma_B$  as the variance attributable to the habitat

281 of origin and  $\sigma_W$  as the residual variance (or for body mass as the sum of the variance attributable  
282 to the residual and to the individual identity; Bertrand *et al.* 2016). We did not include any broods  
283 effects in these models because the variance attributable to the brood is also attributable to the  
284 population of origin. We present the between-habitat variance for each study trait extracted from  
285 the models used to calculate  $Q_{st}$  in Table S5. We calculated  $P_{st}$  as  $Q_{st}$  but used as random  
286 intercepts habitat of origin, the identity of the bird and the observer identity (for handling  
287 aggression and heart rate) along with the significant fixed effects detailed in Dubuc-Messier *et al.*  
288 (2017). For  $P_{st}$  calculation, we calculated  $\sigma_B$  as the variance attributable to habitat of origin and  
289  $\sigma_W$  as the sum of the variance attributable to the observer, to the residual variance and the  
290 individual identity.

291  
292 These models were performed with *MCMCglmm* package (Hadfield 2010) in R using slightly  
293 informative priors (*i.e.*  $V = V_P / n$ ,  $nu = 1$  or  $0.5$ ;  $V_P$  is the total phenotypic variance of the trait  
294 and  $n$  the number of random effects), 10 million iterations, a thinning of 200 and a burn-in phase  
295 of 500. Because the results of the models with different  $nu$  were similar, we used the posterior  
296 distribution of models with  $nu = 1$  in  $Q_{st}$  and  $P_{st}$  calculations. We assessed the presence of  
297 autocorrelation with the function *autocorr* (*MCMCglmm* package). All models showed an  
298 autocorrelation less than  $10^{-4}$ . We checked for model convergence with the function *gewe.diag* of  
299 the *coda* package (Plummer *et al.* 2006). For all traits, we calculated the ratio  $Q_{st} / \text{mean } F_{st}$  for  
300 each sample of the posterior distribution and report the posterior mode of the ratio and its 95%  
301 credibility intervals (calculated using the *HPDinterval* function of the package *lme4*). We  
302 assumed that  $Q_{st}$  differed significantly from  $F_{st}$  when the credibility interval around the ratio did  
303 not include one.

304

305 **Results**

306 Divergence between habitats of origin

307 The Fisher combined probability test method on all studied traits indicated an overall significant  
308 effect of the habitat of origin (chi-squared : 54.647, df=10 and p-value < 0.001). Below we  
309 present the results for each trait separately.

310

311 Behavioural and physiological traits

312 For birds in the common garden experiment, habitat of origin had a significant effect on the two  
313 behavioural traits: blue tits from the deciduous habitat were faster explorers and were more  
314 aggressive to the handler (Table 1; Fig. 1). Birds from the deciduous habitat had a lower heart  
315 rate than birds from the evergreen habitat (Table 1; Fig. 1). We found a trend for an interaction  
316 between habitat of origin and sex for heart rate (L-ratio = 3.360, d.f. = 1, p-value = 0.067):  
317 evergreen males had a higher heart rate than deciduous males [estimate = 1.24 (95% CI: 0.31;  
318 2.17), L-ratio = 6.260, d.f. = 1, p-value = 0.010] but there was no habitat of origin effect for  
319 females (L-ratio = 2.150, d.f. = 1, p-value = 0.142). There was no interaction between sex and  
320 habitat of origin for the two behavioural traits, but there was a difference in exploration score  
321 between sexes (Table 1).

322

323 Morphological traits

324 Habitat of origin also had a significant effect on the two morphological traits: deciduous birds  
325 were heavier and had a longer tarsus than evergreen birds (Table 1; Fig. 1). We did not find any  
326 interaction between habitat of origin and sex for these two traits (tarsus length: L-ratio = 0.226,

327 d.f. = 1, p-value = 0.634; body mass: L-ratio = 0.155, d.f. = 1, p-value = 0.694). Among-  
328 individual differences in body mass were significant and represented 45% of the total variance of  
329 the trait [variance = 0.34 (95% CI: 0.26; 0.46), L-ratio = 421.95, p-value < 0.001].

330

331 Brood effects

332 Differences among broods of origin explained a significant portion (78%) of the total phenotypic  
333 variance in body mass, but not for the other traits (Table S4). Differences among rearing broods  
334 explained a significant portion of the total variance in tarsus length (22%) and a marginally  
335 significant portion of total variance in heart rate (30%, p-value = 0.07) but not for the other traits.

336

337  $Q_{st}$ ,  $P_{st}$  and  $F_{st}$  comparison

338 We found a significant but small genetic differentiation between the two populations [mean  $F_{st}$   
339 over all loci = 0.004 (95% CrI: 0.003; 0.005), p-value < 0.001].  $Q_{st}$  was higher than  $F_{st}$  with non-  
340 overlapping intervals for all traits. The ratio between the  $Q_{st}$  and  $F_{st}$  was significantly greater than  
341 one for all traits. Credibility intervals for  $Q_{st}$  and  $P_{st}$  overlapped for all traits (Table 2).

342

## 343 Discussion

344 Our common garden experiment suggests a genetic divergence in personality, physiological and  
345 morphological traits between two blue tit populations inhabiting contrasted habitats separated by  
346 a small spatial distance in regards to the species dispersal capacity. Adult blue tits originating  
347 from the evergreen habitat displayed slower exploration behaviour, lower handling aggression  
348 (higher docility), faster heart rate, lower body mass and shorter tarsus compared to birds from the  
349 deciduous habitat (Table 1; Fig. 1). These differences are similar to the ones measured in the wild

350 suggesting that plasticity alone is not responsible for the observed phenotypic divergence in the  
351 wild (Charmantier *et al.* 2016; Dubuc-Messier *et al.* 2017). In addition, we found a significant  $F_{st}$   
352 between the two populations, but its low value (0.004) indicates current or past gene flow, in  
353 concordance with previous findings (Szulkin *et al.* 2016). The  $Q_{st}$  -  $F_{st}$  comparisons revealed that  
354 blue tits from these populations are more differentiated for personality, physiological and  
355 morphological traits than they are at the genome-wide level (Table 2). These results suggest that  
356 genetic drift alone does not explain the observed divergence between the two populations and  
357 that differences in selection regimes are responsible for this divergence.

358  
359 The divergence we describe in personality, physiological and morphological traits is likely to be  
360 mainly of genetic origin, since birds from both habitats were raised in identical conditions from  
361 their first week of life to up to five years. In addition, the divergence found in this study for adult  
362 body size is consistent with previous studies that have found divergent selection between the two  
363 populations for morphological traits (Blondel *et al.* 2002; Teplitsky *et al.* 2014) and moderate to  
364 high heritability for these traits (0.29 to 0.51; Teplitsky *et al.* 2014). However, we cannot  
365 completely exclude that early environmental effects such as non genetic inheritance, occurring  
366 before the chicks were sampled from their nest were at least partly responsible for the observed  
367 patterns (Kruuk & Hadfield 2007; Räsänen *et al.* 2007; Bonduriansky & Day 2009; Bouwhuis *et*  
368 *al.* 2010; van Oers *et al.* 2015). Such early environmental effects might be particularly important  
369 for tarsus length, which is usually fixed at fifteen days of age for this species. However, for  
370 behavioural traits, such strong environmental effects lasting for up to five years are unlikely,  
371 since very few studies have reported long-term consequences of early environmental conditions  
372 for the studied traits (Taylor *et al.* 2012; Petelle *et al.* 2015) and because maternal effects are  
373 known to decrease during ontogeny (Cheverud *et al.* 1983; Wilson *et al.* 2007). One way to

374 control for very early environmental effects would be to allow the birds to breed in captivity and  
375 compare the phenotypes in the offspring generation. However, this type of experiment presents  
376 significant challenges that have so far prevented their feasibility in our study system. In  
377 particular, while it is possible to maintain blue tits in aviaries for short time experiments (Reparaz  
378 *et al.* 2014) it is difficult to make them breed in captivity (Lambrechts *et al.* 1999).

379  
380 Some studies have raised concerns regarding  $Q_{st}$  and  $F_{st}$  estimation and their comparison  
381 (Leinonen *et al.* 2013). In particular between-population variance and thus  $Q_{st}$  estimation may be  
382 imprecise when a small number of populations are compared like it is the case in our study  
383 (O'Hara and M€erila 2005; Leinonen *et al.* 2013). However, simulations have shown that a small  
384 number of populations results in a downward bias in  $Q_{st}$  estimation when  $Q_{st}$  is high (O'Hara and  
385 M€erila 2005). Another important concern is whether genetic markers involved in  $F_{st}$  estimation  
386 are truly neutral (Leinonen *et al.* 2013). In this study, we used an  $F_{st}$  calculated from markers that  
387 included the whole genome. Although we filtered SNPs under potential divergent selection, it is  
388 possible that we included potentially non-neutral regions (or that we removed some neutral ones).  
389 However, using microsatellites, Porlier *et al.* (2012) have found a lower  $F_{st}$  (0.001) between the  
390 same populations during a similar time period (year 2009). Hence, although  $Q_{st}$  and  $F_{st}$   
391 comparison have some limitations, these limitations should most probably have limited our  
392 capacity to detect significant  $Q_{st}$  -  $F_{st}$  differences rather than reveal false differences.

393  
394 Environmental heterogeneity, divergent selection and local adaptations  
395 The importance of environmental heterogeneity and gene flow for phenotypic divergence has  
396 mainly been studied for life history and morphological traits and much less for behavioural traits.  
397 Indeed, few studies have disentangled so far the role of plasticity from that of genetic differences

398 in shaping the phenotypic divergence of populations for behavioural traits (Bell 2005;  
399 Dingemanse *et al.* 2007; Herczeg *et al.* 2013; Karlsson *et al.* 2016; Jacquin *et al.* 2016) and even  
400 fewer for highly mobile avian species (Atwell *et al.* 2012; Miranda *et al.* 2013). In addition, to  
401 our knowledge, no studies has until now reported  $Q_{st}$  -  $F_{st}$  comparisons involving personality  
402 traits. This shortage of study is probably due to the fact that personality traits are often considered  
403 plastic and thus less prone to genetic divergence and local adaptations than morphological traits.  
404 Yet, the results of our study suggest a genetic divergence for personality traits and that this  
405 divergence could be as strong as for morphological traits (Table 1 and 2).

406  
407 Past studies in this system and on personality variation suggest that the genetic divergence found  
408 here could be the result of the coevolution of multiple types of traits in response to the ecological  
409 context of each habitat. Indeed, an increasing number of studies are suggesting that life-history  
410 and personality traits could have co-evolved to form a pace-of-life syndrome (Réale *et al.* 2010).  
411 For example, empirical and theoretical studies are suggesting that high investment in early  
412 reproduction at a cost of reduced residual reproductive value (either via survival or future  
413 reproduction) should be associated with boldness, fast exploration, and high aggressiveness  
414 (Réale *et al.* 2010; Wolf *et al.* 2007). Our results on this system are consistent with the pace-of-  
415 life syndrome hypothesis. Blue tits from the deciduous habitat, which are more aggressive and  
416 faster explorers, have a shorter lifespan and a lower residual reproductive value, but larger clutch  
417 sizes than birds from the evergreen habitat (Grosbois *et al.* 2006; Charmantier *et al.* 2016;  
418 Dubuc-Messier *et al.* 2017; Table S1). Our results suggest that these divergences for personality  
419 traits are genetic and the  $Q_{st}$  -  $F_{st}$  comparisons revealed that they are likely the result of divergent  
420 selection pressures rather than drift. In addition, studies on other blue tit or great tit (*Parus major*)  
421 populations have found that the personality phenotype is heritable and related to fitness (van Oers

422 & Sinn 2011; Class *et al.* 2014). Therefore, taken together, our results suggest that the personality  
423 phenotypes of birds living in these habitats could have evolved and be implicated in blue tit  
424 adaptation to local ecological conditions prevailing in each habitat.

425

426 Brood effects

427 We did not find any significant brood-of-origin effect for handling aggression, exploration score,  
428 heart rate, and tarsus length. Since all these traits except heart rate have been shown to be  
429 heritable in previous studies on blue tits (van Oers & Sinn 2011; Class *et al.* 2014; Teplitsky *et al.*  
430 2014), the absence of heritable variance in our analysis is most probably explained by the  
431 relatively small number of broods. Estimating heritability was not the goal of this study, we only  
432 wanted to control for dependence issues associated with the use of sibs.

433

434 The partial cross-fostering manipulation before the common garden experiment revealed a  
435 significant rearing brood effect for tarsus length. This result suggests that the rearing environment  
436 between 2 days to 12 days old can have a significant impact on this morphological trait.  
437 Contrarily to the other traits that are more labile, tarsus length generally stabilises at fifteen days  
438 of age in blue tits. We were, therefore, able to capture the early environmental effect for this trait  
439 by measuring the adult phenotype. We found a marginally significant brood of rearing effect for  
440 heart rate but not for other traits. There may be several reasons for such results. First, these traits  
441 may not be sensitive to the rearing environment. Second, it is possible that - as for brood of origin  
442 - these traits are slightly sensitive to early environmental effect (2 to 12 days) but that we lack  
443 power to detect it.

444  
445 Genetic and environmental effects are not mutually exclusive  
446 Genetic divergence does not preclude a plastic response to ecological conditions specific to each  
447 habitat. For example, in the wild, the phenotypic difference in male heart rate between habitats  
448 was not significant (Dubuc-Messier *et al.* 2017), but using the common garden experiment we  
449 found here a significant difference in male heart rate. It is thus possible that plastic responses of  
450 heart rate to habitat specific ecological conditions in the wild may have hidden the genetic  
451 divergence (Conover & Schultz 1995). In addition, the important temporal variation in mean  
452 handling aggression in the wild shown by Dubuc-Messier *et al.* (2017) in each population,  
453 suggests that individuals can partly adjust their personality phenotype for this trait depending on  
454 the current local conditions. However, for all traits, the  $P_{st}$  between wild birds was not  
455 statistically different from their  $Q_{st}$ , suggesting that environmental effects in the wild might not  
456 result in stronger or weaker differentiation compared to the genetic differentiation.

457  
458 Conclusion  
459 Our study suggests a genetic divergence for personality, physiological and morphological traits  
460 between two blue tit populations that occupy different habitats but that are separated by small  
461 spatial distances compared to the dispersal ability of the species and connected by gene flow. The  
462 present study and past results for this system suggest that these differences are likely due to  
463 different selection pressures and may represent local adaptations. These results thus emphasize  
464 the role of environmental heterogeneity for intra-specific phenotypic diversity and suggest that  
465 genetic population divergence is possible at small spatial scales (relative to their dispersal ability)  
466 for behavioural traits.

467

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471

## 472 **Data accessibility**

473 The dataset will be shared on dryad upon publication.

474

## 475 **References**

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692 Figure 1. Mean phenotypes of blue tits originating from two distinct populations and  
693 habitats (deciduous and evergreen) in Corsica (France) and reared in a common garden. A)  
694 exploration score, B) handling aggression score, C) heart rate during manual restraint (heart  
695 beats/min.), D) tarsus length (mm) and E) adult body mass (g). Boxplots on raw data, the  
696 boxes represent the first and the third quartile, the lines represent the median, the ends of the  
697 whiskers represent the minimum data in the  $1.5 * \text{the interquartile range}$ , dots represent  
698 extreme data. All differences are significant (see Table 1 for details).

699

700 Table 1. Final models describing the phenotype of blue tits originating from two distinct  
 701 populations and habitats (deciduous and evergreen) in Corsica (France) and reared in a common  
 702 garden.

Traits	Terms	Estimates	95% CI	L-ratio	d.f.	p-value
Exploration score	Intercept	-0.32	-0.62; -0.03			
	Habitat of origin	-0.48	-0.78; -0.19	9.70	1	0.002
	Sex	0.26	0.004; 0.52	3.97	1	0.046
	Cohort	0.88	0.59; 1.17	23.91	1	< 0.001
Handling aggression	Intercept	0.45	0.18; 0.72			
	Habitat of origin	-0.82	-1.18; -0.46	14.96	1	< 0.001
Heart rate during restraint (HR)	Intercept	-0.57	-1.06; -0.09			
	Habitat of origin	0.98	0.35; 1.62	8.17	1	0.004
Body mass	Intercept	-1.07	-1.40; -0.74			
	Habitat of origin	-0.33	-0.63; -0.03	4.46	1	0.034
	Sex	-0.56	-0.77; -0.35	25.08	1	< 0.001
	Age	0.27	0.21; 0.33	74.23	1	< 0.001
	Time of day	0.09	0.07; 0.11	75.50	1	< 0.001
Tarsus length	Intercept	-0.25	-0.58; 0.08			
	Habitat of origin	-0.60	-1.00; -0.19	7.74	1	0.005
	Sex	1.04	0.81; 1.28	61.46	1	< 0.001

703 The deciduous habitat, females, and cohort 2010 were set as references in models. Estimates are  
 704 from a model with the brood of rearing and brood of origin in random effect (and individuals  
 705 identity for body mass), variance estimates are shown in Table S4. L-ratio and p-values are from  
 706 the comparison of a full model and a model without the variable of interest. The p-values and L-  
 707 ratio associated with each parameter in initial models before selection are presented in Table S3.  
 708

709 Table 2.  $Q_{st}$  and  $P_{st}$  values (posterior mode) for each trait (and 95% credible interval (CrI)), mean  
 710  $F_{st}$  and  $Q_{st} / F_{st}$  ratio [posterior mode and associated 95% CrI] between two blue tits populations  
 711 originating from distinct populations and habitats (deciduous or evergreen) in Corsica (France) and  
 712 reared in a common garden.

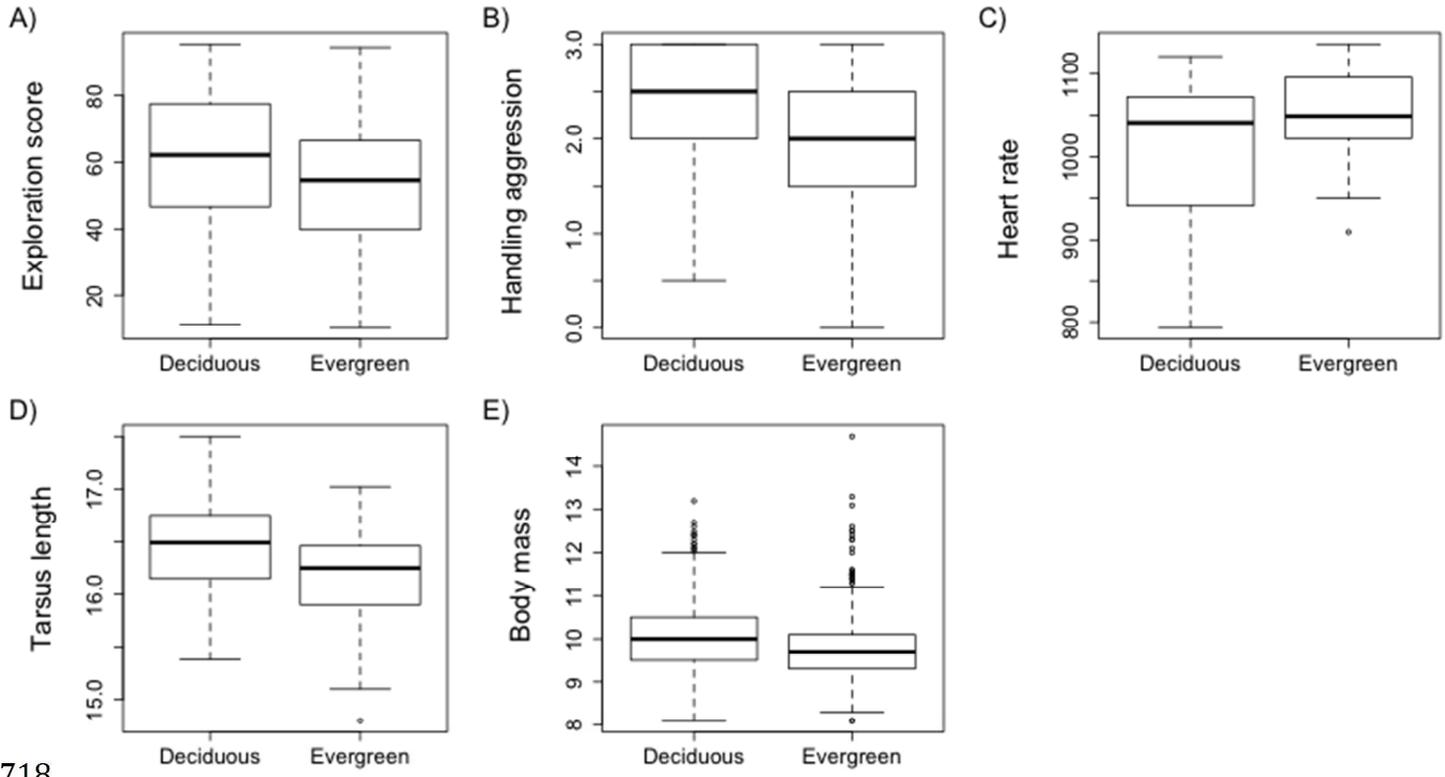
Traits	$Q_{st}$ (95% CrI)	$P_{st}$ (95% CrI)	$Q_{st} / F_{st}$ ratio (95% CrI)
Exploration score	0.084 (0.029; 0.804)	0.063 (0.018; 0.727)	20.982 (7.266; 201.065)
Handling aggression	0.129 (0.034; 0.832)	0.045 (0.011; 0.692)	32.309 (8.525; 208.025)
Heart rate during manual restraint (HR)	0.101 (0.033; 0.846)	0.032 (0.007; 0.562)	25.320 (8.244; 211.475)
Body mass	0.069 (0.018; 0.736)	0.095 (0.030; 0.773)	17.144 (4.541; 183.998)
Tarsus length	0.197 (0.050; 0.872)	0.212 (0.048; 0.864)	49.368 (12.455; 217.881)
<i>Mean <math>F_{st}</math></i>	<i>0.004</i> <i>(0.003; 0.005)</i>		

713  $Q_{st}$  have been calculated from the phenotype of birds raised in a common garden and  $P_{st}$  from the  
 714 phenotype of wild birds.

715

716

717 Figure 1.



718

719

720 **Online supporting information**

721 Table S1. Caterpillar abundance, life-history, morphological and personality phenotypes (mean  
722 (*n*)) of the two Corsican blue tit populations (France) in the wild.

723 Table S2. Blue tit handling aggression scale.

724 Table S3. L-Ratio, degree of freedom and p-values associated with each parameter in initial  
725 models describing the phenotype of blue tits originating from two distinct habitats (deciduous  
726 and evergreen) in Corsica (France) and reared in a common garden.

727 Table S4. Variance components, L-ratio and p-values for studied traits in two blue tits  
728 populations in Corsica (France) reared in a common garden.

729 Table S5. Between-habitat variance (posterior mean and 95% CrI) for each study trait extracted  
730 from the models used to calculate  $Q_{st}$ .

731

732

## Online Supporting Information

### Average caterpillar abundance, life-history, morphological and personality phenotypes

Table S1. Caterpillar abundance, life-history, morphological and personality phenotypes (mean (*n*)) of the two Corsican blue tit populations (France) in the wild.

Populations	Deciduous	Evergreen
First year of monitoring	1993	1976
Caterpillar abundance <sup>1</sup>	762.87	87.10
Annual adult survival probability <sup>2</sup>	0.391 (6)	0.574 (14)
Date of first egg laying (1 = March 1 <sup>st</sup> ) <sup>3</sup>	38.56 (1233)	70.08 (1920)
Male body mass (g) <sup>3</sup>	9.82 (1032)	9.37 (1607)
Female body mass (g) <sup>3</sup>	9.66 (1153)	9.23 (1616)
Male tarsus length (mm) <sup>3</sup>	16.52 (578)	16.27 (789)
Female tarsus length (mm) <sup>3</sup>	16.05 (614)	15.84 (798)
Clutch size <sup>3</sup>	8.50 (1235)	6.61 (1913)
Number of fledglings <sup>3</sup>	6.60 (1092)	4.15 (1273)
Mean exploration speed (cm/s) ± s.d. <sup>4</sup>	13.52 ± 8.39 (176)	10.37 ± 7.49 (117)
Mean handling aggression score ± s.d. <sup>4</sup>	1.69 ± 0.95 (703)	1.49 ± 0.99 (549)
Mean heart rate during manual restraint ± s.d. <sup>4</sup>	963.30 ± 87.80 (159)	976.24 ± 86.99 (91)

<sup>1</sup> mean maximal frass mg/m<sup>2</sup> per day in each population (sampled between 2011 and 2015 during the breeding period using 0.25m<sup>2</sup> trays placed under the forest canopy and collected twice a week, see Zandt et al. 1990 for details about the sampling procedure); <sup>2</sup> Dubuc-Messier et al. In prep; <sup>3</sup> Charmantier et al. 2016 (collected between the first year of monitoring and 2014); <sup>4</sup> Dubuc-Messier et al. 2016).

### Handling aggression scores

The test was done within two minutes after capture and prior to any other manipulation. The handler held the bird with one hand and placed the bird's legs between his forefinger and his thumb to let the bird free to move its tails and wings. The handler pointed the forefinger of his other hand at a spot about 2 to 3 cm in front of the bird's beak and noted if the bird struck at his finger, and the position of its wings and tail. After two seconds in this position, the handler moved his forefinger towards the bird's beak two or three times and recorded its reaction.

Table S2. Blue tit handling aggression scale.

Score	Wings spread	Tail feathers spread	Bird strikes fingers
0	No	No	No
1	No	No	Yes, but only if provoked
2	No	Yes	Yes, spontaneously
3	Yes	Yes	Yes, spontaneously

When the bird displayed one reaction specific to one score and another reaction specific to another score, it received an average score between the two. For example, a bird that struck without any provocation (score 2) but did not have its wings and tail feathers spread (score 1) would be scored as 1.5.

Initial models

Table S3. L-Ratio, degree of freedom and p-values associated with each parameter in initial models describing the phenotype of blue tits originating from two distinct habitats (deciduous and evergreen) in Corsica (France) and reared in a common garden.

Traits	Terms	L-ratio	d.f.	p-value
Exploration score	Cohort	23.912	1	< 0.001
	Sex * Habitat of origin	1.104	1	0.293
	Sex	3.970	1	0.046
	Habitat of origin	9.697	1	0.002
Handling aggression	Time of day	0.258	1	0.612
	Cohort	0.052	1	0.819
	Year of trial	0.001	1	0.973
	Sex * Habitat of origin	0.615	1	0.432
	Sex	0.092	1	0.761
	Habitat of origin	20.592	1	< 0.001
Heart rate during restraint	Mean body mass	0.256	1	0.873
	Sex * Habitat of origin	3.3601	1	0.066
	Sex	1.9081	1	0.167
	Habitat of origin	9.012	1	0.003
	Time of day	0.449	1	0.502
Body mass	Time of day	75.500	1	< 0.001
	Age	74.230	1	< 0.001
	Cohort	0.014	1	0.905
	Sex * Habitat of origin	0.155	1	0.694
	Sex	25.080	1	< 0.001
	Habitat of origin	4.460	1	0.034
Tarsus length	Cohort	0.350	1	0.554
	Sex * Habitat of origin	0.226	1	0.634
	Sex	25.070	1	< 0.001
	Habitat of origin	4.457	1	0.034

The brood of rearing and brood of origin identity are fitted as random effect in all models (and individuals identity for body mass), variance estimates are shown in Table S3. L-ratio and p-values are from the comparison of a full model and a model without the variable of interest.

Variance components

Table S4. Variance components (brood of origin, brood of rearing, and residuals), L-ratio, and p-values for studied traits in two blue tits populations in Corsica (France) reared in a common garden.

Traits	<i>Brood of origin</i>				<i>Rearing broods</i>				<i>Residuals</i>
	Variance (95% CI)	L-ratio	d.f.	p-value	Variance (95% CI)	L-ratio	d.f.	p-value	Variance (95% CI)
Exploration score	0.05 (0.00; 0.15)	0.76	1	0.38	0 (0.00; 0.001)	0	1	1	0.71 (0.56; 0.89)
Handling aggression	0.01 (0.00; 0.13)	0.002	1	0.97	0.08 (0.00; 0.24)	1.52	1	0.22	0.70 (0.54; 0.94)
HR	0.05 (0.00; 0.30)	0	1	1.00	0.25 (0.00; 0.70)	3.35	1	0.07	0.66 (0.42; 1.02)
Tarsus length	0.09 (0.00; 0.29)	1.69	1	0.19	<b>0.14</b> <b>(0.01; 0.37)</b>	4.74	1	0.03	0.44 (0.34; 0.57)
Body mass	<b>0.07</b> <b>(0.01; 0.14)</b>	4.10	1	0.04	0 (0.00; 0.001)	0	1	1.00	0.02 (0.22; 0.26)

L-ratio and p-values are from the comparison of a full model and a model without the variable of interest. Bold indicates significant variance components.

Table S5. Between-habitat variance (posterior mean and 95% CrI) for each study trait extracted from the models used to calculate  $Q_{st}$ .

<b>Traits</b>	<b>Between habitat variance (95% CrI)</b>
Exploration score	1700 (11.88; 2464)
Handling aggression	2.909 (0.021; 4.470)
Heart rate during manual restraint (HR)	34 971 (245.5; 57 937)
Body mass	1.564 (0.015; 3.041)
Tarsus length	0.142 (0.114; 0.181)

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