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Impact of spatial variability in hydraulic parameters on plume migration within unsaturated surficial formations

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Abstract

Heterogeneities in textural properties of surficial formations at field scale result in spatial variations in hydraulic parameters governing unsaturated zone flow. This study aims at quantifying the influence of such variations on solute transport resulting from a localized source of radioactive contaminant at ground level.

The study focuses on three hydraulic parameters related to the Mualem-van Genuchten formalism, namely the saturated hydraulic conductivity $K_s$, the parameter $\alpha$ inversely proportional to the air-entry value, and the parameter $n$ related to the pore-size distribution. Sets of random fields accounting for spatial variability of these parameters are generated using lognormal distributions with different variances and correlation lengths. These random fields are used as inputs to an unsaturated flow and transport model to simulate radionuclide plume migration.

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Each simulated plume is characterized by its size (plume surface area), position (location of center of mass) and shape (elongation ratio) within the unsaturated zone. By comparison with the homogeneous medium, $K_r$, $\alpha$- and $\gamma$- random fields generated with the mean variances computed through the analysis of a global soil database respectively result in average in (i) 25 (variable $K_r$), 20 (variable $\alpha$) and 65% (variable $\gamma$) increase in plume size; (ii) 0.8, 1 and 1.8 m horizontal offsets of the plume center; and (iii) 20, 30 and 50% decrease in plume circularity. In addition, changes in the variance values within one order of magnitude appear to have critical consequences only for the $\gamma$ parameter.

The issue of spatial variability of hydraulic parameters is thus crucial for characterizing the evolution of pollutant plumes within an unsaturated zone and for developing better remediation strategies for industrial sites.

Keywords: Richards equation, Mualem-van Genuchten model, Random fields, Radionuclide migration, Sensitivity analysis.

1. Introduction

Spatial heterogeneities found in porous formations at field scale impact groundwater flow and solute transport. The role of geometrical patterns, especially connected features, in concentrating flow and reducing travel time has been widely highlighted, in both saturated (Knudby and Carrera, 2005; Renard and Allard, 2013) and unsaturated (Appels et al., 2018) zones. Most of the related studies assume that such patterns mainly result from the arrangement of large-scale discrete geological structures with high contrast in hydraulic parameters (Feyen and Caers, 2006; Ye and Khaleel, 2008).
spatial distribution of these parameters can be represented by either several unimodal random fields populating a discontinuous facies model (Matheron et al., 1987; Le Coz et al., 2011; Le Coz et al., 2013) or a single continuous random field whose properties are multimodal (Russo, 2012).

Smaller scale variability, i.e. within individual facies, is known to contribute significantly to the overall hydraulic parameters variability within the unsaturated zone (Botros et al., 2009). When considering such individual facies, the corresponding parameters are generally represented by a single population whose distribution is unimodal, more particularly lognormal for the hydraulic conductivity (Fogg et al., 1998; Paleologos et al., 2015; Tan et al., 2017), and whose spatial variability can be characterized through a two-point spatial covariance (Botros et al., 2009; Russo, 2012). However, identifying a relevant covariance model requires a large amount of data at field scale (e.g., Botros et al., 2009) and is a complex task because the measurements of hydraulic parameters that govern unsaturated zone flow are difficult and time-consuming (Shaap et al., 2004).

The spatial variability within individual facies is thus often neglected, for instance in most of the studies that focus on the spread of radioactive contaminants in the unsaturated zone (Skuratović et al., 2016; Testoni et al., 2017). In particular, Testoni et al. (2017) simulated the barrier effect of the unsaturated zone as a delay and capture system by considering various medium configurations and initial conditions in case of an accidental release of Cesium 137. These simulations were performed by coupling of (i) a distributed one-dimensional unsaturated flow and transport model and (ii) a three-dimensional saturated flow and transport model. This coupling is
thus based on the assumption that the unsaturated water flow is mainly in
the vertical direction. This commonly formulated assumption (Bugai et al.,
2012; Jakimavičiūtė-Maselienė et al., 2016) is questionable since anisotropy
induced by lateral variability in hydraulic parameters can refract flow lines
away from vertical (Gannon et al., 2017).

This study aims at quantifying the influence of the spatial variability in
hydraulic parameters on plume migration in unsaturated surficial formations,
within the framework of a punctual release of radionuclides to the subsurface.
First, the degree of spatial variability in hydraulic parameters at field scale,
for typical soils, is assessed based on the analysis of a global soil database.
Then, numerical simulations are run using a full two-dimensional (2D) flow
and transport model to reproduce lateral flow induced by such spatial vari-
ability.

2. Unsaturated flow in porous media

This section recalls the main parameters governing flow in unsaturated
surficial formations and quantifies their spatial variability in natural environ-
ment. The variance of three of these parameters is computed at field scale
based on soil texture information recorded in a global soil database.

2.1. Governing equation and parameterization

Varibly saturated flow processes in porous media are typically described
by the Richards equation:

$$\frac{\partial \theta(\psi)}{\partial t} = \nabla \cdot (K(\psi) \nabla \psi) + q_{so}$$  \hspace{1cm} (1)
where $\theta$ is the volumetric moisture content \([L^3.L^{-3}]\), $t$ is the time \([T]\), $K$ is the hydraulic conductivity tensor \([L.T^{-1}]\), $\psi$ is the water pressure head \([L]\), $x$ is the vertical coordinate directed upward \([L]\) and $q_{ss}$ represents distributed source (positive) or sink (negative) terms \([L^3.L^{-3}.T^{-1}]\). Solving the Richards equation requires the moisture retention curve and the relative hydraulic conductivity function as inputs describing the links between pressure head, water content and relative hydraulic conductivity. These relationships are based on a model, e.g., the Mualem-van Genuchten model (Mualem, 1976; van Genuchten, 1980):

\[
\theta(\psi) = \theta_s + \left( \frac{\theta_s - \theta_r}{1 + (|\alpha|\psi)^n} \right)^m \quad \text{with} \quad m = 1 - \frac{1}{n} \quad (2)
\]

and

\[
K(\psi) = K_s \theta(\psi) \left[ 1 - (1 - S_e^{\theta(\psi)} \theta_s - \theta_r)^2 \right] \quad \text{with} \quad S_e = \frac{\theta_s - \theta_r}{\theta_s - \theta_r} \quad (3)
\]

where $\theta_r$ and $\theta_s$ are respectively the residual and saturated volumetric water contents \([L^3.L^{-3}]\), $\alpha$ is inversely proportional to the air-entry value \([L^{-1}]\), $n$ is a pore-size distribution index \([-\]) and $K_s$ is the saturated hydraulic conductivity tensor \([L.T^{-1}]\).

Although the hydraulic parameters are linked to physical soil properties, the direct measurement of some of them from samples is subject to experimental limitations (Schaap et al., 2004). Many investigations are thus undertaken to estimate hydraulic parameters using empirical relationships deduced from more readily available data, such as soil texture and bulk density (e.g., Wosten et al., 1999; Schaap et al., 2001; Saxton and Rawls, 2006; Tóth et al., 2015; Zhang and Schaap, 2017). These relationships, commonly
referred as pedotransfer functions (PTFs), are mostly based on regression
analysis of existing soil databases. The ROSETTA PTF (Schaap et al.,
2001) is based on artificial neural network analysis coupled with bootstrap
re-sampling which allows to estimate hydraulic parameters of the Mualem-
van Genuchten model.

The parameter \( \theta_r \) can be quite precisely derived from soil texture using
PTFs (Vereecken et al., 1989); both \( \theta_r \) and \( \theta_s \) can be directly obtained from
measurements under extremely dry or saturated conditions at the subsur-
face respectively (Jadoon et al., 2012). In addition, preliminary sensitivity
analyses performed in temperate climate conditions show that the spatial
variability of these two parameters does not significantly influence the solute
transport. The sensitivity analysis conducted in this study thus only focuses
on the \( K_s, \alpha \) and \( n \) parameters.

2.2. Variability of hydraulic parameters at field scale

2.2.1. Database analysis

The World Soil Information Service (WoSIS) aims at providing consistent
harmonized (i.e., georeferenced, quality-assessed and standardized) soil data
on a global scale based on soil profiles compiled by the International Soil Ref-
erence and Information Center (ISRIC). The WoSIS database currently con-
tains some 96,000 georeferenced soil profiles, described in terms of analytical
and physical soil properties, among which 20% were so far quality-assessed
and standardized (Batjes et al., 2017). The number of recorded data for each
property varies between profiles; yet a majority of samples is described at
least in term of soil texture, i.e. sand, silt and clay contents. These textural
properties are thus used to estimate the Mualem-van Genuchten hydraulic
parameters by means of the ROSETTA PTF. The soil profiles from the WoSIS database are then grouped by locations. A location is defined as a circular area of radius 500 m that contains at least 10 soil profiles. For each of the 49 identified locations, the mean ($\mu$) and variance ($\sigma^2$) of log$K_s$, log$\alpha$ and log$n$ (the parameters $K_s$, $\alpha$ and $n$ are assumed to follow lognormal distributions) are computed based on the whole corresponding soil samples. The Pearson correlation coefficients ($r$) are also computed between the three parameter pairs, i.e., log$K_s$ and log$\alpha$; log$K_s$ and log$n$; and log$\alpha$ and log$n$.

2.2.2. Variances and correlations

On the 49 identified locations, the log-variances (which enables to work with a dimensionless parameter $\sigma^2$) of the hydraulic parameters generally vary within about one order of magnitude (Figure 1) in relation with the mean parameter values. Indeed, $\sigma^2$(log$K_s$) increases from $2.5 \times 10^{-2}$ (1st decile) to $2.5 \times 10^{-1}$ (9th decile) when $\mu(K_s)$ varies from 0.1 to 0.6 m.d$^{-1}$; $\sigma^2$(log$\alpha$) decreases from $8 \times 10^{-2}$ to $8 \times 10^{-3}$ when $\mu(\alpha)$ varies from 0.9 to 2.7 m$^{-1}$; and $\sigma^2$(log$n$) increases from $8 \times 10^{-4}$ to $1 \times 10^{-2}$ when $\mu(n)$ varies from 1.25 to 1.65.

The Pearson correlation coefficients between the parameters also appear to depend on the mean parameter values (Figure 2). For $\mu(K_s)$ lower than 0.2 m.d$^{-1}$, $r$ values are highly scattered: from -0.2 to 0.8 between log$K_s$ and log$\alpha$; from -0.7 to 0.9 between log$K_s$ and log$n$; and from -0.9 to 0.5 between log$\alpha$ and log$n$. For $\mu(K_s)$ higher than 0.2 m.d$^{-1}$, $r$ values are less scattered and tend to stabilize: from 0.3 to 0.7 between log$K_s$ and log$\alpha$; from 0.7 to 1 between log$K_s$ and log$n$; and from -0.1 to 0.6 between log$\alpha$ and log$n$.
3. Methods

This section describes the modeling tools and approaches carried out for assessing the influence of spatial variability in $K_s$, $\alpha$ and $\eta$ on the migration of a radionuclide plume in unsaturated surficial formations. A 2D hydrogeological flow and transport numerical model is run in simulation mode with various input parameter fields representing distinct levels of spatial variability.

3.1. Flow and transport numerical model

The MELODIE numerical code developed by the French Institute for Radiation Protection and Nuclear Safety (IRSN) aims at making available a tool for evaluating the long term safety of a radioactive waste disposal facility (IRSN, 2009; Amor et al., 2014; Amor et al., 2015; Bouzid et al., 2018). This code simulates underground flow and solute transport in saturated or variably saturated porous media based on a mixed finite volume - finite element scheme, namely, Godunov development for the convective term and Galerkin development for the diffusion - dispersion term (Amaziane et al., 2008). In this study, MELODIE is set for solving in 2D (i) the Richards equation describing flow in variably saturated porous media (Eq. 1); and (ii) the following advection-dispersion-reaction equation representing the migration of radionuclides:

$$\nabla \cdot [(D|\nabla| + \omega d)\nabla C - \nabla C] = \omega' \frac{\partial C}{\partial t} + \omega \lambda RC$$  

(4)

where $C$ is the volumetric radionuclide concentration [M.L$^{-3}$], $D$ is the dispersivity tensor [L], $d$ is the molecular diffusion coefficient [L$^2$.T$^{-1}$], $\nabla$ is
the pore water velocity \([L \cdot T^{-1}]\), \(\omega\) and \(\omega'\) are respectively total and effective porosities \([-]\), \(\lambda\) is the decay constant \([T^{-1}]\) and \(R\) is the retardation factor \([-\]).

The modeling domain (Figure 3a) consists of a 2D vertical section of surficial deposits of 100 m large (x axis) by 15 m deep (z axis), discretized in triangles of base 0.5 m and height 0.25 m. The boundary conditions are defined as follows:

1. a fixed head corresponding to the mean water table elevation (7.5 m above the bottom boundary with a 0.004 m.m\(^{-1}\) lateral gradient) is set on both sides of the domain;
2. no-flow conditions are set on the bottom boundary;
3. a time variable flow corresponding to the daily percolation rate, typical from center of France, and estimated from the water balance method (Thornthwaite and Mather, 1955) is imposed on the top boundary (Figure 3b).

A point source of tritium is simulated by setting an activity of 1,000 Bq.d\(^{-1}\) during one month on the top surface node located at the center of the modeling domain. The evolution of the activity within the domain is then simulated during six years with an adaptative time stepping (from \(10^{-20}\) to 1 day) by considering a retardation factor \((R)\) of 1 and a decay constant \((\lambda)\) of \(1.54 \times 10^{-4} \text{ d}^{-1}\) (Table 1). The tritium plume is delimited by the nodes for which the simulated activity is higher than 1 Bq.m\(^{-3}\)\(\text{H}_2\text{O}\).

### 3.2. Parameter fields

First, the internal model parameters are fixed homogeneously for the whole domain as representative of sandy loam material (Table 2). Then, sets
of 100 simulations are run by considering (i) one hydraulic parameter (either $K_s$ or $\alpha$ or $\eta$) variable in space through sets of 100 random fields with similar properties (i.e., same parameters used for the geostatistical simulations of random fields) and (ii) three hydraulic parameters ($K_s$ and $\alpha$ and $\eta$) varying simultaneously through sets of 100 correlated random fields (Table 3).

Hereafter, $Z$ is a random field referring indifferently to one of the hydraulic parameters under consideration. Since $K_s$, $\alpha$ and $\eta$ are assumed to follow lognormal distributions, the logarithm of $Z$ is a gaussian random field such that:

$$\log Z \sim \mathcal{N}(\mu, \sigma^2)$$

where $\mu$ and $\sigma^2$ are the mean and variance of $\log Z$, respectively. The variogram chosen to account for the spatial variability of $\log Z$ is the spherical model, which was shown to fit experimental data of soil hydraulic parameters (Herbst et al., 2006; Botros et al., 2009; Bevington et al., 2016):

$$\gamma(h) = \begin{cases} \frac{3}{2} \frac{|h|}{\delta} - \frac{1}{2} \left( \frac{|h|}{\delta} \right)^3, & \text{if } |h| \leq \delta \\ 1, & \text{if } |h| > \delta \end{cases}$$

where $\gamma(h)$ is the variogram value for pairs of points separated by a distance $|h|$ and $\delta$ is the range (correlation length). In order to take into account the anisotropy of geologic structures, a geometric anisotropy with different ranges along the horizontal and vertical directions (namely $\delta_x$ and $\delta_z$) is introduced. The simulations of $\log Z$ are performed by keeping a constant hazard as far as possible. Thus, the following sensitivity analysis takes into account the effect of changes in simulation parameters, rather than the effect of random sampling (Nguyen et al., 2018).
Each set of 100 random fields is built with specific $\sigma^2$, $\delta_X$ and $\delta_Z$ values (Table 3) defined consistently to studies performed at field scale showing that a significant spatial correlation typically exists within distances ranging from a few meters to 20 m (Taskinen et al., 2008). Yet, the $\mu$ value is fixed so that the geometric mean of $Z$ corresponds to the mean sandy loam properties (Table 2).

3.3. Sensitivity analyses

For each of the three hydraulic parameters under study, the sensitivity of the simulated tritium plume to the properties ($\sigma^2$, $\delta_X$ and $\delta_Z$) of the 100 corresponding random fields is analyzed with regards to the following features (only the unsaturated zone is considered):

1. the surface area of the plume, defined as the number of grid cells with a volumic activity higher than 1 Bq.m$^{-3}$.H$_2$O;

2. the distance between the center of mass of each plume and the center of mass of the plume simulated in the homogeneous medium (where $\sigma^2(\log K_z) = \sigma^2(\log a) = \sigma^2(\log b) = 0$);

3. the elongation ratio ($a/b$) of the equivalent ellipse (Figure 4), defined as the ellipse whose center coincides with the center of mass of the plume and whose semi-minor and semi-major axes lengths (namely $a$ and $b$) are proportional to the eigen values of the inertia matrix.

The number of simulations is set to 100 to ensure that the averages of the previous features reach a stabilization. The modeling conditions (initial and boundary conditions, transient fluxes, source term) are similar for all the
simulations (see section 3.1). Only the soil properties \((\sigma^2, \delta_X\) and \(\delta_Z\)) differ from one simulation to another.

3.4. Additional tests

In order to assess the effect of both radionuclide type and soil texture, two additional tests are conducted by considering different model parameterizations:

1. the tritium (ideal groundwater tracer) is replaced with a reactive radionuclide (e.g., \(^{90}\)Sr) with a retardation factor \((R)\) of 3 and a decay constant \((\lambda)\) of \(6.59 \times 10^{-4} \text{ d}^{-1}\) (Table 1);

2. the sandy loam mean properties are replaced with silty clay loam mean properties, i.e. related to significantly finer texture (Table 2).

4. Results

This section presents the results averaged over each set of simulations for the last modeling time step, i.e. six years after the radionuclide injection. In order to assess the influence of the variance of hydraulic parameters on the plume features, the focus is first put on simulations with one hydraulic parameter varying (Table 3, sets 1., 2., 3.) and then on simulations with the three parameters varying simultaneously (Table 3, sets 4.). Then, the influence of the range values is described.

4.1. Sensitivity to the variance of one hydraulic parameter

Whatever the hydraulic parameter under consideration \((K_s\) or \(\alpha\) or \(\alpha\)), the increase in the variance of the corresponding random field results both
in the increase in the average plume surface areas (from 220, corresponding
to the homogeneous plume, to 350 grid cells) and in the variability of this
surface areas (standard deviation ranging from 0, corresponding to the ho­
mogeneous plume, to 110 grid cells) among the simulated plumes (Figure 5).
The relationships between $\sigma^2$ and the surface areas of the plumes appear to
be similar for the three hydraulic parameters under study when considering
$\sigma^2(\log K_s)/\sigma^2(\log n) = 200$ and $\sigma^2(\log K_s)/\sigma^2(\log a) = 5$.

In addition, the increase in variance for any hydraulic parameter results
in the increase of the offset in both directions of the mean center of mass
regarding the center of the homogeneous plume (Figure 6). The trend is yet
stronger in the horizontal direction (mean offsets ranging from 0.5 to 2 m)
than in the vertical direction (offsets ranging from 0.5 to 0.7 m). Besides,
the relationships between $\sigma^2$ and the offsets are similar when considering
$\sigma^2(\log K_s)/\sigma^2(\log n) = 200$ and $\sigma^2(\log K_s)/\sigma^2(\log a) = 5$.

Finally, the increase in variance results in the decrease of the elongation
ratio (from 0.88, corresponding to the homogeneous plume, to 0.45), re­
gardless of which hydraulic parameter is considered (Figure 7). Thus, when
the variability of hydraulic parameters is high, pollutant plumes are more
likely to have an elongated shape. Besides, the relationships between $\sigma^2$
and the offsets are similar when considering $\sigma^2(\log K_s)/\sigma^2(\log n) = 200$ and
$\sigma^2(\log K_s)/\sigma^2(\log a) = 5$.

4.2. Sensitivity to the variance of the three hydraulic parameters

The mean and standard deviation of the plume surface areas, the mean
offsets between each centers and the homogeneous center and the mean elon­
gation ratios are computed for sets of 100 simulations with the three para­
ters varying simultaneously and in a correlated manner, with $\sigma^2(\log K_s)/\sigma^2(\log n) = 200$ and $\sigma^2(\log K_s)/\sigma^2(\log \alpha) = 5$.

These features are close to the values obtained for the sets with only one parameter varying, considering the corresponding variances (Figures 5, 6 and 7). Here, the results are given for a coefficient $r$ of 0.8. Other values of $r$ coefficient have been tested and lead to similar results.

4.3. Sensitivity to the ranges

Given the values of $\sigma^2$, $\delta x$ and $\delta z$ chosen for this sensitivity analysis, the range in the horizontal direction ($\delta x$) does not seem to have a significant influence on the simulated plume features (size, center of mass and shape), regardless of which hydraulic parameter is considered (Figure 8). The influence of the range in the vertical direction ($\delta z$) appears to be slightly more visible (Figure 9), especially regarding the mean plume surface areas and the location of the mean centers of mass: when the vertical range increases from 1 to 5 m, the mean plume surface areas decrease, especially for $K_s$ and $n$ (diminution from 300 to 250 and from 290 to 260 grid cells respectively) and the horizontal offsets regarding the homogeneous plume center increase of around 0.3 m for the three hydraulic parameters.

4.4. Additional tests

When considering a reactive radionuclide ($R = 3$), the simulated plumes are logically less developed than for the ideal groundwater tracer ($R = 1$). However, both the surface area and the related standard deviation increase significantly when the variance of each of the three hydraulic parameters
considered increase. These features tend to stabilize when the average surface reaches about 130% of the surface simulated in a homogenous medium, consistently with the results for tritium (Figure 10).

The results obtained with mean hydraulic parameters related to a finer texture (silty clay loam) are similar to those above presented for sandy loam (Figure 10a). This could be explained by the fact that silty clay loam mean properties remain sufficiently favorable for leading the infiltration of the water height imposed at the top surface of the model (Figure 3b). The standard deviation computed for highest variances of the hydraulic parameters is yet lower (Figure 10b). This could be explained by the fact that a large distribution of the hydraulic parameters centered on sandy loam material could result in very high hydraulic conductivity values more favorable to the development of preferential flow paths and therefore to the lateral migration of the plume.

5. Discussion

This study allows to quantify the impact of the variance of the random fields representing the logarithms of the hydraulic parameters $K_s$, $\alpha$ and $\eta$ on solute transport within unsaturated surficial formations. By focusing on variances estimated for porous media in natural environment at field scale, it appears that the $\eta$ and, in a lesser extent, $K_s$ parameters are the most critical (Figures 5, 6 and 7). Thus, by comparison with homogeneous parameter fields, random fields generated with $\sigma^2(\log K_s) = 0.14$, $\sigma^2(\log \alpha) = 0.036$ and $\sigma^2(\log \eta) = 0.005$, i.e. mean variances calculated based on the WoSIS database, respectively result in (i) 25, 20 and 65% increase in plume size;
(ii) 0.8, 1 and 1.8 m horizontal offsets of the plume center; and (iii) 20, 30 and 50% decrease in plume circularity. Besides, within the range of variation of $\sigma^2(\log n)$, the average surface areas range from 260 to 350 grid cells. In comparison, within the range of variation of $\sigma^2(\log K_s)$ and $\sigma^2(\log e)$, the mean surface areas only range from 235 to 290, which highlights that $n$ is the most critical parameter. Such results are consistent with previous studies showing the high influence of the spatial variability of the parameter $n$ on flow (Lu and Zhang, 2002 and Tan et al., 2017) in the unsaturated zone.

This study also shows that the simulated plumes are far less sensitive to changes in ranges than changes in variances of the random fields representing the hydraulic parameters, at least for the intervals of values chosen. Yet, the impact of the vertical range appears to be more pronounced than the horizontal range: when the zones of high or low values of hydraulic parameters are more expanded in depth (higher value of vertical ranges), the preferential flow paths are potentially more continuous and the plume surface areas tend to decrease and get closer to the homogeneous plume surface area. The values of horizontal ranges considered here are of the same order of magnitude as the pollutant plumes extensions (a few meters): it could explain the low influence of the changes in horizontal range on the simulated plumes. Nevertheless, these ranges are consistent with analyses of soil parameter variability at field scale (Russo et al., 1997 and Botros et al., 2009). However, for higher values of ranges (from hundreds of meters to a thousand meter), the influence of horizontal range on the flow rate in unsaturated zone was shown to be more significant (Tan et al., 2017).

The sensitivity analysis performed in this study mainly focuses on an
ideal groundwater tracer (tritium) in sandy texture. However, additional
tests show that the results obtained for reactive radionuclides and a finer
texture are comparable (at least for similar infiltration rate). From a general
standpoint, this study thus indicates that taking into account the weak intra-
facies (small scale) spatial variability is critical for characterizing plume mi-
gration within unsaturated zone. This requires to develop in-situ approaches
for estimating more efficiently soil hydraulic parameters and their variabil-
ity at field scale (e.g., Léger et al., 2014; Léger et al., 2016). When these
parameters remain poorly recognized, a set of simulations based on typical
variability deduced from local or global databases can be used for identifying
the plume properties and the related uncertainties at a given time.

6. Conclusion

This work has highlighted the critical influence of the spatial variability
in hydraulic parameters, in particular pore-size distribution index ($n$), on
radionuclide migration in unsaturated surficial formations. In practice, this
work demonstrates the need for well characterizing this spatial variability
in order to quantify and locate soil volumes that need to be removed or
depolluted in the context of remediation of nuclear or industrial sites. It also
quantifies the corresponding uncertainties.

In addition, although the quantitative results presented in this paper are
related to a theoretical example and are dependent on the model settings
(environmental parameters, radionuclide, etc.), the proposed method (ran-
dom field generation, physically-based groundwater transport simulations,
plume properties characterization) can be transposed to other contexts, such
as chemical pollutions or design of devices to contain pollutant spread.

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Figure 1: Log-variances in function of means for $K_s$, $a$ and $n$ for each location built via WoSIS database analysis.

Figure 2: Pearson’s correlation coefficients between the log-parameters in function of $\mu(K_s)$ for each location built via WoSIS database analysis.

Figure 3: (a) Modeling domain with an example of hydraulic conductivity random field and resulting tritium plume. (b) Daily percolation rate and cumulative percolation rate used for the simulations.

Figure 4: Example of pollutant plume within the unsaturated zone with its equivalent ellipse in dotted line. The $a/b$ ratio gives a hint about the elongation of the plume: the more $a/b$ is close to 0, the more the plume has an elongated shape. The triangle highlights the tritium source.

Figure 5: Average plume surface areas (a) and standard deviation of surface areas (b) depending on $\sigma^2$ for the three hydraulic parameters. The squares correspond to the homogeneous medium and the three upside-down triangles on each graph correspond to media where the three hydraulic parameters are varying in a correlated manner. The values of $\delta_x$ and $\delta_z$ are set to 10m and 3m. The shaded rectangles on the axes correspond to the range of variances found at field scale via WoSIS database.

Figure 6: Mean distances with the center of the plume simulated in the homogeneous medium in function of $\sigma^2$ for the three hydraulic parameters. The square corresponds to the homogeneous medium and the six upside-down triangles correspond to the media where the three hydraulic parameters are varying in a correlated manner. The values of $\delta_x$ and $\delta_z$ are set to 10m and 3m. The shaded rectangles on the axes correspond to the range of variances found at field scale via WoSIS database.

Figure 7: Mean $a/b$ ratio in function of $\sigma^2$ for the three hydraulic parameters. The square corresponds to the homogeneous medium and the three upside-down triangles correspond to the media where the hydraulic parameters are varying in a correlated manner. The values of $\delta_x$ and $\delta_z$ are set to 10m and 3m. The shaded rectangles on the axes correspond to the range of variances found at field scale via WoSIS database.
Figure 8: Average (a) and standard deviation (b) of the plume surface areas depending on the horizontal range for the three hydraulic parameters. Average distances (c) with the homogeneous center and mean elongation ratio (d) in function of the horizontal range. The values of $\sigma^2$ are set to 0.2 for log$K_s$, 0.04 for log$\alpha$ and 0.001 for log$n$ and $\delta_z$ is set to 3m.

Figure 9: Average (a) and standard deviation (b) of the plume surface areas depending on the vertical range for the three hydraulic parameters. Average distances (c) with the homogeneous center and mean elongation ratio (d) in function of the vertical range. The values of $\sigma^2$ are set to 0.2 for log$K_s$, 0.04 for log$\alpha$ and 0.001 for log$n$ and $\delta_z$ is set to 10m.

Figure 10: Average (a) and standard deviation (b) of the plume surface areas depending on $\sigma^2$ for the three hydraulic parameters. The black line corresponds to the results obtained for tritium and sandy loam texture (for the parameter $K_s$). The filled symbols correspond to the results for the reactive radionuclide and sandy loam texture. The empty symbols correspond to the results for tritium and silty clay loam texture.
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Table captions (revised)

**Table 1:** Transport parameters used in (Eq. 4) for tritium and for a reactive radionuclide.

**Table 2:** Soil hydraulic (flow) parameters derived from ROSETTA PTF for sandy loam and silty clay loam materials, used in (Eq. 1), (Eq. 2) and (Eq. 3).

**Table 3:** Values of variances, ranges and Pearson's correlation coefficients for each set of 100 simulations.