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## Accepted Manuscript

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1  **$\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  in the Mondego estuary food web: seasonal variation in producers and**  
2 **consumers**

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18  
19  
20 **Abstract**

21 Assessments of temporal variation in stable carbon and nitrogen ratios were used to  
22 examine seasonal trends of the water column and benthic food webs in the Mondego estuary  
23 (Portugal). There was a marked seasonality in weather and water column conditions,  
24 including nutrient supply and chlorophyll concentrations. In spite of the pronounced  
25 environmental changes, we found little evidence of seasonal variation in  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  of  
26 producers and consumers in the Mondego estuary, with a few notable exceptions. Nitrogen  
27 isotope ratios in macrophytes (*Zostera noltii*, *Ulva* sp., *Enteromorpha* sp., *Gracilaria* sp.),  
28 and in two grazers (*Idotea chelipes*, *Lekanesphaera levii*) increased during late summer, with  
29 the highest  $\delta^{15}\text{N}$  values being measured in July, during a period of elevated temperatures and  
30 drought, which may have favored high rates of denitrification and heavier  $\delta^{15}\text{N}$  values. The  
31 results suggest that stable isotope values from macrophytes and selected grazers are useful as  
32 tracers of seasonal changes in nitrogen inputs into estuaries, and that those of consumers  
33 reflect other factors beyond seasonal variations in N and C sources.  
34

35 *Keywords:* Mondego estuary;  $\delta^{13}\text{C}$ ;  $\delta^{15}\text{N}$ ; Coastal eutrophication; Seasonality; Estuarine  
36 species

37

## 38 **1. Introduction**

39 Increased anthropogenic delivery of nutrients to water bodies, both freshwater and  
40 estuarine, has caused detrimental changes in habitat, food web structure, and nutrient cycling  
41 (Valiela et al., 1997; Cole et al., 2004). The resulting eutrophication has many adverse effects  
42 within the estuaries (D'Avanzo et al., 1996; Hauxwell et al., 2003). Increased N loading can  
43 lead to the loss of important estuarine habitats like seagrass meadows (Hauxwell et al.,  
44 2003). Eutrophic estuaries can also suffer from hypoxia and anoxia (Zimmerman and Canuel,  
45 2000), and phytoplankton and macroalgal blooms (Hauxwell et al., 2003).

46 To better understand management of water quality it is important to know the sources,  
47 as well as the amount of inputs of the nutrient limiting production. In the Mondego estuary,  
48 as in most estuarine ecosystems, there is evidence that at least for macroalgal growth,  
49 nitrogen is the limiting factor (Teichberg et al., submitted).  $\delta^{15}\text{N}$  has proven useful as a tracer  
50 of the major source of nitrogen entering coastal waters. Joint use of  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  have  
51 further shown promise as a tool that helps to explain how the external N sources, as well as  
52 the C sources, move up into estuarine food webs. Application of these isotopic ratios has  
53 largely remained an item of research rather than a management tool (Peterson and Fry, 1987;  
54 Cole et al., 2004). The practical utility of stable isotopic ratios to some degree depends on the  
55 relative sensitivity of  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  to seasonal variation.

56 Stable isotopic N ratios might, in addition, change with increased temperatures such  
57 as we might find seasonally, but also as might be forced by global atmospheric warming.  
58 Microbial processes such as denitrification are strongly affected by temperatures (Valiela

59 1995), and higher denitrification could result in notable fractionation of  $\delta^{15}\text{N}$ . This indirect  
60 linkage could furnish heavier N that is taken up by producers.

61 Some studies reported that  $\delta^{15}\text{N}$  and  $\delta^{13}\text{C}$  of producers showed seasonal variation  
62 (Riera and Richard, 1996; 1997; Fourqurean et al., 1997; Kang et al., 1999; Adin and Riera,  
63 2003; Machás et al., 2003; Riera and Hubas, 2003; Pruell et al., 2006), while others did not  
64 (McClelland and Valiela, 1998; Cole et al., 2004). Similarly, some studies showed variation  
65 in consumers (Goering et al., 1990; Riera and Richard, 1996; 1997; Buskey et al., 1999;  
66 Kang et al., 1999; Carman and Fry, 2002; Kibirige et al., 2002; Moens et al., 2002; Adin and  
67 Riera, 2003; Machás et al., 2003; Riera and Hubas, 2003; Vizzini and Mazzola, 2003, 2005),  
68 and others did not (Goering et al., 1990). Knowledge of seasonal variation in stable isotopic  
69 ratios is important as a reflection of biogeochemical and ecological processes, as well as in  
70 regard to sampling schedules and expected variability for applied monitoring schedules.

71 In this paper we examine the seasonal variation in N and C stable isotopic ratios of  
72 producers and consumers within the food web of the Mondego estuary, and compare the  
73 changes in ratios in organisms to the seasonal changes in temperature, precipitation,  
74 dissolved nutrients, and phytoplankton chlorophyll we measured in the Mondego ecosystem.  
75 This comparison aims to discern the degree to which seasonally varying driving factors  
76 might be manifest in the isotopic ratios of the food web, as well as identify the components  
77 of the food web that might be reasonably reliable indicators of changes in nutrient  
78 enrichment and warming.

79

## 80 **2. Methods**

### 81 *2.1. Study site*

82 The Mondego estuary is a relatively small (1600 ha), warm-temperate, polyhaline,  
83 intertidal system located on the Atlantic coast of Portugal, and consists of two arms, north and

84 south (Fig. 1). The southern arm is characterised by large areas of intertidal mudflats (almost  
85 75% of the area) exposed during low tide. The system receives agricultural runoff from  
86 15000 ha of upstream cultivated land (mainly rice fields) and supports a substantial  
87 population, industrial activities, salt-works, and aquaculture farms, and is also the location of  
88 the Figueira da Foz city harbour, which constitutes a tourism centre. All these activities have  
89 imposed a strong anthropogenic impact. A mixture of inputs from sewage effluent,  
90 agricultural runoff, as well as releases from maricultural activity contributes to the nutrient  
91 loads entering the Mondego estuary.

92 In the early 1990s the southern arm was almost silted up in the upstream areas,  
93 causing the river discharge to flow essentially through the northern arm. Consequently, the  
94 water circulation in the southern arm became mostly dependent on the tides and on the small  
95 freshwater input from a tributary, the Pranto River, artificially controlled by a sluice  
96 (Marques et al., 2003). In 1990-1992, the communication between the two arms of the estuary  
97 became totally interrupted in the upstream area due to the completion of stonewalls in the  
98 northern arm banks. Following this interruption, the ecological conditions in the southern arm  
99 suffered a rapid deterioration. The combined effect of an increased water residence time and  
100 of nutrient concentrations became major driving forces behind the occurrence of seasonal  
101 blooms of *Ulva* sp. and a concomitant severe reduction of the area occupied by *Zostera noltii*  
102 beds, previously the richest habitat in terms of productivity and biodiversity (Marques et al.,  
103 1997; 2003). The shift in benthic primary producers affected the structure and functioning of  
104 the biological communities, and through time such modifications started inducing the  
105 emergence of a new selected trophic structure, which has been analysed in abundant literature  
106 (e.g. Dolbeth et al., 2003; Cardoso et al., 2004 a; b; Patrício et al., 2004).

107 From 1998 to 2006 several interventions were carried out to ameliorate the condition of  
108 the system, namely by improving water circulation, which was followed by a partial recovery

109 of the area occupied by *Z. noltii* and the cessation of green *Ulva* sp. blooms (Lillebo et al.,  
110 2005; 2007).

111

## 112 2.2. Sample collection and preparation

113 To assess water quality of the Mondego waters, we collected water samples on a  
114 monthly basis at two sites (*Zostera* site and bare sediment site; Fig. 1), from November 2005  
115 to July 2006. In a companion paper (Baeta et al. submitted) we established that there were no  
116 differences in nutrients or chlorophyll concentrations in samples taken from the two sites,  
117 and so here we treat the samples as replicates. In each sample we measured concentrations of  
118 nitrate ( $\text{NO}_3^-$ ), ammonium ( $\text{NH}_4^+$ ), and phosphate ( $\text{PO}_4^{3-}$ ), and the concentration of  
119 chlorophyll *a*. Samples were immediately filtered (Whatman GF/F glass-fibre filter) and  
120 stored frozen at  $-18^\circ\text{C}$  until the analysis following standard methods described in  
121 Limnologisk Metodik (1992) for  $\text{NH}_4^+$  and  $\text{PO}_4^{3-}$ , and in Strickland and Parsons (1972) for  
122  $\text{NO}_3^-$ , and nitrite  $\text{NO}_2^-$ . The phytoplankton chlorophyll *a* determinations were performed by  
123 filtering 0.5-1.0 l of water through Whatman GF/F glass-fibre filters (Parsons et al., 1985). In  
124 the field and during transportation to the laboratory, samples were stored on ice and protected  
125 from light. Data on monthly precipitation and air temperature derive from the nearby city of  
126 Coimbra (Instituto de Meteorologia, Coimbra forecast station).

127 We measured stable isotopic values in components of the Mondego food web,  
128 including particulate organic matter (POM), sedimentary organic matter (SOM), meiofauna,  
129 seagrass, macroalgae, macrobenthos, zooplankton, and the fish in each of the two sites. To  
130 evaluate the seasonal variation in the isotopic values, we repeated the sampling in November  
131 (2005), and February, May, and July (2006) at the two sites in the south arm of the estuary.  
132 Water samples for POM were collected monthly.

133 POM was obtained by filtering 0.5- 1 l of seawater, from a depth of 0.5 m below the surface,  
134 onto precombusted (450°C, 4h) Whatman GF/F filters (0.45 µm pore size) with a low  
135 pressure vacuum pump. Sediment samples from the upper 1 cm were collected with an  
136 acrylic corer (31 mm of diameter), and analysed for the isotopic composition. For the  
137 meiofauna, sediment samples were collected, and the top 3 cm of each sediment core was  
138 then passed through 500 µm and 38 µm sieves. Meiofauna were examined from the 38 µm  
139 fraction, and samples for isotopic analysis were composites of 50 to 300 individuals.  
140 Seagrass leaves and roots, and macroalgae were collected by hand and gently cleaned of  
141 epiphytic material. Macroinvertebrates were also taken manually from each site, and held in  
142 filtered sea water for 24 h to allow their guts to clear.

143 Zooplankton was collected by towing a Bongo net (0.5 m diameter, 200 µm mesh  
144 size) against the current for 20 min. The zooplankton samples for isotope analysis were  
145 composites of 20 to 200 individuals. Resident (*Atherina boyeri*, *Pomatoschistus microps*,  
146 *Pomatoschistus minutus*, *Syngnathus abaster*, *Syngnathus acus*) and transient (*Dicentrarchus*  
147 *labrax*, *Solea solea*) fish species were collected using a 2 m beam trawl, with 5 mm stretched  
148 mesh size on the cod end. The trawls were carried out during the night, at low water during  
149 spring tides, but only at the bare sediment site. These mobile taxa (fish) could easily manage  
150 the short distance between the two sites (Fig. 1), so it was not considered worthwhile to  
151 collect samples at the two sites.

152 All samples were rinsed with Milli-Q water (filters with POM were rinsed with  
153 ammonium formate to removed the salts), and then freeze-dried. When dry, samples were  
154 ground (filters with POM were kept whole) into a homogenous powder using mortar and  
155 pestle, and combined to make single composite samples of each species/group per site per  
156 sampling date. Samples were then weighed, and loaded into tin capsules. Whole organisms  
157 were used in all cases except for bivalves and decapods, the shells of which were removed,

158 and for fish, only muscle of the dorsal region was analysed. No acidification was applied to  
159 the samples to avoid alterations in the isotopic signal (Mateo et al. 2008).

160 Samples were analysed using an EA-IRMS (Isoprime, Micromass, UK). Isotopic  
161 values were expressed in the  $\delta$  unit notation as deviations from standards (Vienna Pee Dee  
162 Belemnite for  $\delta^{13}\text{C}$  and  $\text{N}_2$  in air for  $\delta^{15}\text{N}$ ) following the formula:  $\delta^{13}\text{C}$  or  $\delta^{15}\text{N} = [(R_{\text{sample}}/$   
163  $R_{\text{standard}}) - 1] \times 10^3$ , where  $R$  is  $^{13}\text{C}/^{12}\text{C}$  or  $\delta^{15}\text{N}/^{14}\text{N}$ . The analytical precision for the  
164 measurement was 0.2‰ for both carbon and nitrogen.

165 The data were analyzed using ANOVA to test the null hypothesis that there were no  
166 significant differences in either the  $\delta^{15}\text{N}$  or  $\delta^{13}\text{C}$  composition of each group/species among  
167 seasons (autumn, winter, spring, summer).

168

### 169 **3. Results and discussion**

#### 170 *3.1. Seasonal ambient conditions*

171 Weather varied substantially during the sampling period (Fig. 2). Temperatures varied  
172 according to the season, with, on average, winter maxima around 15°C rising to 25°C in  
173 summer. In addition, during the period over which we sampled the estuary there were marked  
174 departures from average conditions. In particular, during our last sampling interval in July,  
175 very warm temperatures (near 40 °C) (Fig. 2, top) were brought about by a northern incursion  
176 of an African air mass. The sampling period was also one during which Portugal suffered a  
177 lengthy drought, relative to average long-term precipitation (Fig. 2, bottom).

178 There was a marked seasonality to conditions in the water column (Fig. 3). Nitrate  
179 concentrations were high during winter, and diminished about four-fold during the warmer  
180 months. Concentrations of  $\text{NH}_4^+$  and  $\text{PO}_4^{3-}$  were usually much lower than those of  $\text{NO}_3^-$  (Fig.  
181 3, top). N/P was generally above the 16:1 Redfield ratio during colder months and below it in  
182 warmer months. This suggests that throughout the winter months, P supply might have been

183 the limiting nutrient, while during the summer, N might limit producer growth (Fig. 3,  
184 middle). Chlorophyll *a* concentrations peaked in spring, perhaps drawing down nitrate  
185 concentration during warmer months (Fig. 3, bottom).

186

### 187 3.2. Seasonal changes in isotopic values

188 We collected 45 different taxa and analysed their stable carbon and nitrogen isotopic  
189 composition. These taxa included 5 primary producers, POM, sediment, 21  
190 macroinvertebrates species, 2 meiofauna groups, 8 fish species, and 7 zooplankton taxa  
191 (Table 1).

192 There was a consistent lack of seasonal pattern in the isotopic values in most  
193 compartments of the Mondego ecosystem (Figs. 4, 5). Thirty of the 37 compartments  
194 measured showed no evidence of seasonal changes (Table 2).

195 For  $\delta^{13}\text{C}$  only the copepod *Acartia tonsa* showed a seasonal variation, becoming less  
196 negative in spring and summer. Buskey et al. (1999) showed that *A. tonsa* living over  
197 seagrass beds obtain a larger proportion of their carbon from seagrass than do nearby  
198 populations living over muddy bottoms without seagrass. In our study, during the most  
199 intense periods of the phytoplankton bloom,  $\delta^{13}\text{C}$  values in *A. tonsa* became less negative,  
200 suggesting that more seagrass carbon might have been entering their diets. On the other hand,  
201 the much depleted carbon signatures for the most of the year could be due to a great  
202 contribution of terrestrial organic matter, since several studies have shown that terrestrial  
203 plants have the most depleted  $\delta^{13}\text{C}$  signatures, around  $-26\text{‰}$  (e.g. Vizzini and Mazzola,  
204 2003).

205 The  $\delta^{15}\text{N}$  of producers and consumers consistently lacked significant seasonal  
206 variation across most of the growing season (Table 2), and only on one date in July did it  
207 become significantly higher ( $p < 0.05$ ) in the  $\delta^{15}\text{N}$  of producers. The stable isotopic values of

208 C and N in these producers and consumers therefore showed no seasonal variation. In  
209 contrast, the  $\delta^{15}\text{N}$  of producers rapidly increased seemingly as a seasonal response to certain  
210 conditions.

211 High nitrogen isotopic signatures found in producers in July may have resulted from  
212 seasonal changes in biogeochemical processes, such as denitrification. Denitrification is  
213 temperature-dependent and takes place under anaerobic conditions. This process may lead to  
214 a loss of isotopically light  $^{14}\text{N}$ , which enriches the remaining DIN pool with  $^{15}\text{N}$ . During the  
215 unusually warm event in July there was a strong sulfidic smell, which suggested widespread  
216 anoxia that could have favoured high denitrification rates.

217 The enrichment of  $\delta^{15}\text{N}$  in the producers increased significantly with warmer  
218 temperatures (Fig. 6), as might be expected if a temperature-dependent process such as  
219 denitrification was indeed involved. It is not surprising to find that macrophyte isotopic  
220 values show seasonal changes in N supply: isotopic values of macrophyte fronds change in a  
221 matter of hours to a few days (Teichberg et al. 2007) since their internal nitrogen pools turn  
222 over rather quickly. The importance of the results of Figure 6 are that if indeed global  
223 atmospheric warming increases water temperatures in estuaries such as the Mondego, we can  
224 expect a gradually increase in  $\delta^{15}\text{N}$  in the producers.  $\delta^{15}\text{N}$  values then could therefore be  
225 thought of as indirect indicators of warming.

226 Nitrogen isotope ratios in two isopods also showed seasonal variation, increasing in  
227 July. Both *I. chelipes* and *L. levii* are grazers, feeding on macrophytes (Bamber, 2004), so the  
228 increased nitrogen isotopic ratio in summer could be due to the enrichment found in the  
229 producers during this period, since complete turnover of these populations could occur in a  
230 matter of days if all individuals were equally mobile (Shafir & Field, 1980). Accordingly, the  
231 lack of seasonal variation of  $\delta^{15}\text{N}$  enrichment in most consumers of the Mondego food web  
232 might be related to a slower turnover of internal N pools in consumers (weeks/months)

233 compared to pools in macrophytes (days), position of species in the food web and omnivores  
234 feeding behaviour, and also probably due to the fact that, excepting isopods, the other groups  
235 do not feed directly on fresh macrophytes. This suggests that consumer isotopic values  
236 constitute a more time-integrated reflection of nitrogen isotopic values, as reported by  
237 Vander Zanden et al. (1998).

238         The results of this study show that there were strong seasonal driving variables in the  
239 Mondego estuary. Increased temperature increased  $\delta^{15}\text{N}$  of important producers; this might  
240 be an indirect result of microbial N transformations and suggests that producer  $\delta^{15}\text{N}$  might  
241 become an indicator of larger climatic trends. The influence of this seasonal forcing, as  
242 manifest in stable isotopic ratios of consumers within the Mondego food web, was  
243 surprisingly modest, with most species showing no significant seasonal trends. These results  
244 suggest that the seasonal variation in the various factors we measured (temperature,  
245 precipitation, nutrients, chlorophyll) within the Mondego was not enough to change isotopic  
246 signatures in consumers. This is convenient for monitoring purposes, as it frees sampling  
247 protocols from seasonal schedules.

248

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259

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381 Table 1.  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  of primary producers and consumers collected from Mondego  
 382 estuary. Data are sample size ( $N$ ), and mean  $\delta$  values ( $\pm$  SE), from November 2005 to July  
 383 2006.

Group/species	Abbreviation	$\delta^{13}\text{C}$	$\delta^{15}\text{N}$	$N$
		Mean $\pm$ SE	Mean $\pm$ SE	
<b>Macrophytes</b>				
<i>Enteromorpha</i> sp.	E	-12.9 $\pm$ 0.4	12.7 $\pm$ 1.2	8
<i>Fucus</i> sp.	F	-16.3 $\pm$ 0.5	9.7 $\pm$ 0.4	8
<i>Gracilaria</i> sp.	G	-17.4 $\pm$ 0.7	9.7 $\pm$ 1.0	8
<i>Ulva</i> sp.	U	-11.4 $\pm$ 0.3	12.3 $\pm$ 0.9	8
<i>Zostera noltii</i> (leaves)	Zl	-12.5 $\pm$ 0.1	9.5 $\pm$ 1.5	8
<i>Zostera noltii</i> (roots)	Zr	-12.2 $\pm$ 0.2	11.3 $\pm$ 3.3	8
<b>Particulate organic matter</b>	POM	-22.8 $\pm$ 0.2	5.9 $\pm$ 0.2	18
<b>Sediment</b>	SOM	-21.31 $\pm$ 0.4	4.8 $\pm$ 0.4	8
<b>Amphipoda</b>				
<i>Amphithoe valida</i>	Av	-14.8 $\pm$ 0.3	10.9 $\pm$ 0.2	8
<i>Echinogammarus marinus</i>	Em	-14.7 $\pm$ 0.5	10.6 $\pm$ 0.3	8
<i>Melita palmata</i>	Mp	-15.5 $\pm$ 0.7	9.9 $\pm$ 0.2	8
<b>Bivalvia</b>				
<i>Cerastoderma edule</i>	Ce	-19.1 $\pm$ 0.2	7.8 $\pm$ 0.3	8
<i>Mytilus edulis</i>	Me	-19.2 $\pm$ 0.4	7.2 $\pm$ 0.3	8
<i>Scrobicularia plana</i>	Sp	-17.4 $\pm$ 0.5	9.0 $\pm$ 0.2	8
<b>Decapoda</b>				
<i>Carcinus maenas</i>	Cm	-16.4 $\pm$ 0.3	12.1 $\pm$ 0.2	8
<i>Crangon crangon</i>	Ccr	-15.3 $\pm$ 0.3	11.6 $\pm$ 0.1	8
<b>Gastropoda</b>				
<i>Gibbula umbilicalis</i>	Gu	-12.0 $\pm$ 1.5	10.7 $\pm$ 1.3	3
<i>Hydrobia ulvae</i>	Hu	-11.5 $\pm$ 0.3	9.6 $\pm$ 0.1	8
<i>Littorina litorea</i>	Ll	-11.2 $\pm$ 0.4	12.2 $\pm$ 0.1	4
<b>Isopoda</b>				
<i>Cyathura carinata</i>	Cc	-14.3 $\pm$ 0.4	11.7 $\pm$ 0.2	8
<i>Idotea chelipes</i>	Ic	-14.5 $\pm$ 0.5	9.9 $\pm$ 0.5	8
<i>Lekanesphaera levii</i>	Llev	-12.0 $\pm$ 0.4	8.7 $\pm$ 0.4	8
<b>Meiofauna</b>				
Nematoda	Ne	-16.3 $\pm$ 0.5	11.3 $\pm$ 0.3	8
Copepoda	Co	-16.3 $\pm$ 0.4	10.9 $\pm$ 0.3	8
<b>Polychaeta</b>				
<i>Alkmaria romijni</i>	Ar	-16.7 $\pm$ 0.7	10.7 $\pm$ 0.2	6
<i>Capitella capitata</i>	Cca	-16.4 $\pm$ 0.3	10.8 $\pm$ 0.3	7
<i>Glycera tridactyla</i>	Gt	-14.3 $\pm$ 0.2	13.5 $\pm$ 0.3	5
<i>Hediste diversicolor</i>	Hd	-14.2 $\pm$ 0.2	11.8 $\pm$ 0.3	7
<i>Heteromastus filiformis</i>	Hf	-16.2 $\pm$ 0.2	11.5 $\pm$ 0.2	7
<i>Nephtys cirrosa</i>	Nc	-15.6 $\pm$ 0.7	12.6 $\pm$ 0.3	6
<i>Streblospio shrubsolii</i>	Ssh	-16.9 $\pm$ 0.7	10.8 $\pm$ 0.4	5
<b>Fishes</b>				
<i>Atherina boyeri</i>	Ab	-18.0 $\pm$ 0.6	11.7 $\pm$ 0.7	2
<i>Dicentrarchus labrax</i> (juv)	Dl	-18.2 $\pm$ 0.6	14.2 $\pm$ 0.6	2
<i>Diplodus vulgaris</i> (juv)	Dv	-17.9 $\pm$ 0.5	15.0 $\pm$ 0.3	4
<i>Pomatoschistus microps</i> (juv)	Pm	-22.4 $\pm$ 0.3	14.0 $\pm$ 0.7	2
<i>P. minutus</i>	Pmi	-17.4 $\pm$ 0.4	13.1 $\pm$ 0.3	4
<i>Syngnathus abaster</i>	Sa	-15.0 $\pm$ 0.9	12.3 $\pm$ 0.6	2
<i>S. acus</i>	Sac	-16.2 $\pm$ 0.0	11.4 $\pm$ 0.9	2
<i>Solea solea</i> (juv)	Ss	-23.5 $\pm$ 0.4	14.8 $\pm$ 0.6	4
<b>Zooplankton</b>				
<i>Acartia tonsa</i>	At	-23.2 $\pm$ 0.9	9.7 $\pm$ 0.4	8
<i>Acartia</i> sp.	Asp	-18.5 $\pm$ 0.5	8.4 $\pm$ 0.2	8
Cladocera	Cl	-21.9 $\pm$ 1.0	9.4 $\pm$ 0.8	3
Mysidacea (juv)	My	-18.9 $\pm$ 0.7	8.1 $\pm$ 0.4	4
<i>Pomatochistos</i> sp. (larvae)	Pl	-20.0 $\pm$ 0.5	9.8 $\pm$ 0.3	3
<i>Sagitta friderici</i>	Sf	-18.2 $\pm$ 0.6	11.1 $\pm$ 0.3	6
Zoeae (brachyura)	Zo	-18.7 $\pm$ 0.4	9.0 $\pm$ 0.2	6

385 Table 2. ANOVA results testing seasonal differences for C and N isotope ratios of  
 386 groups/species collected in the Mondego estuary.\* =  $p < 0.05$ , \*\* =  $p < 0.01$ ; the absence of \*  
 387 means no significant differences.

Group/Species	$\delta^{13}\text{C}$		$\delta^{15}\text{N}$	
	F	df	F	df
<b>Macrophytes</b>				
<i>Enteromorpha</i> sp.	0.983	3	30.678**	3
<i>Fucus</i> sp.	0.822	3	3.990	3
<i>Gracilaria</i> sp.	1.051	3	2.801	3
<i>Ulva</i> sp.	3.615	3	49.339**	3
<i>Zostera noltii</i> (leaves)	2.004	3	32.239**	3
<i>Zostera noltii</i> (roots)	1.229	3	32.247**	3
<b>Particulate organic matter</b>	1.926	3	2.046	3
<b>Sediment</b>	5.412	3	3.294	3
<b>Amphipoda</b>				
<i>Amphithoe valida</i>	0.218	3	0.77	3
<i>Echinogammarus marinus</i>	0.648	3	1.178	3
<i>Melita palmata</i>	1.074	3	4.091	3
<b>Bivalvia</b>				
<i>Cerastoderma edule</i>	0.159	3	0.204	3
<i>Mytilus edulis</i>	0.797	3	0.899	3
<i>Scrobicularia plana</i>	1.167	3	1.856	3
<b>Decapoda</b>				
<i>Carcinus maenas</i>	0.397	3	0.526	3
<i>Crangon crangon</i>	0.315	3	0.561	3
<b>Gastropoda</b>				
<i>Hydrobia ulvae</i>	0.347	3	1.348	3
<b>Isopoda</b>				
<i>Cyatura carinata</i>	2.063	3	1.797	3
<i>Idotea chelipes</i>	0.532	3	11.204*	3
<i>Lekanesphaera levii</i>	0.116	3	10.327*	3
<b>Meiofauna</b>				
Nematoda	0.903	3	0.435	3
Copepoda	0.853	3	4.471	3
<b>Polychaeta</b>				
<i>Alkmaria romijni</i>	1.203	2	1.774	2
<i>Capitella capitata</i>	0.659	3	0.669	3
<i>Glycera tridactyla</i>	0.77	3	0.359	3
<i>Hediste diversicolor</i>	4.305	3	0.645	3
<i>Heteromastus filiformis</i>	0.131	3	4.557	3
<i>Nephtys cirrosa</i>	1.943	3	0.552	3
<i>Streblospio shrubsolii</i>	2.966	2	1.746	2
<b>Fishes</b>				
<i>Diplodus vulgaris</i> (juv)	7.428	3	4.261	3
<i>P. minutus</i>	4.668	3	3.875	3
<i>Solea solea</i> (juv)	3.939	3	7.186	3
<b>Zooplankton</b>				
<i>Acartia tonsa</i>	15.897*	3	3.628	3
<i>Acartia</i> sp. (marine species)	5.837	3	4.245	3
<i>Sagitta friderici</i>	8.585	2	3.613	2
Mysidacea	4.016	3	1.320	3
Zoea ( <i>C. maenas</i> )	0.787	2	14.653	2

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389

**Figure captions**

390 Fig. 1. Mondego estuary map showing sampling sites: *Zostera* and bare sediment sites (grey  
391 circles).

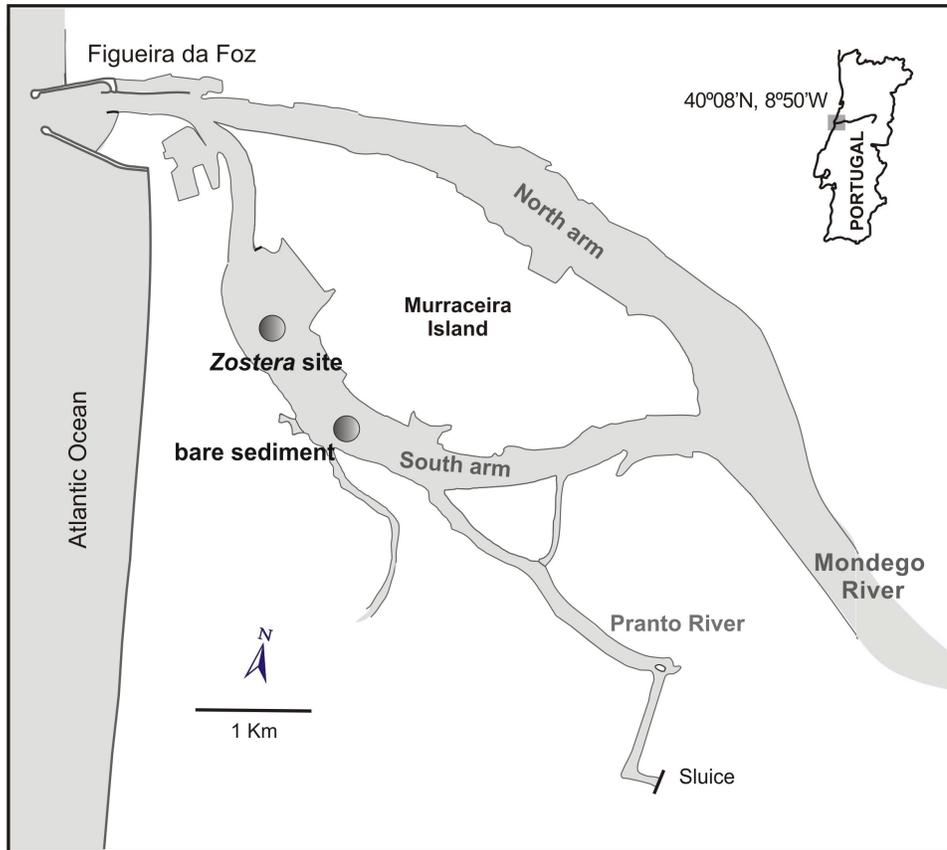
392 Fig. 2. Top: daily maximum and minimum air temperature, from November 2005 to July  
393 2006 (black lines), and daily maximum and minimum air temperature means for 1961-1990  
394 (grey lines). Black rectangles show when were the sampling periods. Bottom: precipitation  
395 values, from November 2005 to July 2006 (black bars), and precipitation means for 1961-  
396 1990 (grey bars).

397 Fig. 3. Mean of nitrate, ammonium, and phosphate (top), DIN/PO<sub>4</sub><sup>3-</sup> (middle), and  
398 phytoplankton chlorophyll *a* (bottom) concentrations, from the two sampling sites, *Zostera*  
399 and bare sediment sites, in the Mondego estuary, from November 2005 to July 2006.

400 Fig. 4. C stable-isotopic values (mean ± SE) for all the groups/species collected in the  
401 Mondego estuary, from November 2005 to July 2006. Abbreviations of species/groups are  
402 shown in table 1.

403 Fig. 5. N stable-isotopic values (mean ± SE) for all the groups/species collected in the  
404 Mondego estuary, from November 2005 to July 2006. Abbreviations of species/groups are  
405 shown in table 1.

406 Fig. 6. N stable-isotopic ratio for primary producers (seagrass and macroalgae) collected in  
407 the Mondego estuary. N stable-isotopic data are plotted against the air temperature.

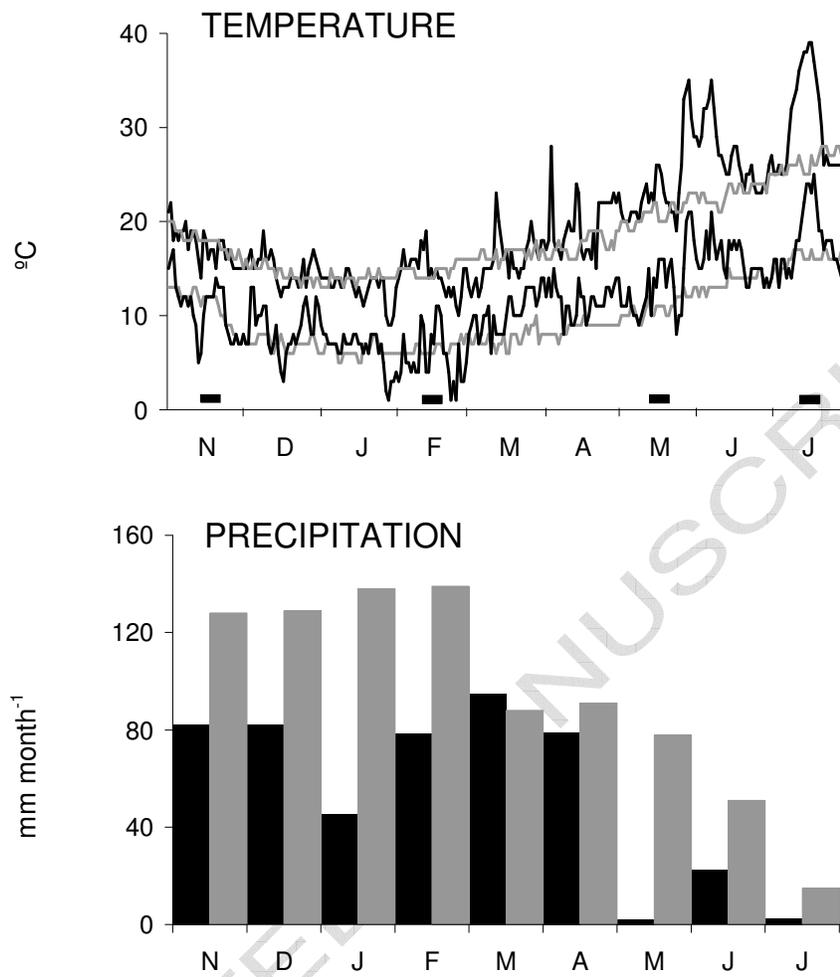


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Fig. 1

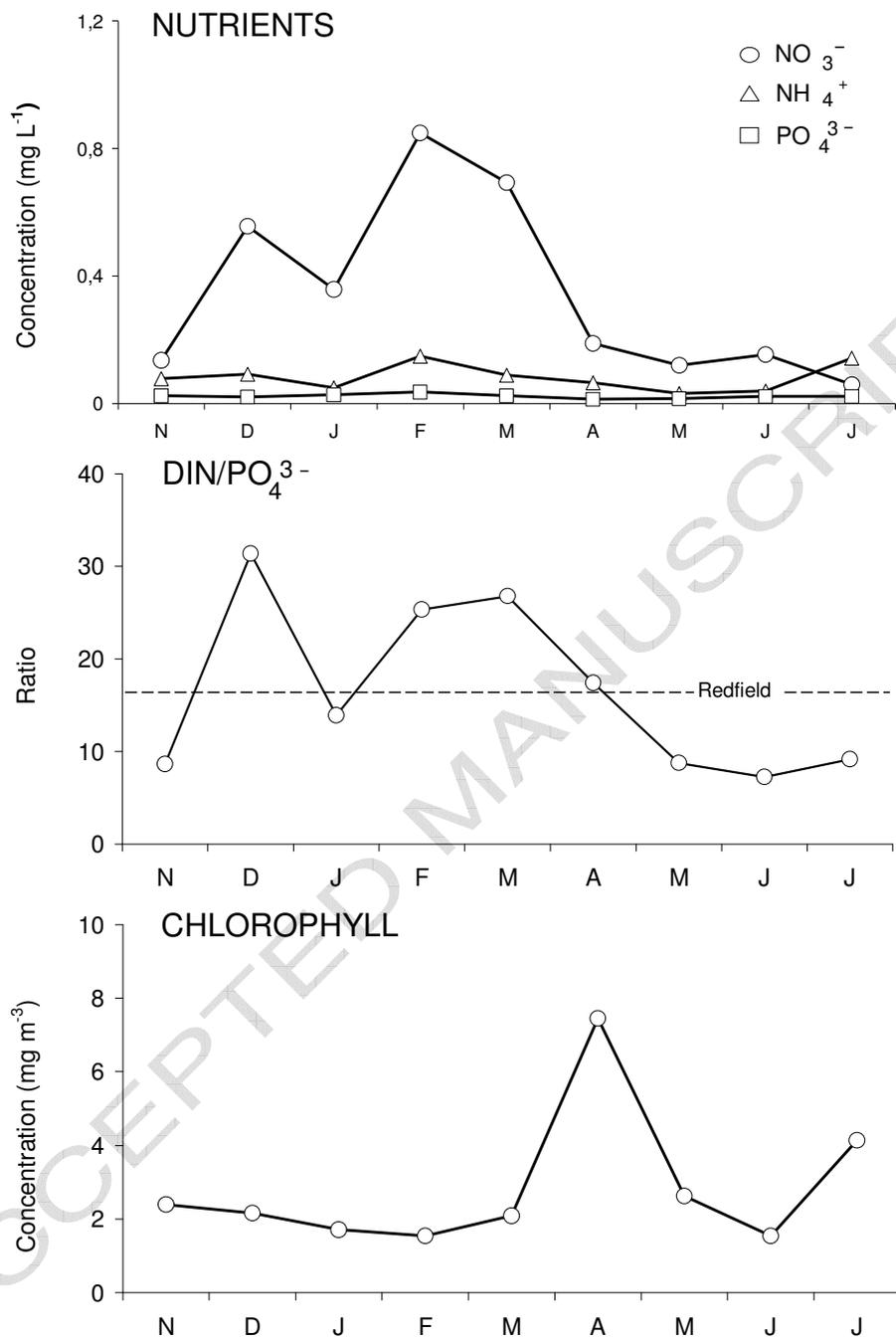
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Fig. 2

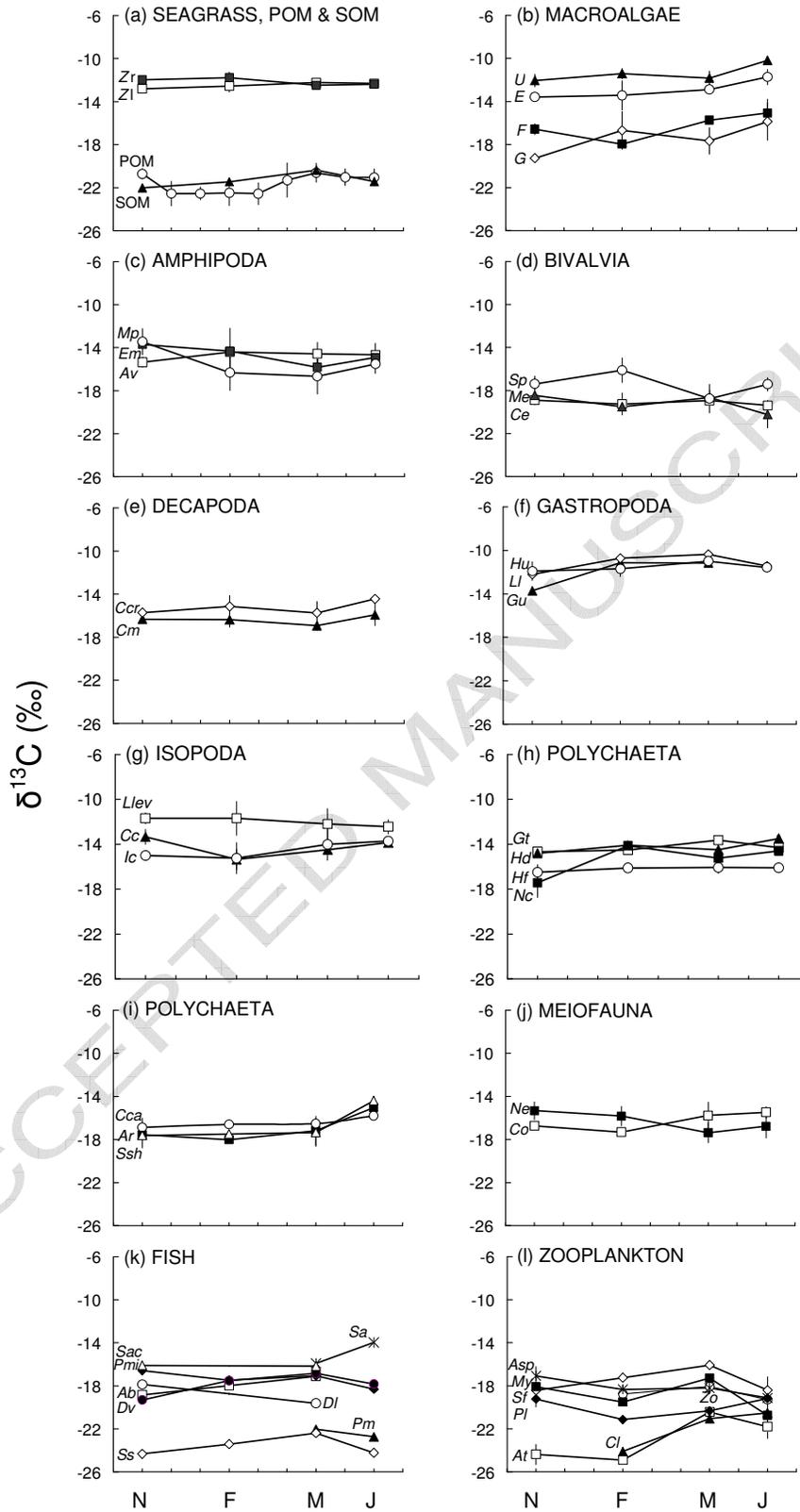
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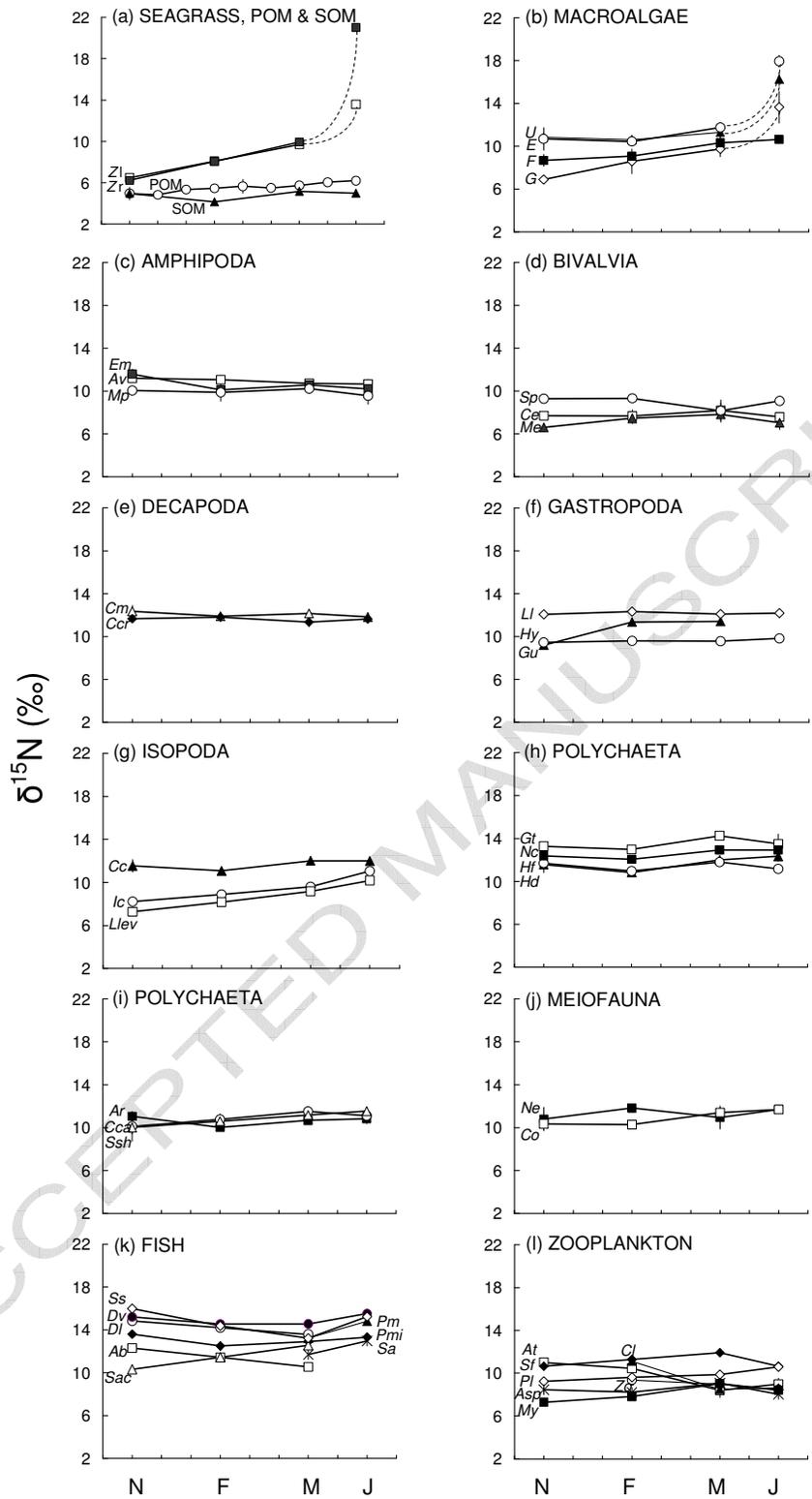
Fig. 3.

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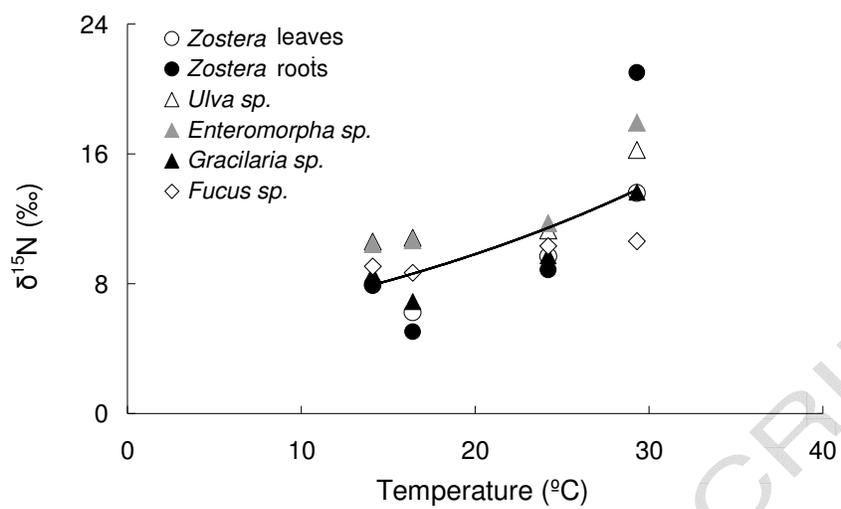


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Fig. 4



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Fig. 6

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