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Rare earth elements in French stream waters – revisiting the geochemical continental cycle using FOREGS dataset

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Abstract

The geochemical behavior of rare earth elements (REE) has been investigated mainly in geological systems where these elements represent the best proxies of processes involving the occurrence of an interface between different media. This behavior is assessed according to REE concentrations recorded along the REE series normalized with respect to upper continental crust. In this study based on a field approach, the geochemical behavior of REE was investigated in French stream waters. This study is based on FOREGS (Forum of European Geological Surveys) Geochemical dataset that consists on a sampling at regular mesh on all Europe. In France, 119 stream water samples were extracted in drainage basins $<100 \text{ km}^2$. The aim of the study is that of describing the spatial variation of REE and finding the hydro-topo-geochemical factors that affect their distribution by means of a Multivariate Factorial Kriging.

On the basis of their atomic number and of the results of a preliminary Principal Component Analysis three REE have been selected (La, Eu and Lu) and five physicochemical properties (pH, organic carbon, carbonates, Fe, Mn). A cokriging has been applied that shows a similar spatial organization of REE: higher values are especially observed in the Aquitaine basin. In order to investigate more deeply on the different sources of variation acting in the study area, a factorial cokriging is applied. The first 2 regionalized factors have been estimated to give a synthetic description of the studied process at the different selected spatial scales. At higher spatial scales (250 km) environmental parameters like Fe, carbonates, pH, supposed to be ascribed to the rock's nature or to other geological larger scale processes (i.e., hydrographic network and topography), have shown to affect REE distribution. At short range, only Eu and Mn weigh more, which are ascribed to the process of liberation of Mn oxides in rivers that also release the REE sorbed onto these oxides.

Keywords lanthanide, river water, organic matter, Fe and Mn oxides

1. Introduction

Rare earth elements (REE) represent a group of fifteen elements, which share common physiochemical properties and therefore often occur together (McLennan and Taylor, 2012). Over the past fifteen years, REE became of critical importance to many green-technology products but also for medical applications, and therefore are of great economic interest (e.g., (Tepe et al., 2014; Guyonnet et al., 2015). In this context as highlighted by Kulaksız and Bau (2013), the continuous development of new technologies and new substances has led to strongly increased release of REE into natural waters, although their toxicological effects and the potential implications for the ecosystem are often not fully understood. But in aquatic systems, REE concentrations are low compared to their concentrations in rocks with regards to their slight solubility (e.g., Noack et al., 2014). Therefore, it appears important first, to assess and to fully understand the occurrence and fate of aqueous REE in an environment where REE release from the anthroposphere may be considered as negligible.

In aquatic systems, solution and interface chemistry appears to be the major factor controlling the REE concentration (e.g., Elderfield et al., 1990; Sholkovitz, 1995). Rare earth elements can form strong complexes with a number of different ligands. For convenience, REE concentrations of continental waters are usually normalized to Upper Continental Crust (UCC; McLennan, 2001), which produces smooth REE_{UCC} distribution patterns. The REE patterns result from the combination of several processes able to induce their fractionation. These processes are themselves controlled by several physicochemical mechanisms and parameters. Three processes can be distinguished: (i) precipitation/dissolution, (ii) sorption onto colloids and particles, and (iii) complexation in solution with organic and inorganic ligands. The resulting REE pattern therefore corresponds to the REE pattern for mineral sources that are modified by the sorption/complexation with ligands, colloids and particles. This results in a wide range of diverse REE patterns, which can be characterized by a

depletion or enrichment degree of light REE (LREE) relative to heavy REE (HREE) or by whether or not anomalies occur. Indeed, individual REE may show anomalous behavior in natural waters: redox-sensitive Ce and Eu may show anomalies that may be used as redox- and/or temperature-proxies (e.g., Bau, 1991; De Baar et al., 1988), and La, Gd and Lu may show small anomalies due to subtle differences between the stabilities of REE complexes (Bau, 1999; Byrne and Kim, 1990). In addition to these natural anomalies, anthropogenic anomalies of Gd and, recently, La and Sm have been reported from natural waters (e.g., Kulaksız and Bau, 2013), and references therein).

To better understand the occurrence of aqueous metals and especially REE, the FOREGS (Forum of European Geological Surveys) aims at providing a basis for formulating policies and legislation concerning the management of harmful elements and to define their corresponding safety levels (Salminen et al., 2005). A first attempt was performed in Italy, Sweden and Europe by Imrie et al. (2008); Lado et al. (2008); Petrosino et al. (2013); Sadeghi et al. (2013). These studies mostly focus on topsoil dataset and investigate the main factors explaining REE variation in the soil solution; however, the FOREGS database contains descriptions of other environmental media (stream and floodplain sediments, and stream water). The database has been processed by Imrie et al. (2008); Lado et al. (2008), Petrosino et al. (2013) and Sadeghi et al. (2013) by means of multivariate approaches to interpolate all variables, metals or REE concentrations. They used factorial kriging for Imrie et al. (2008), regression kriging for Lado et al. (2008) and inverse distance weighted interpolation and principal component analysis for Sadeghi et al. (2013). The results associate spatial distribution of elements with different factors operating at several scales. For example, Imrie et al (2008) highlight four factors and amongst them (i) a short scale (72 km): concentrations may be explained by parent material geology, land use and organic matter content; (ii) a

medium scale (296 km): concentrations may be explained by major structural division of European continent and the distribution of calcareous rocks. Petrosino et al. (2013) show that in Italy and Sweden, REE concentration in all sampling medias are related to the geological context. In the particular case of stream water, they found that Swedish waters are more concentrated in REE than Italian waters. These authors relate high REE concentrations to water acidity (linked to vegetation and felsic rocks). Whereas, in Italy, pH is higher and the watershed are mostly composed of calcareous rocks which mostly explain the lower REE contents as already highlighted by Johannesson and Burdige (2007) or Deberdt et al. (2002).

As the chemistry of stream waters is influenced by several landscape factors which are related to geology, topography, climate, and vegetation (e.g., Andersson and Nyberg, 2009; Gaillardet et al., 2014); the aim of this study is to further investigate the FOREGS dataset focusing on France. A special attention will be given to REE patterns spatial distribution. For such needs, the FOREGS stream dataset was processed by means of geostatistical methods, especially by factor kriging analysis. These approaches have three steps described as follows: (i) modeling the coregionalization of the set of variables, (ii) analyzing the correlation structure between the variables by applying principal component analysis; (iii) cokriging specific factors at each characteristic scale. The 14 REE concentrations available in the dataset were all used as inputs, as well as some physicochemical properties: pH, carbonate alkalinity, Fe, Mn and organic carbon. The obtained maps allow visualizing the factors which integrate REE spatial variability. This variability is discussed with some landscape factors, especially topography and upstream/downstream location.

2. Materials and methods

2.1 The study area

The geology of France results of a succession of events related to assembly and disruption of Gondwana and Pangea megacontinents and climate changes. Geographically, France has acquired a rugged topography giving it a wide range of outcropping terranes spanning from Proterozoic to Cenozoic (Figure 1). France can be divided into four geological terranes. (i) France is mostly covered by Mesozoic and Cenozoic deposits which correspond to intracratonic sedimentary basins like Paris Basin or Aquitaine Basin (Biteau et al., 2006; Guillocheau et al., 2000). These extensive areas have been scarcely deformed and are characterized by small dip values and concentric rock deposits. The tabular structure involves typical landscapes like alluvial plains, plateau and hills. In addition to intracratonic genesis, other basins correspond to grabens formed during Alps orogeny (Rhein and Limagnes graben). These units are generally depressions filled with Cenozoic sediments. (ii) These Mesozoic and Cenozoic sedimentary units lay on a basement composed of Paleozoic and Proterozoic rocks (Ballèvre et al., 2009; Faure et al., 2009). The basement is widely metamorphic and magmatic and is surrounded by discordant Mesozoic sedimentary cover. It constitutes the essential part of eroded mountain ranges like Vosges, Armorican Massif or French Massif Central erected during Cadomian and Variscan orogeny (Ballèvre et al., 2009; Faure et al., 2009). These reliefs have been strongly flattened by erosion, except to Vosges and French Massif Central which have been surrected again during Alps orogeny (Faure et al., 2009). (iii) Alps and Pyrenees are recent mountain ranges still erecting since Upper Cretaceous (Choukroune, 1992; Lagabrielle and Lemoine, 1997; Rosenbaum and Lister, 2005; Vissers and Meijer, 2012). These ranges form a complex association between meso-cenoizic rocks which have been heavily deformed due to fault and folding action and

basement units. The resulting landscape is high mountains (Valla et al., 2011). (iv) Recent volcanic ranges are located massively in the Massif Central (Michon and Merle, 2001). These units are mainly composed of basalts, trachytes, and rhyolites, aged from Neogene (Cantal) to Pleistocene (Chaîne des Puys).

2.2 The FOREGS dataset

The FOREGS program uses standardized field, analytical and quality control procedures to produce reliable reproducible geochemical data over Europe (Salminen et al., 1998). Therefore, the FOREGS initiated a program to construct a geochemical database with the aim of compiling the first geochemical atlas of Europe (Fedele et al., 2008). The field manual by Salminen et al. (1998) is the basis for a decade-long project, involving geochemists from 26 countries, which led to the publication of the Geochemical Atlas of Europe (De Vos and Tarvainen, 2006; Salminen et al., 2005) . As a result, a large geochemical database is now available free for public use¹ (De Vos and Tarvainen, 2006; Salminen et al., 2005).

2.2.1 Sampling strategy

The FOREGS sampling grid (Salminen et al., 1998; Tarvainen et al., 2005) was based on the Global Terrestrial Network (GTN) grid composed of 160 x 160 km cells and developed for the purpose of Global Geochemical Baselines Mapping (Darnley et al., 1995). For each cell, five randomly generated sites were selected, according to the following scheme:

- Point number 1 is located in the NE quadrant of the GTN grid cell;
- Point number 2 in the NW quadrant;

¹ <http://weppi.gtk.fi/publ/foregsatlas/index.php>

- Point number 3 in the SW quadrant;
- Point number 4 in the SE quadrant;
- Point number 5 is randomly located in anyone of the four quadrants of the GTN grid cell.

As a result, France was divided into 25 cells and 119 sample sites were determined. Based on former randomly generated points, five nearest small drainage basins of $<100 \text{ km}^2$ were selected. For each cell, a larger drainage basin (area 1,000-6,000 km^2), to which the small drainage basin is connected, was selected. The floodplain sediment samples were collected either from a suitable point near its outlet with the sea or the confluence point with another major river system. In this study, French stream water samples were selected. The dataset consists of 119 sampling sites (one sample per site). However, 4 sites located in Corsica were removed of the dataset to obtain a homogeneous spatial distribution. Samples were collected during two periods of the winter: from November 1998 to December 1998; and from March 1999 to October 1999. Sampling during rainy periods and flood events was avoided. According to FOREGS stream water sampling procedure, running stream water was collected from the small, second order, drainage basins ($<100 \text{ km}^2$). Physico-chemical parameters (pH, temperature and electrical conductivity) were measured at the site while several stream water samples were collected. ICP-MS analyses were performed on a 100 mL sample filtered to $0.45 \mu\text{m}$ (Salminen et al. 1998).

2.2.2 Chemical analyses and quality controls

As described in Sandström et al. (2005): stream water samples were acidified to 1% v/v with nitric acid and stored at less than 8°C . The samples were analyzed by both inductively coupled plasma quadrupole mass spectrometry (ICP-QMS) and inductively coupled plasma atomic emission spectrometry (ICP-AES), using Perkin Elmer Sciex ELAN

5000A and Spectro Flame M instruments respectively, in accordance with the German norms DIN 38406-29 (ICP-MS) and DIN 38406-22 (ICP-AES). Analyses were performed by BGR (Bundesanstalt für Geowissenschaften und Rohstoffe, Hannover, Germany). The primary chemicals used to prepare the calibration and quality control standards and reagents were of analytical reagent grade. Multi-element standard solutions for calibration were prepared from Claritas SPEX/Certiprep stock solutions. International certified reference materials (NIST 1640, NIST 1643d and SLSR-4) were included in every batch of 20 samples. Indium was used as an internal standard. The accuracy of the methods for all determinants is better than $\pm 10\%$, the bias is within $\pm 3\%$, and the repeatability at the 95% confidence interval is better than 5% at concentrations an order of magnitude above the limit of quantification. Limits of quantification for all cations and trace metals are given in Sandström et al. (2005).

2.3 Geographic information system and multivariate geostatistical approach.

Geographic information system (GIS) analysis was performed with ArcGis 10 software. Distance between sample site and outlet was calculated using the BD Carthage (Sandre Eau France) dataset which is the French official dataset about hydrographic network. Geological context of sampling sites was obtained by GIS intersection with 1/1,000,000 geological units (BRGM One Geology). Concentrations of 14 REE and physicochemical properties (pH, carbonate alkalinity as HCO_3 , Fe, Mn and organic carbon) were used as inputs for the geostatistical method. Descriptive statistics (mean, standard deviation sd, median med, and median absolute deviation MAD) were performed on all variables and normality of data and homogeneity of variances were verified. As data behave following a not normal distribution, non-parametric Kruskal-Wallis tests were performed to quantify differences in variables within geological contexts.

The multivariate spatial data were analyzed by cokriging and Factor Kriging Analysis (FKA) which is a geostatistical method developed by Matheron (1982). The FKA consists of decomposing the set of original second-order random stationary variables $\{Z_i(x), i = 1, \dots, n; \}$ into a set of reciprocally orthogonal regionalized factors $\{Y_v^u(x), v = 1, \dots, n; u = 1, \dots, N_s\}$ where N_s is the number of spatial scales, through transformation coefficients a_{iv}^u (loadings components score) combining the spatial with the multivariate decomposition:

$$Z_i(x) = \sum_{u=1}^{N_s} \sum_{v=1}^n a_{iv}^u Y_v^u(x)$$

The three basic steps of FKA are the following:

- (i) Modeling the coregionalization of the set of variables, using the so called Linear Model of Coregionalization (LMC);
- (ii) Analyzing the correlation structure between the variables, by applying Principal Component Analysis (PCA) at each spatial scale;
- (iii) Cokriging specific factors at each characteristic scale and mapping them.

2.3.1 Linear Model of Coregionalization

The LMC, developed by Journel and Huijbregts (1978), considers all the studied variables as the result of the same independent physical processes, acting at different spatial scales u . The $n(n+1)/2$ simple and cross variograms of the p variables are modeled by a linear combination of N_s standardized variograms to unit sill $g^u(h)$. Using the matrix notation, the LMC can be written as:

$$\Gamma(h) = \sum_{u=1}^{N_s} B^u g^u(h)$$

where $\Gamma(h) = [\gamma_{ij}(h)]$ is a symmetric matrix of order $n \times n$, whose diagonal and non-diagonal elements represent simple and cross variograms for lag h ; $B^u = [b^u_{ij}]$ is called coregionalization matrix and it is a symmetric semi-definite matrix of order $n \times n$ with real elements b^u_{ij} at a specific spatial scale u . The model is authorized if the functions $g^u(h)$ are authorized variogram models. In the LMC the spatial behavior of the variables is supposed resulting from superimposition of different independent processes working at different spatial scales. These processes may affect the behavior of experimental semi-variograms, which can then be modeled by a set of functions $g^u(h)$. The choice of number and characteristics (model, sill, range) of the functions $g^u(h)$ is quite delicate and can be made easier by a good experience of the studied phenomena (Chilès and Guillen, 1984). Fitting of LMC is performed by weighed least-squares approximation under the constraint of positive semi-definiteness of the B^u , using the iterative procedures developed by Goulard (1989). The best model was chosen, as suggested by Goulard and Voltz (1992), by comparing the goodness of fit for several combinations of functions of $g^u(h)$ with different ranges in terms of the weighted sum of squares.

2.3.2 Regionalized Principal Component Analysis

Regionalized Principal Component Analysis consists of decomposing each coregionalization matrix B^u into two other diagonal matrices: the matrix of eigenvectors and the diagonal matrix of eigenvalues for each spatial scale u through the matrix A^u of order $n \times n$ of the transformation coefficients a^u_{iv} (Wackernagel, 2003). The transformation coefficients a^u_{iv} in the matrix A^u correspond to the covariances between the original variables $Z_i(x)$ and the regionalized factors $Y^u_v(x)$.

2.3.3 Mapping multivariate spatial information

The behavior and relationships among variables at different spatial scales can be displayed by interpolating the regionalized factors $Y_v^u(x)$ using cokriging and mapping them (Castrignanò et al., 2007, 2000). The cokriging system in FKA has been widely described by Wackernagel (2003).

3. Results

3.1 Relationships between stream data and geology.

Table 1 shows the descriptive statistics of all variables between the geological contexts. The majority of the 115 samples is located in sedimentary areas ($n=85$), among them 40 are carbonate derived rocks, 9 chalks, 12 clays and 24 sand and sandstones. The others are located in metamorphic ($n=18$), acidic plutonic ($n=8$) and acidic volcanic ($n=3$) contexts. Most of physicochemical properties and selected REE show significant differences between the geological contexts: pH, carbonate alkalinity, Fe, La and Lu. This suggests the influence of surrounding rocks geology on water chemistry. Water samples taken from sedimentary areas are characterized by higher carbonate alkalinity concentrations (from 69mg/L to 134 mg/L) and alkaline pH value (from 7.68 to 8.02) which is explained by carbonates presence in sedimentary rocks. Higher Fe concentrations (176 mg/L and 224 mg/L respectively) are observed in both acidic plutonic and metamorphic areas. Regarding La, Eu and Lu concentrations, higher concentrations are observed in samples taken from metamorphic and plutonic areas (0.1491 $\mu\text{g/L}$ and 0.1374 $\mu\text{g/L}$ for La, respectively) whereas lower concentrations are derived from carbonate rocks (0.0280 $\mu\text{g/L}$ for La). This is consistent with

REE sources essentially located in metamorphic and magmatic rocks (from 50 to 100 mg/kg for La; Henderson, 1984; McLennan, 1989) whereas REE content in carbonates is low (10 mg/kg for La; Turekian and Wedepohl, 1961).

3.2 Principal Component Analysis

The classical statistical technique based on the PCA has been applied in order to investigate on the behavior of the variables in relation to the principal components. Since geochemical data are compositional, every data set should be opened, prior to its statistical treatment, using a preferred method from a variety of suggested methods (Sadeghi et al., 2014). In this study, statistical PCA has been performed using the ln-transformed data sets (Pawłowsky-Glahn and Buccianti, 2011). The generated results are given in Table 2. The PCA, performed in this study, using ln-transformed data has successfully opened the data. Using PCA, two principal components were extracted that cumulatively explained 90.6% of the total data variability and with an eigenvalue greater than 1 (Table 2). The other components had an eigenvalue less than 1 and were not used. Generally, the first components account for most of the variability contained in the data set (Johnson and Wichern, 2002). In the case study the first two principal components were used in the analysis, mainly due to the presence of such correlations with the real properties, as shown in the circle of correlations presented in Figure 2. The circle of correlation shows the proximity of the variables inside a unit circle and is useful to evaluate the affinity and the antagonism between the variables. Statement can easily be made about variables which are located near the circumference of the unit circle. In our case the first component is highly correlated with all REE and also Fe, while the Mn, OC, pH, carbonate alkalinity are located a little further away from the circumference and appear to have a different behavior. They all are correlated negatively with the F2. From an analysis of

the circle it is possible to see that all the HREE are positioned in the upper part and all the LREE in the lower part of the semi circumference. Europium shows a slightly different behavior as it is more detached from the circumference, and is close to Fe. From the PCA it appears that the first regionalized factor explains the behavior of the REE and separates them from the other elements. No substantial additional information is provided by the F2.

3.3 Coregionalization analysis

The variables are highly shifted from the Gaussian distribution so they were normalized and standardized to mean 0 and variance 1. On the basis of the results of principal component analysis, a choice has been made between the 19 variables. The chosen parameters are Fe, Mn, carbonate alkalinity, organic carbon and pH for the physicochemical properties together with La, Lu and Eu for the REE. La is representative of LREE, Lu of HREE and Eu has been chosen because of his peculiar behavior (McLennan and Taylor, 2012). Using GAMV of GSLIB library (Deutsch and Journel, 1992) the experimental variograms and cross-variograms of the 3 REEs and Fe, Mn, carbonate alkalinity, organic carbon and pH has been obtained with lag separation distance equal to 20 km. No relevant anisotropy was observed in the variogram maps and the experimental variograms looked upper bounded. The linear coregionalization model has been obtained using the LCMFIT2 program (Pardo-Iguzquiza and Dowd, 2002) to fit the 36 experimental variograms. The LMC was fitted using two spatial structures: a spherical model with a range of 120 km and a spherical model with a range of 250 km. The linear coregionalization models (direct and cross-variograms) (not shown) appear well spatially structured also due to the absence of the not-spatially correlated component (nugget effect). The spatial cross-correlation is shown in Table 3. The appropriateness of the LCM and the basic structures was evaluated with a cross-validation test by calculating the mean error and the variance of standardized error, which were quite close

to 0 (varying between -0.033 and 0.0145) and 1 (varying between 0.9 and 1.2), respectively. These results mean that the estimates were unbiased and the estimation variance reproduced the experimental variance accurately.

3.4 Factorial kriging

Using FACTOR2d program (Pardo-Iguzquiza and Dowd, 2002) a factorial cokriging is used to estimate the first 2 regionalized individual factors that, at the cost of an acceptable loss of information, have given a synthetic description of the process in study at the different selected spatial scales. As no nugget effect has been modeled, the short-range and long-range components of the first two regionalized factors were selected. The long range component (250 km) of the first two factors explains most variance (80.3% and 17.0%, respectively) while the short range component explains just the 50.0% and 20.2% for F1 and F2 respectively, which is less representative. The long range component of the first factor shows to be the most explicative as it is correlated with most of the elements positively, such as Fe (0.796), La (0.814), Lu (0.603), organic carbon (0.639) and negatively with HCO_3^- (-0.607), pH (-0.648) (Table 4). The F1 at long range synthesizes the long range variability of the whole elements with a localized higher value zone in the central and south-eastern part. The short range component (120 km) of the first factor is mostly correlated with Eu (0.773) and Mn (0.687). Europium shows a different behavior being mostly explained by the F1 at the short spatial scale. However, in order to get more precise information about its structure, additional data should be collected to infer the variability at a smaller spatial scale. As far as the F2, it does not show to be correlated at all with the REE but it is correlated more with the physico-chemical parameters, such as with the pH at small spatial scale (0.580).

3.5 Cokriging

Cokriging was applied to the transformed data to obtain the estimates which were then back-transformed to express them in the original variables. The spatial maps of the eleven variables were obtained by cokriging on a $10\text{ km} \times 10\text{ km}$ square grid. Figures 3, 4 and 5 represent, respectively, the first regionalized factor at short range and at long range, the 3 selected REEs and the 5 physicochemical properties. From an analysis of the cokriged maps it is possible to see that REE exhibit all similar spatial distributions, with extended higher values in the south-western part, in correspondence with the depression of Aquitaine basin (low area surrounded by higher land and usually characterized by interior drainage). Other zones of localized higher value correspond with the Alpine valley corridors of the Rhône, the hydrographic network of the Seine basin, hydrographic network of the Loire with its effluents. Another zone of higher values is localized in correspondence of the Oise basin, in the Paris basin. The anisotropy (SW-NE for the southern part and NW-SE for the northern part of France) showed in the maps of REE is coherent with the direction of propagation of rivers except for the Aquitaine basin where the high concentrations are due to the interior drainage of the low area. Organic carbon and Mn show similar behavior having higher zone values concentrated in the same areas of the REE, while carbonate alkalinity and pH, show a totally contrasting behavior with the rest of the variables (being negatively correlated).

4. Discussion

4.1. Assessment of factor kriging analysis

The long range component of the first factor shows to be the most explicative (80.3% of the variance) as it synthesizes the long range variability of most of the elements (Fe, La, Lu,

organic carbon, carbonate alkalinity and pH) with a localized higher value zone in the central and south-eastern part. Therefore at high spatial scales environmental parameters like Fe, carbonates, pH, are supposed be ascribed to the rock's nature or to other larger scale processes (i.e., river network). The structure of the hydrographic network is determined by a complex of physicogeographic conditions especially by climate, by the topography, and by the geological structure of the locality. Indeed, hydrographic network and topography have shown to affect REE distribution (e.g., Kohler et al., 2014). The short range component of the first factor explains just the 50.0% of the variance and is mostly correlated with Eu (0.773) and Mn (0.687). So at short range, just the Eu and Mn weigh more, which are ascribed to the process of Mn oxides release in rivers simultaneously with REE sorbed on these oxides (Koeppenkastrop and De Carlo, 1993). Indeed, Mn oxides behave like Fe oxides, and as evidenced by Steinmann and Stille (2008) saturation index of goethite decrease with transport and resulting REE fractionation in the stream water.

To sum up, only the first regionalized factor corresponding to long range is representative of the conjoint variability of the elements in study. The short range variability has not proved to explain the behavior of the variables satisfactorily. The sampling scale adopted is too wide and can give sufficient information for a correlation scale of hundreds of kilometers. To be able to infer the variation at a smaller scale, further sampling on a finer spatial scale would be needed.

4.2. Rare earth elements as tracer of hydrological transfer

Research results from past twenty years clearly show an influence of topography on stream water chemistry. Moreover several authors have proved that topography was the attribute that had the major influence on stream water chemistry (Andersson and Nyberg,

2009; Ogawa et al., 2006). The influence of topography is important because it controls the water subsurface contact time (Beven and Kirkby, 1979; Dillon and Molot, 1997; McGuire et al., 2005; Wolock et al., 1990). Topography is of great significance in hydrology, affecting soil water content, flowpaths and residence times (Nyberg, 1995), and subsequently the chemical composition of surface waters (Beven, 1986; Wolock et al., 1989). Such a feature was thus tested on REE as previously shown by Köhler et al. (2014): in boreal catchments REE export is mostly strongly controlled by landscape type. In continental systems, percolation of rain water through the rocks will result in low-temperature chemical weathering reactions that will slowly break down the primary minerals, possibly resulting in mobility of the REE. The chemistry of groundwater is clearly very dependent on the physicochemical environments through which it has passed.

A focus on the Garonne and Dordogne systems, in correspondence to the Aquitaine basin characterized by higher values and anisotropy, emphasized this feature. Indeed, both river systems originate in Massif Central, where low-temperature chemical weathering of acidic magmatic rocks occurs. Moreover, topography is escarped and results in low residence time. It is highlighted on figure 6, where REE patterns of considered samples depicted low REE concentration and negative cerium anomaly. Positive europium anomaly can be interpreted as a result of water rock interaction with feldspar from bedrocks (i.e., basalts; Steinmann and Stille, 2008). Both river systems then encountered the Aquitaine basin, with a more flat topography with higher residence time. It results in higher REE concentrations (figure 6), with middle REE enriched patterns, corresponding to an organic sedimentary input. Overall, spatial variability results in REE and organic carbon concentrations increase and a pH decrease (figure 7).

Wetlands would play a key role in the regulation of REE concentrations in the environment as earlier proposed (Davranche et al., 2014). Indeed REE are released in wetland

bound to colloidal organic matter as also observed in watershed or rivers (Shiller, 2010; Stolpe et al., 2013). Moreover, figure 7 shows the global control of dissolved REE concentrations by pH as previously highlighted for Nd by Johannesson and Burdige (2007) or Deberdt et al. (2002). Indeed, pH can significantly influence the speciation and thus the behavior of the REE (Pourret et al., 2007; Tang and Johannesson, 2003). A decrease in pH will favor solution of the REE and thus their transport either as organic complexes or as free ions. In figure 8, cerium anomaly in these two river water systems are reported as a function of the distance to the outlet. Figure 8 shows that the two river water systems plot along a single trend reflecting the Ce anomaly amplitude gradual reduction as the distance to the outlet decreases. It must be noted that a few points (n=3) have a different behavior from this trend and correspond to more organic water (associated to wetlands). As already proposed by Pourret et al. (2010) the likely reason for the Ce anomaly amplitude gradual reduction observed mainly relies on the fact that, in low permeability aquifers, water table generally reaches organic soil horizons in bottomland domains, thus allowing incorporation of large quantities of organic colloids in the aquifer bottomland part. This feature is not seen in the aquifer upland part where the water table always remains far below the upper, organic-rich soil horizons. Considering results from this study, it appears that the general feature of shallow groundwaters flowing into aquifers developed onto low permeability bedrock (Pourret et al., 2010) can be expanded to river water systems.

5. Concluding remarks

In order to further understand REE patterns spatial distribution, the FOREGS stream dataset was processed by means of geostatistical methods. The obtained maps allow visualizing the factors which integrate REE spatial variability. Cokriging shows a similar

spatial organization of REE: higher values are observed in the Aquitaine basin, more locally in the Alpine valley corridors of the Rhone, and along some tributaries of the Loire and the Seine. A factorial cokriging was applied to investigate more deeply the different sources of variation acting in the study area. The first 2 regionalized factors have been estimated to give a synthetic description of the studied process at the different selected spatial scales. At large spatial scales (250 km) environmental parameters like Fe, carbonates, pH, are supposed be ascribed to the rock's nature (plutonic, volcanic and metamorphic versus sedimentary) or to other larger scale processes, such as hydrographic network and topography have shown to affect REE distribution. The factor kriging analysis used in this study reveals the spatial patterns of REE in stream water. REE are positively correlated to Fe and Mn and negatively to carbonate alkalinity and pH. At short range, only Eu and Mn weigh more, which are ascribed to the process of liberation of Mn oxides in rivers that also releases the REE sorbed onto these oxides. This variability is discussed with some landscape factors, especially topography and upstream/downstream location. However, the short range variability cannot be explained satisfactorily by the adopted sampling as the sampling scale is too wide and can give sufficient information for a correlation scale of hundreds of km. To be able to infer the variation at a smaller scale, further sampling on a finer spatial scale would be needed.

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673 **Figure and table captions**

674 Fig. 1 Simplified maps of (a) elevation and (b) geological units.

675

676 Fig. 2 Principal component analysis: circle of correlations.

677

678 Fig. 3 First regionalized factor at short (structure 1) and at long range (structure 2).

679

680 Fig. 4 Cokrigged maps of (a) La, (b) Eu, and (c) Lu.

681

682 Fig. 5 Cokrigged maps of physicochemical properties (a) Fe, (b) HCO_3^- , (c) Mn, (d) organic
683 carbon, and (e) pH.

684

685 Fig. 6 Upper continental crust(UCC)-normalized REE patterns in samples from (a) Dordogne
686 river and (b) Garonne river (UCC values are from McLennan 2001).

687

688 Fig. 7 Concentrations of Nd in river samples as a function of pH (Garonne-Dordogne
689 watershed).

690

691 Fig. 8 Cerium anomaly as a function of distance to outlet (Garonne-Dordogne watershed).

692

693 Table 1 Descriptive statistics of selected REE and physicochemical properties.

694

695 Table 2 Principal component analysis: obtained components.

696

697 Table 3 Structural correlation coefficients.

698

699 Table 4 Correlation between factors and original variables.

Table 1 Descriptive statistics of selected REE and physicochemical properties (to be continued).

Geological context	pH					HCO3 (mg/L)				OC (mg/L)				Fe (µg/L)				Mn (µg/L)			
	n	mean	sd	med	MAD	mean	sd	med	MAD	mean	sd	med	MAD	mean	sd	med	MAD	mean	sd	med	MAD
acidic plutonic rock	8	6.99	0.42	7.00	0.20	30.12	31.74	18.17	5.24	5.02	3.28	4.05	1.80	176.48	114.55	150.00	81.05	16.24	10.69	13.75	6.95
acidic volcanic rock	3	7.03	0.85	6.70	0.30	31.14	8.61	34.00	3.95	1.82	0.77	1.85	0.73	74.90	30.96	83.00	18.00	11.42	6.07	10.20	4.15
Metamorphic rock (schist. gneiss...)	18	7.42	0.67	7.20	0.30	50.80	47.08	34.63	12.50	5.03	3.79	4.57	2.70	224.47	161.12	229.00	154.78	23.73	23.55	15.50	10.45
Carbonate sedimentary rock	40	8.02	0.67	8.10	0.10	230.87	114.18	277.98	65.84	2.89	3.70	1.97	0.90	46.78	68.82	18.23	9.60	17.45	28.19	3.61	3.24
Chalk	9	7.79	0.67	7.75	0.10	276.11	69.14	300.23	24.08	4.50	3.44	3.41	1.71	65.33	93.12	32.60	24.02	14.16	10.19	11.30	7.10
Clay	12	7.68	0.67	7.80	0.25	163.49	133.96	113.15	91.75	3.61	2.25	3.68	1.74	108.56	126.34	60.77	45.12	34.11	44.31	20.55	13.90
Sand and sandstone	24	7.80	0.67	8.00	0.20	174.60	130.85	131.67	106.31	4.35	2.96	3.80	1.95	137.25	211.21	37.40	28.39	25.73	28.36	15.71	11.18
no data	1																				
Kruskal-Wallis p-value		***				***				*				***				NS			

***: p<0.001, **: p<0.01, *: p<0.05, NS = non-significant.

Table 1 Descriptive statistics of selected REE and physicochemical properties (continued).

Geological context	n	La (µg/L)				Eu (µg/L)				Lu (µg/L)			
		mean	sd	med	MAD	mean	sd	med	MAD	mean	sd	med	MAD
acidic plutonic rock	8	0.1491	0.0905	0.1490	0.0680	0.0083	0.0036	0.0070	0.0020	0.0031	0.0014	0.0030	0.0010
acidic volcanic rock	3	0.0717	0.0071	0.0730	0.0050	0.0040	0.0010	0.0040	0.0010	0.0013	0.0006	0.0010	< 0.001
Metamorphic rock (schist, gneiss...)	18	0.1374	0.1076	0.1250	0.0930	0.0095	0.0057	0.0080	0.0040	0.0031	0.0019	0.0030	0.0020
Carbonate sedimentary rock	40	0.0280	0.0441	0.0170	0.0130	0.0055	0.0043	0.0050	0.0030	0.0013	0.0009	0.0010	< 0.001
Chalk	9	0.0720	0.1333	0.0190	0.0090	0.0049	0.0069	0.0020	0.0010	0.0018	0.0016	0.0010	< 0.001
Clay	12	0.0850	0.1267	0.0380	0.0240	0.0080	0.0087	0.0050	0.0030	0.0020	0.0018	0.0010	< 0.001
Sand and sandstone	24	0.0941	0.1278	0.0340	0.0260	0.0080	0.0078	0.0050	0.0030	0.0025	0.0026	0.0010	< 0.001
no data	1												
Kruskal-Wallis p-value		***				*				***			

***: p<0.001, **: p<0.01, *: p<0.05, NS = non-significant.

Table 2 Principal component analysis: obtained components.

Component	Eigenvalue	Variance (%)	Cumulative variance
F1	22.66	84.5	84.5
F2	1.62	6.1	90.6

Table 3 Structural correlation coefficients

Short range h = 120 km								
	Eu	Fe	HCO ₃	La	Lu	Mn	OC	pH
Eu	1.000							
Fe	0.640	1.000						
HCO ₃	-0.174	-0.227	1.000					
La	0.772	0.525	-0.474	1.000				
Lu	0.682	0.444	-0.409	0.568	1.000			
Mn	0.644	0.856	-0.156	0.533	0.315	1.000		
OC	0.735	0.409	0.000	0.251	0.364	0.400	1.000	
pH	-0.261	-0.181	0.403	-0.191	-0.331	-0.191	0.000	1.000
Long range h = 250 km								
	Eu	Fe	HCO ₃	La	Lu	Mn	OC	pH
Eu	1.000							
Fe	0.879	1.000						
HCO ₃	-0.987	-0.845	1.000					
La	0.734	0.958	-0.718	1.000				
Lu	0.865	0.995	-0.823	0.957	1.000			
Mn	0.109	0.561	-0.087	0.752	0.572	1.000		
OC	0.428	0.796	-0.355	0.858	0.802	0.873	1.000	
pH	-0.739	-0.922	0.750	-0.981	-0.920	-0.705	-0.752	1.000

Table 4 Correlation between factors and original variables.

<i>Short range h = 120 km</i>		
	Factor 1	Factor 2
Eu	0.773	0.146
Fe	0.493	0.106
HCO ₃	-0.256	0.437
La	0.441	-0.044
Lu	0.519	-0.148
Mn	0.687	0.183
OC	0.378	0.252
pH	-0.296	0.580
Explained variance (%)	50.0	20.2
<i>Long range h = 250 km</i>		
	Factor 1	Factor 2
Eu	-0.436	0.294
Fe	0.796	-0.074
HCO ₃	-0.607	0.452
La	0.814	0.104
Lu	0.603	-0.042
Mn	0.374	0.423
OC	0.639	0.393
pH	-0.648	-0.029
Explained variance (%)	80.3	17.0

Figure 1

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Figure 2

Figure 3
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Figure 6

Figure 7

Figure 8

Authorship Confirmation

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As corresponding author I, Olivier Pourret, hereby confirm on behalf of all authors that:

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