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Colonization of heavy metal polluted soils by Collembola: preliminary

experiments in compartmented boxes

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Abstract

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Two-week laboratory experiments were carried out in plastic boxes separated in two connected compartments filled with a neutral soil (mull humus) at pH 7.7 and an acid soil (moder humus) at pH 4.3, containing their original faunas. Migration of Collembola from one compartment to the other was allowed through a perforated wall. The mull was contaminated with three concentrations of lead (as lead acetate) at 50, 6,000 and 60,000 ppm. When combined with moder in the adjacent compartment, 6 times more individuals and 5 times more species were observed in the mull at the highest concentration applied, compared to mull combined with itself. Parisotoma notabilis proved to be highly sensitive to lead, and shifted to the moder compartment even at the lowest concentration. Densities of other mull species such as Pseudosinella alba were affected by medium to high concentrations of lead but these species did not move to the moder soil despite their high motility. Acidophilic species living in moder only, such as Willemia anophthalma, Xenylla tullbergi, Proisotoma minima and Xenylla tullbergi, colonized contaminated mull treatments with densities increasing with lead concentrations but the result of this process was erratic. Folsomia manolachei was present in both humus forms but was much more abundant in the moder. This species colonized mull at medium to high lead concentrations, where it restored totally or partly its original abundance in the uncontaminated mull. These results suggested differences between mull and moder populations of Folsomia manolachei.

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1. Introduction

The sensitivity of Collembola (Hexapoda) to heavy metals has been the subject of recent investigations, pointing on the heritability of resistance in populations submitted to long-term contamination of soils (Posthuma, 1990; Posthuma et al., 1993; Tranvik et al., 1994). Several studies have been carried out on collembolan communities, showing that shifts in species composition occurred under the influence of heavy metal contamination, together with a decrease in biodiversity (Hågyar and Abrahamsen, 1990; Filser and Hölscher, 1997; Bruus Pedersen et al., 1999).

The sensitivity of different species of Collembola to soil acidity has been recognized for a long time (Gisin, 1943; Mertens, 1975; Hågvar and Abrahamsen, 1984; Loranger et al., 2001). Experiments with artificial acid rain have shown that species living only in acid soils such as *Willemia anophthalma*, *Micranurida pygmaea* and *Mesaphorura yosii* were favoured by the application of sulphuric acid. Other species, present in acid soils but more abundant in neutral soils, such as *Parisotoma notabilis*, were eliminated by the same experimental conditions (Bååth et al., 1980; Hågvar and Kjøndal, 1981; Hågvar, 1984).

The presence of high concentrations of aluminum, iron and other metals in the soil solution when soils are very acid (Bergkvist, 1987; Lundström et al., 2000), and the increased ecological effects of heavy metals at low pH (Crommentuijn et al., 1997), led us to postulate that tolerance to heavy metals and tolerance to acidity could be ecologically related. Specifically, animals living in acid soils, either at the species or the sub-species level, could be more tolerant to heavy metals than animals commonly living in neutral or weakly acidic soils. The threshold of pH 5, below which changes in collembolan species composition have been recorded (Ponge, 1983; Ponge, 1993; Ponge, 2000a), is also a threshold under which aluminum passes into the soil solution as a metal cation and may exert severe toxicity, mostly by negatively interacting with phosphorus metabolism (Clarkson, 1969; James and Riha, 1984; Berggren et al., 1990). Given a common detoxication strategy in Collembola and other Hexapoda, i.e. by periodically renewing the midgut epithelium (Joosse and Buker, 1979; Hopkin,

1995), it can be hypothesized that species living in naturally acidic soils (and thus thought tolerant or resistant to free metallic forms) are able to colonize neutral or weakly acidic soils contaminated with heavy metals as a consequence of human activities such as mining, agriculture and industry.

It is intended to test this hypothesis in the long-term by inoculating acid humus profiles, with their complete acido-tolerant collembolan communities, into metal-contaminated sites. Prior to undertaking field experiments, a preliminary study was conducted in order to verify whether Collembola living in acid soils are able to leave their original habitat to colonize a neutral soil polluted with heavy metals.

2. Materials and Methods

Short-term experiments were conducted in experimental boxes where animals were free to move between compartments, one filled with an acid soil (moder humus), and another filled with a neutral soil (mull humus) treated with lead acetate at different concentrations. Two soils were chosen, a Dysmoder, characterized by the accumulation of organic matter in the form of small enchytraeid faeces, and an Eumull, characterized by a rapid mixing of organic matter to mineral matter through earthworm activity (Brêthes et al., 1995; Ponge, 1999). These soils contain very different collembolan communities (Ponge, 1993; Chagnon et al., 2000). Lead acetate was chosen as the metal salt in order to avoid possible toxic effects of the anion. Given the rapid disappearance of acetate in a biologically active soil (Verschueren, 2001), most lead becomes rapidly complexed by soil organic matter (Sarret et al., 1997; Balabane et al., 1999).

Topsoil horizons, with their complete original fauna, were collected in May 2000. The Eumull material (pH 7.7) was collected in a rendzina soil in the park adjacent to the laboratory (Brunoy, 20km south of Paris) under oak (*Quercus robur* L.) and hornbeam (*Carpinus betulus* L.), with a ground flora of ivy (*Hedera helix* L.) and dog's mercury (*Mercurialis perennis* L.). The sparse litter horizon was discarded, together with aerial parts of ground vegetation, and the top 10cm of the A horizon were collected over a square meter area. The Dysmoder material (pH 4.3) was collected over the same area in a pine stand (*Pinus sylvestris* L.) located in the Senart forest near the laboratory. The ground

cover was mainly bramble (*Rubus fruticosus* L.) and bracken [*Pteridium aquilinum* (L.) Kuhn]. The superficial litter (OL horizon) was discarded, and only the OF horizon (fragmented litter) and the thick OH horizon (humified litter, 10cm thick) were collected. In both cases the soil was thoroughly crumbled by hand, most woody subterranean parts of vegetation, earthworms and stones being discarded.

Experiments were carried out in polystyrene boxes (175x115mm, 65mm deep) which were divided in two compartments by a 2mm thick millboard (Isorel®) division, pierced with 4mm diameter holes at a rate of 30 holes per wall. The division allowed passage by fauna but prevented physical contact between the two soils. Eumull was treated with lead acetate/deionized water solutions at three dilution rates, giving rise to three concentrations in the soil at 50, 6,000 and 60,000 ppm, referred to as low, medium and high concentration, respectively. These concentrations were chosen on the basis of preliminary assays. Control boxes (Eumull or Dysmoder) received deionized water only. Different combinations were established with five replicates each:

15	E0/E0	Eumull versus Eumull, without any addition of lead in the two compartments
16	E1/E1	Eumull versus Eumull, with lead at 50 ppm in the two compartments
17	E2/E2	Eumull versus Eumull, with lead at 6,000 ppm in the two compartments
18	E3/E3	Eumull versus Eumull, with lead at 60,000 ppm in the two compartments
19	E0/E1	Eumull versus Eumull, with lead at 50 ppm in only one compartment (E0/E1 =
20		unpolluted Eumull compartment, E1/E0 = polluted Eumull compartment)
21	E0/E2	Eumull versus Eumull, with lead at 6,000 ppm in only one compartment (E0/E2 =
22		unpolluted Eumull compartment, E2/E0 = polluted Eumull compartment)
23	E0/E3	Eumull versus Eumull, with lead at 60,000 ppm in only one compartment (E0/E3 =
24		unpolluted Eumull compartment, E3/E0 = polluted Eumull compartment)
25	E0/D	Eumull versus Dysmoder, without any addition of lead acetate in the Eumull (E0/D =
26		Eumull compartment, D/E0 = Dysmoder compartment)
27	E1/D	Eumull versus Dysmoder, with lead at 50 ppm in the Eumull (E1/D = Eumull
28		compartment, D/E1 = Dysmoder compartment)
29	E2/D	Eumull versus Dysmoder, with lead at 6,000 ppm in the Eumull (E2/D = Eumull
30		compartment, D/E2 = Dysmoder compartment)

1	E3/D	Eumull versus Dysmoder, with lead at 60,000 ppm in the Eumull (E3/D = Eumull
2		compartment, D/E3 = Dysmoder compartment)
3	D/D	Dysmoder versus Dysmoder
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5	The e	xperimental boxes were incubated in the laboratory at 15°C in darkness during two
6	weeks. At the	end of the incubation period arthropods were extracted by the dry-funnel method. The
7	identification	of Collembola was done to the species level under a light microscope at 400x
8	magnification.	After extraction of the microarthropods, the water pH of the dried soil was measured
9	electrometrica	lly in a 1:3 (w:w) soil:water slurry.
10		
11	Data	(pH values, abundance of species or groups of species, number of species) were
12	analysed by o	one-way ANOVA, using boxes as replicates. For treatments where both compartments
13	were filled wit	h the same soil (E0/E0, E1/E1, E2/E2, E3/E3, D/D), average values between paired
14	compartments	were used for comparisons with other treatments. When necessary data were log-
15	transformed ir	order to ensure additivity of variances. Comparisons among means following ANOVA
16	were achieved	d a posteriori by the SNK procedure (Glantz, 1997). When homogeneity of variances
17	could not be	achieved through log-transformation of the data, non-parametric Mann-Whitney rank
18	tests were per	formed in place of ANOVA (Glantz, 1997).
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20	3. Results	
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22	3.1. Comparis	on between Eumull and Dysmoder populations
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24	A tota	of 37 collembolan species were found over the whole set of samples (Table 1). Among
25	these, 17 mos	t frequent species were selected for further analyses (Table 2).
26		
27	Total	abundance of Collembola and species composition of Eumull and Dysmoder soils used
28	for the experi	ment differed to a great extent, as well as pH values (Table 2). At the end of the
29	experiment Co	ollembola were twice more abundant in Dysmoder compared to Eumull (Table 2, Fig. 1,
30	compare E0/E	0 to D/D), and the number of species was significantly higher (Fig. 2). Some species

were common to both soils, viz. *Dicyrtoma fusca, Folsomia manolachei, Friesea truncata, Isotomiella minor, Lepidocyrtus lanuginosus, Megalothorax minimus, Paratullbergia callipygos* and *Sphaeridia pumilis*. These species were called ubiquitous species. Nevertheless, among this group, some species were much more abundant in one or other of the soils. For instance *Folsomia manolachei* was 6 times more abundant in Dysmoder than in Eumull (Fig. 3), while *Isotomiella minor* was twice more abundant in Eumull than in Dysmoder (Table 1). Among species living only in Eumull, called Eumull species, the most frequent ones were *Heteromurus nitidus, Parisotoma notabilis* and *Pseudosinella alba*. The most frequent species living only in Dysmoder, called Dysmoder species, were *Micranurida pygmaea, Proisotoma minima, Pseudosinella mauli, Sminthurinus signatus, Willemia anophthalma* and *Xenylla tullbergi.* It must be noticed that these features describe the two soil samples at the end of the experimental period. They refer only to one square meter (the area over which the soil was sampled) in each site and they include possible changes due to experimental conditions, thus they do not picture the original population.

The efficiency of the perforated wall for allowing animals to pass from a compartment to another can be assessed by observing that *Parisotoma notabilis*, absent from the Dysmoder soil used for the experiment, was retrieved in this soil when paired to Eumull at densities equal or even higher than those of Eumull (Fig. 4). Conversely, all Dysmoder species were retrieved in the Eumull compartment when these soils were paired (Fig. 2).

3.2. Experiments on Eumull alone

Eumull treatments where both compartments were given the same concentration of lead (E0/E0, E1/E1, E2/E2 and E3/E3) allowed us to quantify, in our experimental conditions, the short-term impact of lead contamination on collembolan communities. Although soil pH increased significantly at the highest treatment, the increase was small (from 7.7 to 7.8), thus no departure from neutrality was observed in polluted Eumull (Table 3). The total abundance of Collembola, as well as the number of species, decreased only at high concentration (Table 3, Figs. 1 and 2). The general trend of decreased abundance at high concentration only was apparent for the most abundant species, *Isotomiella minor*, which comprised 57% of the collembolan population in unpolluted Eumull.

Lepidocyrtus lanuginosus and Heteromurus nitidus exhibited also a decrease in abundance at high concentration only. Nevertheless some species did not follow this general trend. Pseudosinella alba and Sphaeridia pumilis decreased significantly in abundance at medium concentration and totally disappeared at high concentration. Folsomia manolachei (Fig. 3) decreased progressively and significantly in abundance at all three concentrations but never disappeared totally. It was the most abundant species at the end of the experiment at high concentration. Parisotoma notabilis (Fig. 4) decreased only at medium then nearly disappeared at high concentration. Paratullbergia callipygos and Friesea truncata displayed an increase from nil to medium concentration (significant in P. callipygos, not significant in F. truncata), then disappeared totally at high concentration.

Treatments with Eumull in both compartments but with only one of them polluted with lead allowed to test the hypothesis of a possible refuge effect of unpolluted zones within a polluted site. Increasing the lead concentration in the compartment adjacent to unpolluted Eumull did not affect abundance and species richness of collembolan communities in the unpolluted compartment (Table 4, Figs. 1 and 2). Effects observed in the polluted compartment were similar to those observed when both compartments were equally polluted, i.e. a decrease in abundance of species richness at high concentration only (Table 5, Figs. 1 and 2). At the species level, *Folsomia manolachei* exhibited a significant decrease in unpolluted Eumull at medium and high concentrations (Table 4, Fig. 3). In the adjacent (polluted) compartment, this species exhibited a strong significant decrease at high concentration only, contrary to the experiment with both compartments equally polluted, where it exhibited a significant decrease as soon as low concentration was applied (Table 3, Fig. 3). No change in pH was observed in the different treatments (Tables 4 and 5).

3.3. Experiments on Eumull paired with Dysmoder

Treatments with Eumull combined with Dysmoder allowed us to test the hypothesis that collembolan species living in acid soils were able to colonize neutral soils polluted with heavy metals. The collembolan community of Eumull exhibited some significant changes when freely communicating with Dysmoder (Table 6, Figs. 1 and 2). In the absence of lead, a significant decrease was observed in the abundance of ubiquitous species, although no significant change occurred in the total

abundance nor in the total number of species (compare E0/E0 with E0/D). At the species level, only *Pseudosinella alba* was affected by the presence of Dysmoder, its abundance decreasing by a factor of ten (Table 1). Other species did not react significantly.

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After lead application, the total population of Eumull increased significantly at low and medium concentrations (compare E1/D and E2/D to E0/D). This increase, not observed when Eumull was alone (Table 3, Fig. 1), was due to a strong increase in the abundance of ubiquitous species. At high concentration, the total abundance of Collembola decreased, but was six times higher than when Eumull was alone (Tables 3 and 6, Fig. 1). The improvement due to the presence of Dysmoder was even more pronounced if we take into account the number of species. In the presence of Dysmoder, no decrease in the number of species occurred at the high concentration of lead (Table 6, Fig. 2), although this number was divided by six when Eumull was used alone. The decrease in the abundance of Parisotoma notabilis observed at medium and high concentration when Eumull was used alone (Table 3) was smoothed, although still significant, in the presence of Dysmoder (Table 6, Fig. 4). The same phenomenon was observed, but in a more pronounced way, in Folsomia manolachei, an ubiquitous species. At medium concentration the abundance of this species in Eumull more than doubled under the influence of Dysmoder (Fig. 3, Tables 3 and 6). At high concentration its abundance increased sixfold compared to Eumull alone. The abundance of Isotomiella minor nearly doubled after lead application at low concentration (Table 6), then significantly decreased at high concentration but its abundance was three times that observed in the absence of Dysmoder (Table 3). The Lepidocyrtus lanuginosus population collapsed at high concentration of lead in the absence of Dysmoder, and its abundance remained unchanged below this theshold (Table 3). In the presence of Dysmoder L. lanuginosus exhibited a significant increase in abundance when lead concentration increased, reaching a maximum at medium concentration, then decreased at high concentration (Table 6) but remained 38 times more abundant than in the absence of Dysmoder (Table 3). A small although significant increase (around a tenth unit) in soil pH was observed at high concentration only (Table 6), of the same order as when Eumull was alone (Table 3), and the presence of Dysmoder in a paired compartment did not influence the pH of Eumull.

In the absence of lead a few animals belonging to the Dysmoder group of species (*Micranurida pygmaea*, *Pseudosinella mauli*, *Sminthurinus signatus*, *Willemia anophthalma*, *Xenylla tullbergi*) were found in the Eumull compartment, and their number increased with lead concentration (Table 6, Fig. 1). Nevertheless this trend was not significant, due to strong departures from a box to another (see standard error values on Table 6).

The Dysmoder compartment was affected by the presence of the Eumull compartment. In the absence of lead in the Eumull compartment, the number of ubiquitous species decreased significantly in adjacent Dysmoder, as well as the density of the Dysmoder species *Willemia anophthalma*, and the pH increased slightly but significantly (Table 7, Figs. 1 and 2). When lead was applied to Eumull, the number and the abundance of Eumull species increased with lead concentration in the Dysmoder compartment, their abundance reaching a level nearly twice that of unpolluted Eumull (Table 6). This was mostly due to *Parisotoma notabilis*, the abundance of which at medium and high concentration of lead reached in Dysmoder a level 2.8 and 2.5 times that of original Eumull (Fig. 4, Table 7). Other species increased their abundance in Dysmoder when lead was applied to Eumull, as for instance the ubiquitous *Isotomiella minor*, the abundance of which trebled when lead was at medium concentration, and the ubiquitous *Lepidocyrtus lanuginosus*, the abundance of which near trebled at high concentration (Table 7). The pH of Dysmoder increased with lead concentration (Table 7).

4. Discussion

Despite the absence of significant colonization of the polluted neutral soil by strongly acidophilic species such as *Willemia anophthalma*, which was yet well-represented in the acid soil used for the experiment, migration movements were clearly demonstrated between the polluted mull and the moder. In the presence of lead, moder acted as a refuge for the isotomid *Parisotoma notabilis*, coming from the polluted mull. This species is known to live both in acid and neutro-acidocline soils (Ponge 1993), although it proved to be highly sensitive to experimental acidification with sulphuric acid (Bååth et al., 1980; Hågvar and Kjøndal, 1981; Hågvar, 1984), preferred alkalinity in pH-preference tests (Van Straalen and Verhoef, 1997) and was favoured by ash or lime application (Abrahamsen et al., 1980; Hågvar and Abrahamsen, 1980; Vilkamaa and Huhta, 1986). It should be noticed that

1 P.notabilis, also known for its sensitivity to heavy metals (Tranvik et al., 1993; Bruus Pedersen et al.,

1999), was strongly depressed in our experiments at medium concentration of lead (6,000 ppm).

Contrary to moder, unpolluted mull did not act as a refuge for animals escaping contamination by

heavy metals. Several reasons could explain this unexpected phenomenon, such as too many

predators or a lack of food resources and habitats, but our experiment was not designed to explore

these factors.

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The colonization of the polluted mull by the isotomid Folsomia manolachei coming from the moder was clear. This species was present in both humus forms, but it was much more abundant in the moder, which acted as a source for the polluted mull. At medium lead concentration and only in the presence of moder the abundance of this species increased in polluted mull compartments to a level so high that it cannot be explained by another cause than a colonization by specimens coming from the moder. Although the appearance of typical acidophilic species in the polluted mull was not significant, their presence in lead-polluted mull compartments, even at high concentration (60,000 ppm), indicated that they were able to colonize soils polluted by heavy metals, but were poorly attracted to them. One possible reason for the lack of a more intense colonization of the polluted mull by the mycetophagous poduromorph Willemia anophthalma (Ponge, 1991; Ponge, 2000b) could be the scarcity of fungi in the mull at neutral pH used for the experiments, and a subsequent lack of attraction by fungal odour (Bengtsson et al., 1988; Sadaka-Laulan et al., 1998), while fungi where abundant in the acidic moder, as ascertained by visual inspection. Another reason could be the lack of motility of this species, which has particularly short legs compared to that of entomobryomorph (entomobryid and isotomid) species. Isotomid species such as Folsomia manolachei and Parisotoma notabilis have a less specialized feeding habit, their diet being mainly composed of humified organic matter (Ponge, 1991; Ponge, 2000b). They are also much more motile than poduromorphs, due to possessing longer legs and functional furcula (Hopkin, 1997), and these two species are indifferent to soil pH (Ponge, 1993). These attributes make them better able to move to other places when environmental conditions become unfavourable or when more space or food are available somewhere else.

If we compare populations of *Folsomia manolachei* from the acid and neutral soils used for our experiments, it appears that mull specimens were more sensitive to lead application than moder ones. The existence of populations tolerant or intolerant to heavy metals in *Folsomia manolachei* could be compared to the case of *Orchesella cincta* where heritability of tolerance to heavy metals has been demonstrated (Posthuma, 1990; Posthuma et al., 1992; Posthuma et al., 1993). In our experiments animals attracted to a soil contaminated by heavy metals did not come from a polluted site but from an acidic coniferous forest soil. The importance of tolerant strains of *Folsomia manolachei* for the possible reclamation of polluted sites should be stressed, given its wide occurrence in all types of soils (Rusek, 1989; Mateos and Selga, 1991; Ponge, 1993). As in our experiments with lead acetate, this species recolonized soil cores polluted with copper sulphate where it exhibited higher densities than in the unpolluted control (Filser and Hölscher, 1997), and was attracted to copper sulphate solutions in choice experiments (Filser et al., 2000). The same phenomenon was observed with soil polluted by copper-containing fungicides (Filser et al., 2000). In a naturally lead-contaminated site Hågvar and Abrahamsen (1990) classified *Folsomia quadrioculata* (a nearby species) as tolerant to lead, its abundance being not affected at concentrations as high as 5,000 ppm.

Changes in pH cannot explain the above mentioned movements of collembolan populations from a humus type to another. Although soil pH was significantly affected by the application of lead acetate, the rise observed at the end of the experiment in the neutral soil was very weak (at most 0.1 pH unit), thus it was not high enough to force some species to escape from polluted compartments. A possible attraction of acidophilic species by a decrease of pH in the mull soil can be disregarded for the same reason. We cannot rule out possible toxic (or attractive) effects of the acetate moiety, but the absence of acidification of the soil at the end of the experiment led us to conclude that such effects, if any, were probably temporary. Nevertheless, acetate toxicity could explain at least partly, together with an increase in osmotic pressure of the soil solution (Heungens and Van Daele, 1984), the collapse in collembolan abundance and diversity we observed at high concentration. Lead, acetate and osmotic pressure acted probably both directly and indirectly on soil collembolan communities. Indirect effects could be suspected to occur through changes in predatory pressure (Grelle et al., 2000) and quantity and quality of the microflora (Tranvik and Eijsackers, 1989; Hopkin, 1994) which may have occurred during the two weeks of the experiment.

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2	5. References
3	
4	Abrahamsen, G., Hovland, J., Hågvar, S., 1980. Effects of artificial acid rain and liming on soil
5	organisms and the decomposition of organic matter. In: Hutchinson, T.C., Havas, M. (Eds.),
6	Effects of acid precipitation on terrestrial ecosystems. Plenum Publishing Corporation, New
7	York, pp. 341-362.
8	
9	Bååth, E., Berg, B., Lohm, U., Lundgren, B., Lundkvist, H., Rosswall, T., Söderström, B., Wiren, A.,
10	1980. Effects of experimental acidification and liming on soil organisms and decomposition in
11	a Scots pine forest. Pedobiologia 20, 85-100.
12	
13	Balabane, M., Faivre, D., Van Oort, F., Dahmani-Muller, H., 1999. Mutual effects of soil organic matter
14	dynamics and heavy metals fate in a metallophyte grassland. Environmental Pollution 105, 45-
15	54.
16	
17	Bengtsson, G., Erlandsson, A., Rundgren, S., 1988. Fungal odour attracts soil Collembola. Soil
18	Biolpgy and Biochemistry 20, 25-30.
19	
20	Berggren, D., Bergkvist, B., Falkengren-Grerup, U., Folkeson, L., Tyler, G., 1990. Metal solubility and
21	pathways in acidified forest ecosystems of South Sweden. The Science of the Total
22	Environment 96, 103-114.
23	
24	Bergkvist, B., 1987. Soil solution chemistry and metal budgets of spruce forest ecosystems in S.
25	Sweden. Water, Air, and Soil Pollution 33, 131-154.
26	
27	Brêthes, A., Brun, J.J., Jabiol, B., Ponge, J.F., Brun, J.J., 1995. Classification of forest humus forms: a
28	French proposal. Annales des Sciences Forestières 52, 535-546.

1	Bruus Pedersen, M., Axelsen, J.A., Strandberg, B., Jensen, J., Attrill, M.J., 1999. The impact of a
2	copper gradient on a microarthropod field community. Ecotoxicology 8, 467-483.
3	
4	Chagnon, M., Hébert, C., Paré, D., 2000. Community structures of Collembola in sugar maple forests:
5	relations to humus type and seasonal trends. Pedobiologia 44, 148-174.
6	
7	Clarkson, D.T., 1969. Metabolic aspects of aluminium toxicity and some possible mechanisms for
8	resistance. In: Rorison, I.H. (Ed.), Ecological aspects of the mineral nutrition of plants.
9	Blackwell Scientific Publications, Oxford, pp. 381-397.
10	
11	Crommentuijn, T., Doornekamp, A., Van Gestel, C.A.M., 1997. Bioavailability and ecological effects of
12	cadmium on Folsomia candida (Willem) in an artificial soil substrate as influenced by pH and
13	organic matter. Applied Soil Ecology 5, 261-271.
14	
15	Filser, J., Hölscher, G., 1997. Experimental studies on the reactions of Collembola to copper
16	contamination. Pedobiologia 41, 173-178.
17	
18	Filser, J., Wittmann, R., Lang, A., 2000. Response types in Collembola towards copper in the
19	microenvironment. Environmental Pollution 107, 71-78.
20	
21	Gisin, H., 1943. Ökologie und Lebengemeinschaften der Collembolen im schweizerischen
22	Exkursionsgebiet Basels. Revue Suisse de Zoologie 50, 131-224.
23	
24	Glantz, S.A., 1997. Primer of biostatistics, 4 th ed. McGraw-Hill, New York, p. 473.
25	
26	Grelle, C., Fabre, M.C., Leprêtre, A., Descamps, 2000. Myriapod and isopod communities in soils
27	contaminated by heavy metals in northern France. European Journal of Soil Science 51, 425-
28	433.

1	Hågvar, S., 1984. Effect of liming and artificial acid rain on Collembola and Protura in coniferous
2	forest. Pedobiologia 27, 341-254
3	
4	Hågvar, S., Abrahamsen, G., 1980. Colonisation by Enchytraeidae, Collembola and Acari in sterile soil
5	samples with adjusted pH levels. Oikos 34, 245-258.
6	
7	Hågvar, S., Abrahamsen, G., 1984. Collembola in Norwegian coniferous forest soils. III. Relations to
8	soil chemistry. Pedobiologia 27, 331-339.
9	
10	Hågvar, S., Abrahamsen, G., 1990. Microarthropods and Enchytraeidae (Oligochaeta) in naturally
11	lead-contaminated soil: a gradient study. Environmental Entomology 19, 1263-1277.
12	
13	Hågvar, S., Kjøndal, B.R., 1981. Effects of artificial acid rain on the microarthropod fauna in
14	decomposing birch leaves. Pedobiologia 22, 409-422.
15	
16	Harasymek, L., Sinha, R.N., 1974. Survival of springtails Hypogastrura tullbergi and Proisotoma
17	minuta on fungal and bacterial diets. Environmental Entomology 3, 965-968.
18	
19	Heungens, A., Van Daele, E., 1984. The influence of some acids, bases and salts on the mite and
20	Collembola population of a pine litter substrate. Pedobiologia 27, 299-311.
21	
22	Hopkin, S.P., 1994. Effects of metal pollutants on decomposition processes in terrestrial ecosystems
23	with special reference to fungivorous soil arthropods. In: Ross, S.M. (Ed.), Toxic metals in soil-
24	plant systems. John Wiley, Chichester, pp. 303-326.
25	
26	Hopkin, S.P., 1995. Deficiency and excess of essential and non-essential metals in terrestrial insects.
27	In: Harrington, R., Stork, N.E. (Eds.), Insects in a changing environment. Symposia of the
28	Royal Entomological Society of London 17, 251-270.
29	
30	Hopkin, S.P., 1997. Biology of the springtails. Oxford University Press, Oxford, p. 330.

1	
2	James, B.R., Riha, S.J., 1984. Soluble aluminum in acidified organic horizons of forest soils. Canadian
3	Journal of Soil Science 64, 637-646.
4	
5	Joosse, E.N.G., Buker, J.B., 1979. Uptake and excretion of lead by litter-dwelling Collembola.
6	Environmental Pollution 18, 235-240.
7	
8	Loranger, G., Bandyopadhyaya, I., Razaka, B., Ponge, J.F., 2001. Does soil acidity explain altitudinal
9	sequences in collembolan communities? Soil Biology and Biochemistry 33, 381-393.
10	
11	Lundström, U.S., Van Breemen, N., Bain, D.C., Van Hees, P.A.W., Giesler, R., Gustafsson, J.P.,
12	Ilvesniemi, H., Karltun, E., Melkerud, P.A., Olsson, M., Riise, G., Wahlberg, O., Bergelin, A.,
13	Bishop, K., Finlay, R., Jongmans, A.G., Magnusson, T., Mannerkoski, H., Nordgren, A.,
14	Nyberg, L., Starr, M., Tau Strand, L., 2000. Advances in understanding the podzolization
15	process resulting from a multidisciplinary study of three coniferous forest soils in the Nordic
16	Countries. Geoderma 94, 335-353.
17	
18	Mateos, E., Selga, D., 1991. Efecto de los incendios forestales sobre las poblaciones de colémbolos
19	edáficos en bosque mediterráneo. Revue d'Écologie et de Biologie du Sol 28, 19-30.
20	
21	Mertens, J., 1975. L'influence du facteur pH sur le comportement de Orchesella villosa (Geoffroy,
22	1764) (Collembola, Insecta). Annales de la Société Royale de Zoologie de Belgique 105, 45-
23	52.
24	
25	Ponge, J.F., 1983. Les Collemboles, indicateurs du type d'humus en milieu forestier. Résultats
26	obtenus au Sud de Paris. Acta Oecologica, Oecologia Generalis 4, 359-374.
27	
28	Ponge, J.F., 1991. Food resources and diets of soil animals in a small area of Scots pine litter.
29	Geoderma 49, 33-62.
30	

1	Ponge, J.F., 1993. Biocenoses of Collembola in atlantic temperate grass-woodland ecosystems.
2	Pedobiologia 37, 223-244.
3	
4	Ponge, J.F., 1999. Horizons and humus forms in beech forests of the Belgian Ardennes. Soil Science
5	Society of America Journal 63, 1888-1901.
6	
7	Ponge, J.F., 2000a. Acidophilic Collembola: living fossils? Contributions from the Biological
8	Laboratory, Kyoto University 29, 65-74.
9	
10	Ponge, J.F., 2000b. Vertical distribution of Collembola (Hexapoda) and their food resources in organic
11	horizons of beech forests. Biology and Fertility of Soils 32, 508-522.
12	
13	Posthuma, L., 1990. Genetic differentiation between populations of Orchesella cincta (Collembola)
14	from heavy metal contaminated sites. Journal of Applied Ecology 27, 609-622.
15	
16	Posthuma, L., Hogervorst, R.F., Joosse, E.N.G., Van Straalen, N.M., 1993. Genetic variation and
17	covariation for characteristics associated with cadmium tolerance in natural populations of the
18	springtail Orchesella cincta (L.). Evolution 47, 619-631.
19	
20	Posthuma, L., Hogervorst, R.F., Van Straalen, N.M., 1992. Adaptation to soil pollution by cadmium
21	excretion in natural populations of Orchesella cincta (L.) (Collembola). Archives of
22	Environmental Contamination and Toxicology 22, 146-156.
23	
24	Rusek, J., 1989. Collembola and Protura in a meadow-forest ecotone. In: Dallai, R. (Ed.), Third
25	international seminar on Apterygota. University of Siena, Siena, pp. 413-418.
26	
27	Sadaka-Laulan, N., Ponge, J.F., Roquebert, M.F., Bury, E., Boumezzough, A., 1998. Feeding
28	preferences of the collembolan Onychiurus sinensis for fungi colonizing holm oak litter
29	(Quercus rotundifolia Lam.). European Journal of Soil Biology 34, 179-188.
30	

1	Sarret, G., Manceau, A., Hazemann, J.L., Gomez, A., Mench, M., 1997. EXAFS study of the nature of
2	zinc complexation sites in humic substances as a function of Zn concentration. Journal de
3	Physique, Section IV 7, 799-802.
4	
5	Tranvik, L., Bengtsson, G., Rundgren, S., 1993. Relative abundance and resistance traits of two
6	Collembola species under metal stress. Journal of Applied Ecology 30, 43-52.
7	
8	Tranvik, L., Eijsackers, H., 1989. On the advantage of Folsomia fimetarioides over Isotomiella minor
9	(Collembola) in a metal polluted soil. Oecologia 80, 195-200.
10	
11	Tranvik, L., Sjögren, M., Bengtsson, G., 1994. Allozyme polymorphism and protein profile in
12	Orchesella bifasciata (Collembola): indicative of extended metal pollution? Biochemical
13	Systematics and Ecology 22, 13-23.
14	
15	Van Straalen, N.M., Verhoef, H.A., 1997. The development of a bioindicator system for soil acidity
16	based on arthropod pH preferences. Journal of Applied Ecology 34, 217-232.
17	
18	Verschueren, K., 2001. Handbook of environmental data on organic chemicals, 4 th ed. John Wiley and
19	Sons, New York.
20	
21	Vilkamaa, P., Huhta, V., 1986. Effects of fertilization and pH on communities of Collembola in pine
22	forest soil. Annali Zoologici Fennici 23, 167-174.
23	

1	6. Figure captions
2	
3	Fig. 1 Abundance of Collembola (mean and standard error of five replicates) in experimental
4	compartments (1 dm ³ each). E = Eumull. D = Dysmoder. E0 = Eumull without lead. E1 =
5	Eumull with 50 ppm lead. E2 = Eumull with 6,000 ppm lead. E3 = Eumull with 60,000 ppm
6	lead. Opposite bars relate to adjacent compartments
7	
8	Fig. 2 Number of species of Collembola (mean and standard error of five replicates) in experimental
9	compartments (1 dm ³ each). E = Eumull. D = Dysmoder. E0 = Eumull without lead. E1 =
10	Eumull with 50 ppm lead. E2 = Eumull with 6,000 ppm lead. E3 = Eumull with 60,000 ppm
11	lead. Opposite bars relate to adjacent compartments
12	
13	Fig. 3 Abundance of Folsomia manolachei (mean and standard error of five replicates) in experimental
14	compartments (1 dm³ each). E = Eumull. D = Dysmoder. E0 = Eumull without lead. E1 =
15	Eumull with 50 ppm lead. E2 = Eumull with 6,000 ppm lead. E3 = Eumull with 60,000 ppm
16	lead. Opposite bars relate to adjacent compartments
17	
18	Fig. 4 Abundance of Parisotoma notabilis (mean and standard error of five replicates) in experimental
19	compartments (1 dm ³ each). E = Eumull. D = Dysmoder. E0 = Eumull without lead. E1 =
20	Eumull with 50 ppm lead. E2 = Eumull with 6,000 ppm lead. E3 = Eumull with 60,000 ppm
21	lead. Opposite bars relate to adjacent compartments

Table 1. List of collembolan species found in the experiment, 0: absence, 1: presence

	Eumull	Dysmoder	Ubiquitous	Dubious
Allacma fusca (Linné, 1758)	1	0	0	0
Arrhopalites caecus (Tullberg, 1871)	1	0	0	0
Arrhopalites sericus Gisin, 1947	0	1	0	0
Dicyrtoma fusca (Lucas, 1842)	0	0	1	0
Dicyrtomina minuta (Fabricius, 1783)	1	0	0	0
Folsomia manolachei Bagnall 1939	0	0	1	0
Friesea truncata Cassagnau, 1958	0	0	1	0
Heteromurus nitidus (Templeton, 1835)	1	0	0	0
Isotomiella minor (Schäffer, 1896)	0	0	1	0
Isotomurus palustris (Müller, 1776)	1	0	0	0
Kalaphorura burmeisteri (Lubbock, 1873)	1	0	0	0
Lepidocyrtus lanuginosus (Gmelin, 1788)	0	0	1	0
Megalothorax minimus (Willem, 1900)	0	0	1	0
Mesaphorura gr krausbaueri (Börner, 190	0	0	1	0
Mesaphorura gr sylvatica (Rusek, 1971)	1	0	0	0
Mesaphorura gr yosii (Rusek, 1967)	0	0	1	0
Micranurida pygmaea (Börner, 1901)	0	1	0	0
Monobella grassei (Denis, 1923)	1	0	0	0
Neanura muscorum (Templeton, 1835)	1	0	0	0
Onychiurus pseudogranulosus Gisin, 1951	1	0	0	0
Orchesella cincta (Linné, 1758)	1	0	0	0
Paratullbergia callipygos (Börner, 1902)	0	0	1	0
Parisotoma notabilis (Schäffer, 1896)	1	0	0	0
Proisotoma minima (Absolon, 1901)	0	1	0	0
Pseudachorutes parvulus Böener, 1901	1	0	0	0
Pseudosinella alba (Packard, 1873)	1	0	0	0
Pseudosinella mauli Stomp, 1972	0	1	0	0
Sminthurinus aureus (Lubbock, 1862)	1	0	0	0
Sminthurinus signatus (Krausbauer, 1898)	0	1	0	0
Sphaeridia pumilis (Krausbauer, 1898)	0	0	1	0
Stenaphorurella denisi (Bagnall, 1935)	1	0	0	0
Tomocerus minor (Lubbock, 1862)	0	0	1	0
Willemia anophthalma Börner, 1901	0	1	0	0
Willemia buddenbrocki Hüther, 1959	1	0	0	0
Xenylla grisea Axelson, 1900	1	0	0	0
Xenylla tullbergi Börner, 1903	0	1	0	0
Xenyllodes armatus (Axelson, 1903)	0	0	0	1

Table 2. Mean abundance (±S.E.) of most frequent collembolan species and groups of species and water pH in Eumull and Dysmoder at the end of the experiment. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, D = Dysmoder species, U = ubiquitous species

		E	December
_		Eumull	Dysmoder
U	Dicyrtoma fusca	1.8±0.6	0.6±0.2
U	Folsomia manolachei	18.5±1.3b	113.4±18.6a
U	Friesea truncata	0.5±0.3	0.6±0.1
Ε	Heteromurus nitidus	1.3±0.3a	0b
U	Isotomiella minor	51.3±3.8a	25.7±3.4b
U	Lepidocyrtus lanuginosus	1.6±0.3b	4.8±0.6a
U	Megalothorax minimus	0.5±0.5b	8.9±0.6a
D	Micranurida pygmaea	0b	1.9±0.6a
U	Paratullbergia callipygos	0.6±0.3b	3.1±0.5a
Ε	Parisotoma notabilis	8.7±1.3a	0b
D	Proisotoma minima	0b	0.4±0.1a
Ε	Pseudosinella alba	2.0±0.4a	0b
D	Pseudosinella mauli	0b	0.7±0.3a
D	Sminthurinus signatus	0b	1.4±0.3a
U	Sphaeridia pumilis	2.1±0.7	2.9±0.5
D	Willemia anophthalma	0b	28.6±3.0a
D	Xenylla tullbergi	0b	0.3±0.1a
	Total abundance	90.4±5.5b	194.6±17.2a
	Abundance of Eumull species	13.3±1.6a	0b
	Abundance of Dysmoder spec	0b	33.3±2.5a
	Abundance of ubiquitous spec	77.1±4.4b	161.3±16.4a
	Total number of species	8.9±0.4b	12.0±0.3a
	Number of Eumuli species	4.8±0.5a	0b
	Number of Dysmoder species	0 b	5.2±0.4a
	Number of ubiquitous species		9.6±0.5a
	Water pH	7.66±0.03a	4.29±0.01b

Table 3. Mean abundance (\pm S.E.) of most frequent collembolan species and groups of species and water pH in Eumull after application of lead acetate in the two compartments of experimental boxes, compared to control boxes without addition of lead. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, U = ubiquitous species. E0/E0 = No Pb, E1/E1 = Pb at 50 ppm, E2/E2 = Pb at 6,000 ppm, E3/E3 = Pb at 60,000 ppm

		E0/E0	E1/E1	E2/E2	E3/E3
U	Dicyrtoma fusca	1.8±0.6	1.2±0.3	2.0±0.7	0.5±0.2
U	Folsomia manolachei	18.5±1.3a	13.3±1.8b	8.1±1.0c	0.9±0.7d
U	Friesea truncata	0.5±0.3ab	1.5±0.5a	1.6±0.4a	0b
Ε	Heteromurus nitidus	1.3±0.3a	0.6±0.2ab	1.0±0.2ab	0.3±0.2b
U	Isotomiella minor	51.3±3.8a	53.7±10.6a	46.8±7.9a	0.7±0.7b
U	Lepidocyrtus lanuginosus	1.6±0.3a	1.3±0.4a	1.2±0.3a	0.2±0.1b
U	Megalothorax minimus	0.5±0.5	0.1±0.1	0	0
U	Paratullbergia callipygos	0.6±0.3bc	1.0±0.2b	1.9±0.4a	0c
Ε	Parisotoma notabilis	8.7±1.3a	5.8±0.3a	2.3±0.3b	0.2±0.2c
Ε	Pseudosinella alba	2.0±0.4a	1.3±0.3ab	0.6±0.3bc	0c
U	Sphaeridia pumilis	2.1±0.7a	1.3±0.5ab	0.3±0.3b	0b
	Total abundance	90.4±5.5a	86.3±14.5a	68.3±9.6a	2.8±1.5b
	Abundance of Eumull species	13.3±1.6a	12.7±1.3a	6.4±0.7b	0.5±0.2c
	Abundance of ubiquitous spec	77.1±4.4a	73.6±13.7a	61.9±9.0a	2.3±1.3b
	Total number of species	8.9±0.4a	10.3±0.9a	9.1±0.3a	1.5±0.5b
	Number of Eumull species	4.8±0.5a	6.4±0.7a	5.8±1.0a	0.6±0.2b
	Number of ubiquitous species	6.4±0.7a	6.8±0.7a	6.0±0.0a	2.0±0.6b
	Water pH	7.66±0.03b	7.63±0.03b	7.65±0.01b	7.8±0.02a

Table 4. Mean abundance (±S.E.) of most frequent collembolan species and groups of species and water pH in unpolluted Eumull after application of lead acetate in the adjacent Eumull compartment, compared to control boxes without addition of lead. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, U = ubiquitous species. E0/E0 = No Pb, E0/E1 = Pb at 50 ppm, E0/E2 = Pb at 6,000 ppm, E0/E3 = Pb at 60,000 ppm

		E0/E0	E0/E1	E0/E2	E0/E3
U	Dicyrtoma fusca	1.8±0.6	1.4±0.4	1.8±0.4	0.8±0.6
U	Folsomia manolachei	18.5±1.3a	16.8±1.9ab	12.4±0.5bc	9.6±2.1c
U	Friesea truncata	0.5±0.3	1.0±0.8	0.8 ± 0.4	0.8±0.4
Ε	Heteromurus nitidus	1.3±0.3a	0.4±0.2b	0.6±0.2ab	0b
U	Isotomiella minor	51.3±3.8	59.2±10.5	51.2±10.2	33.2±7.7
U	Lepidocyrtus lanuginosus	1.6±0.3	1.4±0.5	1.6±0.4	1.4±0.7
U	Megalothorax minimus	0.5±0.5ab	1.0±0.3a	0b	0b
U	Paratullbergia callipygos	0.6±0.3	0.4±0.2	1.6±0.7	1.6±0.6
Ε	Parisotoma notabilis	8.7±1.3	7.8±0.7	8.8±2.7	3.8±0.9
Ε	Pseudosinella alba	2.0±0.4	1.8±0.6	1.4±0.5	0.6±0.4
U	Sphaeridia pumilis	2.1±0.7	1.4±0.5	1.6±0.9	0.8±0.2
	Total abundance	90.4±5.5	94.0±13.8	84.6±8.9	55.2±8.5
	Abundance of Eumull species	13.3±1.6	11.2±0.9	13.6±2.7	6.8±2.2
	Abundance of ubiquitous spec	77.1±4.4	82.8±13.1	71.0±9.7	48.4±8.6
	Total number of species	8.9±0.4	9.4±0.9	10.8±0.7	8.6±0.9
	Number of Eumull species	4.8±0.5	3.2±0.5	4.8±0.6	3.2±0.9
	Number of ubiquitous species	6.4±0.7	6.2±0.6	6.0±0.4	5.4±0.6
	Water pH	7.66±0.03	7.68±0.01	7.66±0.02	7.78±0.02

Table 5. Mean abundance (\pm S.E.) of most frequent collembolan species and groups of species and water pH in lead acetate-polluted Eumull adjacent to unpolluted Eumull, compared to control boxes without any addition of lead in both compartments. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, U = ubiquitous species. E0/E0 = No Pb, E1/E0 = Pb at 50 ppm, E2/E0 = Pb at 6,000 ppm, E3/E0 = Pb at 60,000 ppm

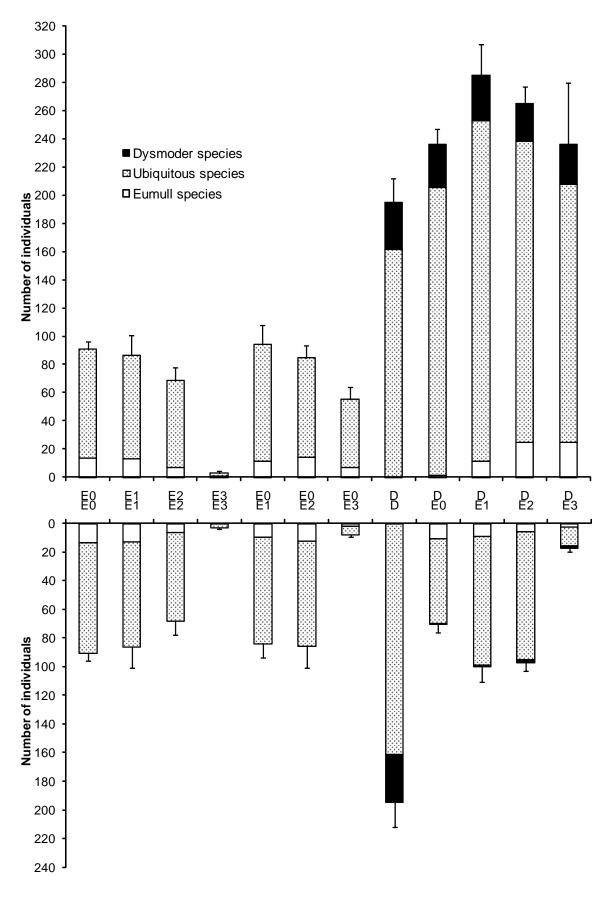
		E0/E0	E1/E0	E2/E0	E3/E0
U	Dicyrtoma fusca	1.8±0.6	1.4±0.4	1.6±0.4	0.8±0.6
U	Folsomia manolachei	18.5±1.3a	14.6±0.7a	14.8±2.0a	1.2±0.7b
U	Friesea truncata	$0.5 \pm 0.3b$	0.8±0.5b	2.0±0.4a	0.2±0.2b
Ε	Heteromurus nitidus	1.3±0.3	0.6 ± 0.4	0.6±0.2	0.2 ± 0.2
U	Isotomiella minor	51.3±3.8a	52.6±8.4a	50.0±13.6a	1.4±0.7b
U	Lepidocyrtus lanuginosus	1.6±0.3	1.6±0.7	2.6±1.2	1.8±0.7
U	Megalothorax minimus	0.5 ± 0.5	0.8±0.5	0.4 ± 0.4	0
U	Paratullbergia callipygos	0.6 ± 0.3	0.8 ± 0.4	0.8±0.4	0
Ε	Parisotoma notabilis	8.7±1.3a	6.0±0.6b	5.2±1.0b	0.4±0.4c
Ε	Pseudosinella alba	2.0±0.4a	1.6±0.4ab	1.0±0.5ab	0.2±0.2b
U	Sphaeridia pumilis	2.1±0.7	1.6±0.9	1.4±0.7	0.4 ± 0.2
	Total abundance	90.4±5.5a	84.0±9.7a	85.8±15.2a	7.8±2.0b
	Abundance of Eumull species	13.3±1.6a	9.6±1.0a	12.2±2.6a	2.0±0.7b
	Abundance of ubiquitous spec	77.1±4.4a	74.4±9.1a	73.6±14.0a	5.8±2.0b
	Total number of species	8.9±0.4a	9.0±0.7a	10.8±1.0a	4.8±0.7b
	Number of Eumull species	4.8±0.5a	3.4±0.7ab	4.6±0.9a	1.6±0.5b
	Number of ubiquitous species	6.4±0.7a	5.8±0.4a	6.2±0.7a	3.0±0.8b
	Water pH	7.66±0.03	7.67±0.01	7.73±0.04	7.73±0.05

Table 6. Mean abundance (\pm S.E.) of most frequent collembolan species and groups of species and water pH in Eumull adjacent to Dysmoder, at different levels of lead application in Eumull, compared to control boxes with unpolluted Eumull in both compartments. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, D = Dysmoder species, U = ubiquitous species. E0/E0 and E0/D = No Pb, E1/D = Pb at 50 ppm, E2/D = Pb at 60,000 ppm, E3/D = Pb at 60,000 ppm

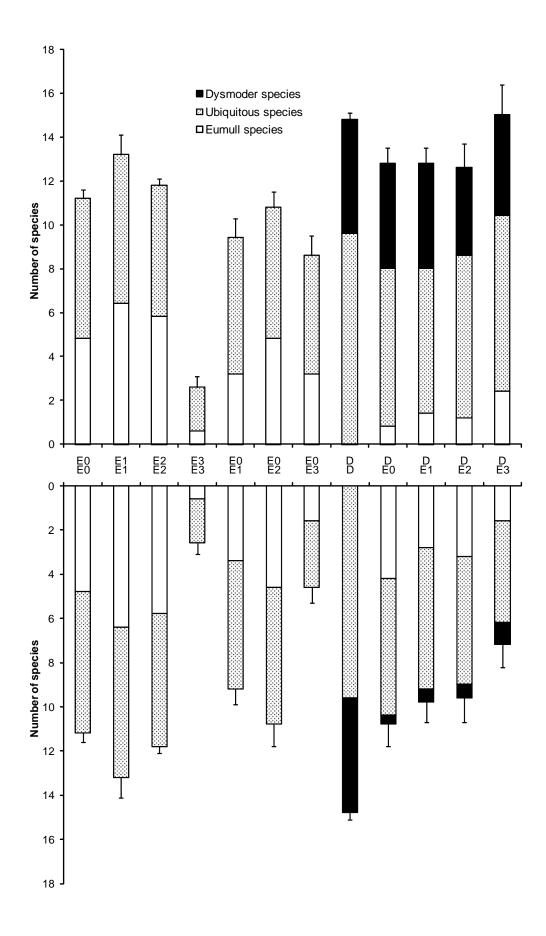
		F0/F0	FO/D	E4/D	F0/D	
		E0/E0	E0/D	E1/D	E2/D	E3/D
U	Dicyrtoma fusca	1.8±0.6	1.6±0.7	1.4±0.5	1.0±0.6	0.8±0.4
U	Folsomia manolachei	18.5±1.3a	12.4±1.5a	12.2±1.7a	18.8±3.9a	5.6±0.6b
U	Friesea truncata	0.5±0.3ab	0.4±0.2ab	1.4±0.5ab	2.0±0.7a	0 b
Ε	Heteromurus nitidus	1.3±0.3	0.6 ± 0.4	0.2±0.2	0.4±0.2	0.2±0.2
U	Isotomiella minor	51.3±3.8ab	38.4±6.4b	69.0±10.8a	57.6±6.6ab	2.0±0.4c
U	Lepidocyrtus lanuginosus	1.6±0.3b	2.2±0.4b	3.2±1.0ab	7.6±2.3a	3.6±0.4a
U	Megalothorax minimus	0.5±0.5ab	0b	1.2±0.4a	0.4±0.2ab	0b
D	Micranurida pygmaea	0	0	0	0.2±0.2	0
U	Paratullbergia callipygos	0.6±0.3ab	2.4±1.0a	0.6±0.4ab	0.6±0.2ab	0 b
Ε	Parisotoma notabilis	8.7±1.3a	6.2±1.6ab	6.2±2.2ab	$3.0\pm0.8b$	1.6±0.7b
Ε	Pseudosinella alba	2.0±0.4a	0.2±0.2b	0.4±0.2b	1.2±0.4ab	0.2±0.2b
D	Pseudosinella mauli	0	0.2±0.2	0.2±0.2	0	0
D	Sminthurinus signatus	0	0.2±0.2	0	0	0
U	Sphaeridia pumilis	2.1±0.7	1.6±0.4	0.6±0.2	0.8±0.6	0.8±0.2
D	Willemia anophthalma	0	0	0.2±0.2	0.6 ± 0.6	1.6±0.7
D	Xenylla tullbergi	0	0	0.2±0.2	1.6±1.6	0.2±0.2
	Total abundance	90.4±5.5ab	70.0±6.3b	99.6±11.2a	97.8±6.1a	17.4±2.6c
	Abundance of Eumull species	13.3±1.6a	10.4±1.1a	9.2±3.0a	5.8±0.7a	2.4±1.0b
	Abundance of Dysmoder spec	0	0.4 ± 0.4	0.6 ± 0.4	2.4±1.7	2.0±0.9
	Abundance of ubiquitous spec	77.1±4.4a	59.2±5.7b	89.8±10.8a	89.0±6.6a	13.0±1.2c
	Total number of species	8.9±0.4	10.8±1.0	9.8±0.9	9.6±1.1	7.2±1.0
	Number of Eumull species	4.8±0.5a	4.2±0.7a	2.8±0.6ab	3.2±0.5ab	1.6±0.5b
	Number of Dysmoder species	0	0.4±0.4	0.6 ± 0.4	0.6 ± 0.4	1.0±0.5
	Number of ubiquitous species	6.4±0.7	6.2±0.5	6.4±0.7	5.8±0.8	4.6±0.4
	Water pH	7.66±0.03b	7.64±0.03b	7.61±0.01b	7.68±0.02b	7.78±0.05a

Table 7. Mean abundance (±S.E.) of most frequent Collembolan species and groups of species and water pH in Dysmoder adjacent to Eumull, at different levels of lead application in Eumull, compared to control boxes with Dysmoder in both compartments. Means followed by a common letter indicate groups which do not differ significantly. E = Eumull species, D = Dysmoder species, U = ubiquitous species. E0/E0 and E0/D = No Pb, E1/D = Pb at 50 ppm, E2/D = Pb at 6,000 ppm, E3/D = Pb at 60,000 ppm

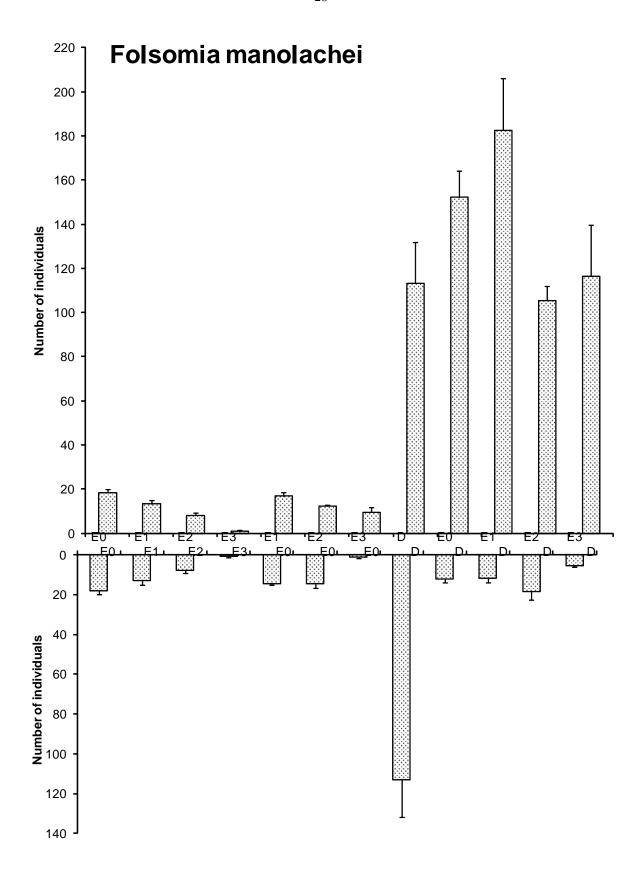
		D/D	D/E0	D/E1	D/E2	D/E3
U	Dicyrtoma fusca	0.6±0.2	1.2±0.5	0	1.2±0.4	1.0±0.0
U	Folsomia manolachei	113.4±18.6ab	152.2±12.0ab	182.6±23.6a	105.6±6.4b	116.6±23.2ab
U	Friesea truncata	0.6±0.1	1.2±0.6	1.4±0.4	0.4±0.4	1.8±0.5
Ε	Heteromurus nitidus	0	0	0	0	0.2±0.2
U	Isotomiella minor	25.7±3.4b	29.4±3.2b	37.4±6.2b	84.6±7.7a	43.0±11.6b
U	Lepidocyrtus lanuginosus	4.8±0.6b	5.4±1.3b	7.4±1.2b	4.4±0.9b	13.2±1.1a
U	Megalothorax minimus	8.9±0.6b	8.6±0.8b	9.6±0.8b	12.4±1.4a	4.6±0.5c
D	Micranurida pygmaea	1.9±0.6	2.2±0.5	1.4±0.7	1.2±0.8	0.6±0.4
U	Paratullbergia callipygos	3.1±0.5a	2.6±0.9ab	0.4±0.2c	2.2±0.7ab	0.8±0.2bc
Ε	Parisotoma notabilis	0c	1.0±0.4c	10.6±3.6b	24.4±4.5a	22.0±5.1a
D	Proisotoma minima	0.4±0.1b	3.0±1.1ab	6.0±0.6a	5.0±1.4ab	6.4±1.9a
D	Pseudosinella mauli	0.7±0.3	0.8±0.4	0.8±0.5	0.4±0.2	0.4±0.2
D	Sminthurinus signatus	1.4±0.3	1.0±0.4	1.0±0.3	0.6±0.2	1.8±0.6
U	Sphaeridia pumilis	2.9±0.5	2.6±1.1	1.6±0.7	0.6±0.4	0.8±0.2
D	Willemia anophthalma	28.6±3.0a	18.8±2.6bc	20.8±5.9abc	10.8±2.5c	17.4±0.9b
D	Xenylla tullbergi	0.3±0.1c	4.6±2.2	1.6±1.1	8.4±5.7	1.8±1.0
	Total abundance	194.6±17.2b	235.8±11.0ab	284.8±22.0a	264.4±12.2a	235.8±43.6ab
	Abundance of Eumull species	0c	1.2±0.5c	11.2±3.7b	24.6±4.6a	24.6±6.2a
	Abundance of Dysmoder spec	33.3±2.5	30.4±4.0	32.2±7.1	26.4±6.6	28.6±3.8
	Abundance of ubiquitous spec	:161.3±16.4	204.2±11.3	241.4±21.1	213.4±13.4	182.6±35.5
	Total number of species	12.0±0.3	13.0±0.7	12.8±0.7	12.6±1.1	15.0±1.4
	Number of Eumull species	0c	0.8±0.4bc	1.4±0.2b	1.2±0.2b	2.4±0.5a
	Number of Dysmoder species	5.2±0.4	4.8±0.5	4.8±0.5	4.0±0.4	4.6±0.6
	Number of ubiquitous species	9.6±0.5a	7.2±0.6b	6.6±0.5b	7.4±0.7b	8.0±0.5ab
	Water pH	4.29±0.01d	4.33±0.02c	4.51±0.07abc	4.46±0.02b	4.65±0.03a



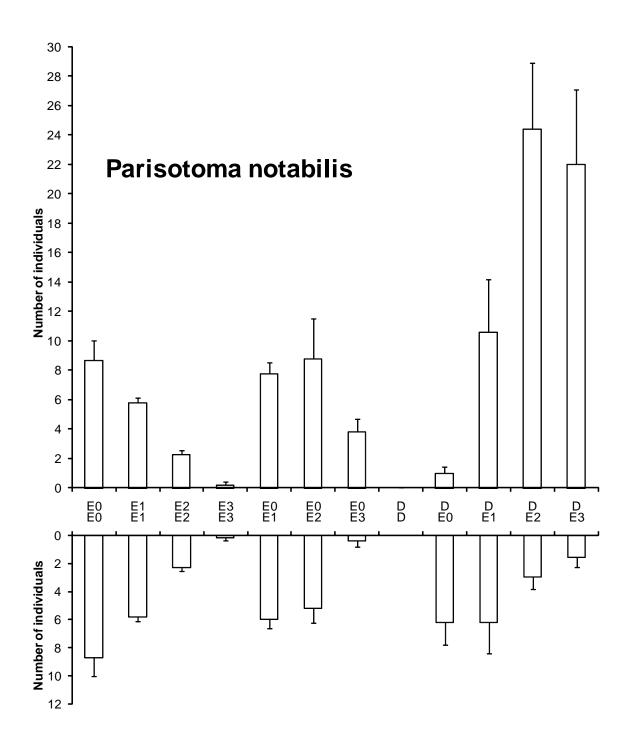
2 Fig. 1



2 Fig. 2



2 Fig. 3



2 Fig. 4