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Changes in carbon density for three old-growth forests on Changbai Mountain, Northeast China: 1981–2010

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Abstract

• **Introduction** The old-growth forests on Changbai Mountain have been well protected from human activities and provide a living laboratory for studying forest carbon sequestration under natural environmental conditions.

• **Objectives** We used data from permanent plots measured periodically in 1981 and 2010 to quantify carbon densities for Korean pine-broadleaf mixed forest, coniferous forest and Erman's birch forest on Changbai Mountain.

• **Results** Carbon pools were divided into tree stems, leaves, branches, coarse woody debris, tree roots, and soil. Although the mixed forest experienced minor wind damage, every forest component except for coarse woody debris experienced increases in carbon density, and the total forest carbon density increased from 233 to 317 t C ha⁻¹. The coniferous forest was severely damaged by wind, so carbon content in trees decreased but the total forest carbon density still increased from 298 to 327 t C ha⁻¹. The birch forest gained much carbon in trees but the soil carbon pool remained relatively

stable, and its total carbon density increased from 226 to 281 t C ha⁻¹. The old-growth forest was more resilient to disturbance than previously thought. The positive increases in carbon for the three old-growth forests suggest that forest landscapes on Changbai Mountain are indeed carbon sinks.

Keywords Carbon sequestration · China forestry · Long-term observation · Temperate forest · Boreal forest

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1 Introduction

Carbon budgets and cycling in forest ecosystems associated with human-induced climate change have been important research topics worldwide during the past few decades. Forest ecosystems store 86% of terrestrial vegetation carbon (Olson et al. 1983) and 73% of soil carbon on Earth (Post et al. 1982). The average carbon density of forests is about 89 t C ha⁻¹ (Dixon et al. 1994), much higher than that of grasslands (21 t C ha⁻¹) and croplands (5 t C ha⁻¹) (Ajtay et al. 1979). Forest carbon fixation capability is 20–100 times higher than that of cropland (Xu and Liu 1992). Therefore, protecting forest ecosystems is an important human action that can play a vital role in offsetting CO₂ emissions and mitigating global climate change.

Changbai Mountain in Northeast China is covered with various vegetation types, including Korean pine-broadleaf mixed forest, coniferous forest and Erman birch forest, and also tundra, and is equivalent geographically to extending from the temperate zone to the North Pole. At the same time, there is a large area of old-growth forests that is under strict protection within the Changbai Mountain Nature Reserve (Shao et al. 1996). These forests provide baselines that reveal the potential for continued carbon sequestration in many places where these forests grow (Rhemtulla et al.

2009). However, the carbon sequestration of Changbai's forests is still quantitatively unclear. It is important to determine if forest carbon sequestration on Changbai Mountain also supports the theory that old-growth forests are carbon sinks found elsewhere on Earth (Luyssaert et al. 2008; Zhou et al. 2006).

Although research on forest structure, function, and productivity on Changbai Mountain was initiated in the early 1980s (e.g., Li et al. 1981), systematic research on forest carbon stock and sequestration did not begin until the 2000s. Earlier studies focused on carbon cycling in individual biomes (e.g., Dai et al. 2002). Zhang et al. (2007) made the first effort to examine the spatiotemporal pattern of ecosystem processes in the Changbai Mountain Nature Reserve by integrating simulation modeling, GIS, remote sensing data, and field-based observations. They found that Changbai's forests as a whole served as a carbon sink and that the broadleaf and Korean pine (*Pinus koraiensis*) mixed forest displayed the highest net primary production (NPP 1.084 kg C m⁻² year⁻¹). Zhu et al. (2010) used 22 forest plots to investigate the altitudinal changes in C storage in three components (vegetation, detritus, and soil) of three forest ecosystems (broadleaf and Korean pine mixed, coniferous, and Erman birch forests) on Changbai Mountain. They suggested that while C density of the three ecosystem components showed distinct altitudinal patterns, carbon storage and partitioning among the different components in temperate forests on Changbai Mountain varied greatly with forest type and altitude.

Clearly there are many unknowns and uncertainties regarding forest carbon stocks and their dynamics on Changbai Mountain. The purpose of this study is to shed more light on carbon sequestration for these three forest types based on long-term ground observations on Changbai Mountain. We do so by addressing the basic scientific question: are old-growth forests on Changbai Mountain carbon sinks or carbon sources?

2 Methods

2.1 Study sites

This study was conducted for three old-growth forests in the Changbai Mountain Biosphere Nature Reserve (41°43'–42°26'N, 127°42'–128°17'E), one of the largest nature reserves in China. In 1980, the reserve became affiliated with the World Biosphere Reserve Network under UNESCO's Man and the Biosphere Project, and is known as Changbai Mountain Biosphere Reserve (CBR) (Tang et al. 2010). CBR is characterized by a mountain climate with strong winds in spring and winter. As altitude increases from 442 m to 2,623 m, the mean annual temperature decreases from 4.9°C to -7.3°C and the mean annual

precipitation increases from 600 mm to 1,340 mm (Zhang et al. 1980). Influenced by the climate gradient, clear altitudinal vegetation zones are formed from the base to the top of Changbai Mountain (Wang et al. 1980). These include: (1) Korean pine-broadleaf mixed forest (740–1,100 m), dominated by Korean pine (*Pinus koraiensis*), Amur linden (*Tilia amurensis*), Manchurian ash (*Fraxinus mandshurica*) and other broadleaf species; (2) coniferous forest (1,100–1,700 m), dominated by Yezo spruce (*Picea jezoensis* var. *komarovii*), Korean pine, Manchurian fir (*Abies nephrolepis*) and Erman's birch (*Betula ermanii*); and (3) Erman's birch forest (1,700–2,000 m), composed solely of this species.

2.2 Data sources

In 1981, 1 ha permanent plots were established in different vegetation zones on the north slope of Changbai Mountain. Every tree ≥ 8 cm in DBH within the plots was identified and its DBH and height were measured. These original measurements and their derivatives for the Korean pine-broadleaf mixed forest at 740 m, the coniferous forest at 1,620 m, and the Erman's birch forest at 1,990 m were used in this article for computing carbon density in 1981. In 2010, the three permanent plots were re-measured following the same procedure used in 1981.

We measured coarse woody debris (CWD, diameter ≥ 2.5 cm, including fallen trees, standing dead trees and stumps) within the plots in 2010. The recorded information includes species (*S*); DBH for standing dead trees or diameter at large-end and small-end for fallen trees and stumps (*D*); length for fallen trees and stumps or height for standing dead trees (*L*); and decay class (1–5) using the classification system reported by Yan et al. (2006) (*C*_{CWD}).

We excavated one soil pit to a depth of 50 cm in each plot, and took soil samples at five depths (0–10, 10–20, 20–30, 30–40, 40–50 cm) in the mineral soil (excluding the organic layer) by using a 100-cm³ soil sampler. Samples were oven-dried in the laboratory at 105°C and weighed to calculate bulk density. Samples were then sub-sampled for measurement of organic matter concentration by wet combustion with K₂Cr₂O₇.

2.3 Carbon computations

The biomass of all living trees, including above-ground (stems, branches, leaves) and below-ground (roots) were estimated by using allometric species-specific equations (JPFD, 2003; Li et al. 1981; Zhong 2009). Carbon density was then calculated as the product of dry mass and assumed C concentration of 50%.

CWD density and carbon content for different decay classes and different species have been measured in previous research (Zhou 2007). Carbon density of CWD for each forest type in 2010 was calculated by species and

decay class. We obtained CWD biomass for the three forest types in 1981 from Liu and Wang (1992), Dai et al. (2000) and Yang et al. (2002). CWD carbon density in 1981 was also calculated as the product of dry mass and assumed C concentration of 50%.

We calculated the soil organic carbon densities for the three forest types in 2010 using the following equation:

$$D = \sum_{i=1}^n 0.58 \times T_i \times B_i \times O_i \quad (1)$$

where D is soil organic carbon density (t C ha^{-1}); n is the number of soil layers; 0.58 is the Bemmelen index that converts organic matter concentration to organic C concentration; and D_i , B_i and M_i are thickness (cm), bulk density ($\text{g } 100 \text{ cm}^{-3}$) and organic matter concentration in soil layer i , respectively.

We calculated the soil carbon densities for the three forest types by referring to published data from Cheng et al. (1981) and Pei et al. (1981). We obtained soil carbon density for each forest type in 1981 based on soil organic matter, soil bulk density and soil thickness (50 cm, the same as in 2010). The above equation was then used to calculate carbon densities in the three forests.

Carbon content from forest floor vegetation was not included in this study, since it accounts for only a small fraction ($< 2\%$) of total forest carbon storage and is relatively stable over time.

3 Results

3.1 Korean pine-broadleaf mixed forest

The total carbon density of Korean pine-broadleaf mixed forest was 233 t C ha^{-1} in 1981 and 317 t C ha^{-1} in 2010, representing an increase of 84 t C ha^{-1} or 36.1% during the past three decades (Table 1). Carbon density in almost all

Table 1 Changes in carbon density (t C ha^{-1}) between 1981 and 2010 for Korean pine-broadleaf mixed forest on Changbai Mountain

Component		1981	2010
Above-ground	Stems	93.3	127
	Leaves	2.85	3.89
	Branches	12.4	16.9
	Coarse woody debris	8.13	7.53
Sub-total		117	156
Below-ground	Roots	25.5	34.8
	Soil	90.8	126
Sub-total		116	161
Total		233	317

the components increased consistently except for that in CWD, which decreased slightly from 8.13 to 7.53 t C ha^{-1} from 1981 to 2010. The increase in carbon density below ground (38.6%) exceeded that for above ground (33.5%).

Above ground carbon density was nearly the same as that below ground in 1981, but was higher than the latter in 2010. Tree stems contained the most carbon among above-ground components, followed by branches or CWD, depending on the year. Carbon density in soil was over three times higher than that in tree roots, and even exceeded that in tree stems in 2010.

3.2 Coniferous forest

The total carbon density of the coniferous forest was 298 t C ha^{-1} in 1981 and 327 t C ha^{-1} in 2010, representing an increase of 29 t C ha^{-1} or 9.74% during the past three decades (Table 2). Carbon density in all living tree components decreased, but that in CWD doubled (from 26.7 to 55.1 t C ha^{-1}). Carbon density in soil increased from 124 to 148 t C ha^{-1} or by 19.9% between 1981 and 2010. The decreases in carbon density in living tree components were consistent, ranging from 15% to 18%.

In 1981, above-ground carbon density was nearly the same as below-ground density, while it was slightly lower than the latter in 2010. Tree stems contained the most carbon among above-ground components throughout this period. Carbon density in soil was almost 5–8 times higher than that in tree roots, and even exceeded that in tree stems. Carbon density in soil was the highest among all the forest components.

3.3 Erman's birch forest

The total carbon density of the birch forest was 226 t C ha^{-1} in 1981 and 281 t C ha^{-1} in 2010, representing an increase of 55 t C ha^{-1} or 24.3% during the past three decades (Table 3). Carbon density in all the above-ground components increased by 72.4%. Although soil carbon

Table 2 Changes in carbon density (t C ha^{-1}) between 1981 and 2010 for coniferous forest on Changbai Mountain

Component		1981	2010
Above-ground	Stems	105	88.1
	Leaves	5.70	4.67
	Branches	15.3	12.8
	Coarse woody debris	26.7	55.1
Sub-total		152	160
Below-ground	Roots	22.3	18.9
	Soil	124	148
Sub-total		146	167
Total		298	327

Table 3 Changes in carbon density (t C ha^{-1}) between 1981 and 2010 for Erman's birch forest on Changbai Mountain

Component		1981	2010
Aboveground	Stems	47.7	80.3
	Leaves	0.95	1.61
	Branches	5.73	9.63
	Coarse woody debris	1.72	5.25
Sub-total		56.1	96.7
Belowground	Roots	14.3	24.1
	Soil	156	160
Sub-total		170	184
Total		226	281

density was nearly static between 1981 and 2010, carbon in roots increased by 68.2%. Below-ground carbon density increased by 8.36%.

Although above-ground carbon density was much lower than below-ground density, the gap narrowed over time. The majority of above-ground carbon was stored in tree stems. Carbon density in soil was the highest among all the forest components.

3.4 A comparison among three forest types

In both 1981 and 2010, the coniferous forest had the highest carbon density, followed by the mixed forest and birch forest. All three forests experienced increases in above-ground, below-ground and total carbon densities, with the exception of tree components of the coniferous forest. For the mixed forest, the increase in carbon density resulted simultaneously from growth of above-ground living tree components and carbon accumulation in the soil (Table 1). For the coniferous forest, this increase resulted primarily from carbon accumulation in CWD and the soil (Table 2); while for the birch forest, this increase resulted only from the growth of living tree components (Table 3). Since soil was the largest carbon pool, especially for the coniferous and birch forests, changes in soil carbon density had a major influence on the total below-ground carbon density in those three forests (Tables 1–3). This suggests that carbon density in living trees is higher at lower elevations, whereas that of soil carbon density is greater at higher elevations.

4 Discussion

4.1 Implications

From 1981 to 2010, the mixed, coniferous and birch forests on Changbai Mountain experienced increases in above-ground,

below-ground and total carbon densities. During this time, the total carbon densities of the three old-growth forests increased by 84, 29, and 55 t C ha^{-1} , respectively. These results confirm that the forests have acted as a carbon sink during the past three decades, having sequestered 2.79, 0.97 and 1.83 $\text{t C ha}^{-1} \text{ year}^{-1}$, respectively. These numbers are higher than those reported for other old-growth forests (Harmon et al. 2004; Luyssaert et al. 2008).

We also found that the above-below ground carbon ratio varied among the three forest types. The aboveground components contributed about 47%, 27% and 74% of the carbon sink in the mixed, conifer and birch forests, respectively; while 53%, 73% and 26% of the carbon sink was contributed by belowground components (mainly soil organic carbon) during the past three decades (Fig. 1).

Carbon density in all the components of the mixed forest increased consistently from 1981 to 2010, suggesting that this forest type is a stable carbon sink. Because the mixed forest is the most extensive type in the greater Changbai Mountain area, the entire forest landscape is likely a carbon sink (Zhang et al. 2007). This is even more likely due to China's Natural Forest Conservation Program (Zhang et al. 2000), which has banned logging from many forests on Changbai Mountain since 1998. Because the mixed forest plot was located inside the Changbai Mountain Nature Reserve, where human disturbance is limited, the increase in carbon density in the mixed forest suggests that the existing climatic conditions have been favorable to the growth of this forest during the past three decades.

A decrease in living tree carbon density (16.9 t C ha^{-1}) and an increase in CWD carbon density (28.4 t C ha^{-1}) in the coniferous forest suggest that wind disturbances have occurred in this forest between 1981 and 2010. Forest gaps were formed and, as a result, forest floors should receive more solar radiation, which could result in higher soil

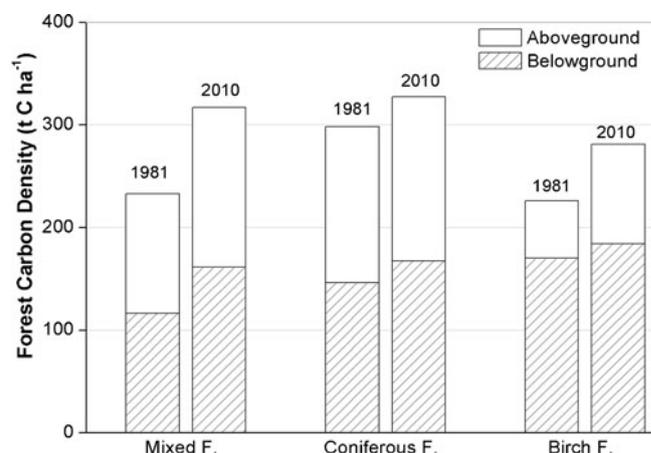


Fig. 1 A comparison of the carbon densities for three forest types on Changbai Mountain

temperature for the coniferous forest. However, the wind disturbances did not seem to damage the soil carbon pool despite the fact that the stability of that pool in the boreal forest is sensitive to temperature increases (Goulden et al. 1998). The soil organic carbon in the coniferous forest increased by 24 tC ha^{-1} during the three decades and contributed about $0.8 \text{ tC ha}^{-1} \text{ year}^{-1}$ to the carbon sink. This result was consistent with those reported by Zhou et al. (2006) for subtropical evergreen old-growth forests. It is possible that extremely low temperatures and high winds led to a decrease in the decomposition rate of litter and CWD and an increase in organic carbon storage over time. It should be noted, however, that, despite the wind damage, the coniferous forest was still a carbon sink, thereby further enhancing the likelihood that forests on Changbai Mountain are a carbon sink. In effect, the old-growth forest appears to be more resilient to disturbance than we initially anticipated (Goulden et al. 1998).

The birch forest is composed almost entirely of Erman's birch, reflecting the fact that this tree species can tolerate a windy and cold environment. A sharp increase in living tree carbon density may be driven by increased temperature, although forest mortality increased as indicated by a three-fold increase in CWD. Wind disturbances between 1981 and 2010 may have exceeded the tolerance of Erman's birch.

4.2 Certainty vs uncertainty

The estimates of carbon density for living trees and changes in this density over time are relatively reliable because: (1) the three permanent forest plots were selected because they were representative of typical forest types on Changbai Mountain (Dai et al. 2011); (2) carbon contents in living trees were computed by using species-specific biomass equations developed from trees sampled in nearby forests; and (3) repeated measurements over time were used to quantify changes in forest biomass and carbon. Because a single 1-ha plot was used for each forest type, we were unable to compute the precision of carbon density estimates.

Carbon contents in CWD increased in the three forests, particularly in the coniferous forest. Major wind throw started to occur in the mid-1980s (Shao et al. 1996) and has caused continuous damage to forests on Changbai Mountain since then (Tang et al. 2010). Trees at higher elevations experienced the most damage and spruce/fir trees were more vulnerable than birch to strong winds in winter. This is why there were many more dead trees in the coniferous forest than in others.

There was great variation in soil carbon density on Changbai Mountain, especially in the mixed and coniferous forests. It is worth noting that the measurements of soil carbon densities in 1981 for the three forest types did not take changes of bulk density with depth of soil into account

although it was known that bulk density typically increases with soil depth (Wu et al. 2004). But the data we used were consistent over time. Therefore, we have reasonable confidence about the changes in soil carbon density over time for each forest.

In this research we did not calculate carbon accumulation in litter and organic layers. Past studies suggest that the carbon storage in litter is relatively small and stable in general, accounting for only 1–4% of the total forest carbon storage (Fang et al. 2002; Zhong 2009; Zhu et al. 2010).

5 Conclusions

The mixed, coniferous and birch forests, representing those from low to high elevations on Changbai Mountain, experienced different carbon processes over the period from 1981 to 2010. The mixed forest at a lower elevation was not faced with as much wind damage as were the two higher-elevation forests. The relatively favorable environment at lower elevations aided carbon accumulation in every component of the mixed forest. In contrast, wind-throw increased forest mortality and created canopy gaps in coniferous and birch forests. The natural disturbances did not damage the soil carbon pool in either of these forests. Total forest carbon densities for the coniferous and birch forests also increased from 1981 to 2010. The positive increases in carbon densities in the three forests indicate that the old-growth forest is more resilient to disturbance than previously thought, and that the forest landscape on Changbai Mountain is indeed a carbon sink. Our finding at the local scale is consistent with the suggestion of Luyssaert et al. (2008) and Zhou et al. (2006) that old-growth forests can continue to accumulate carbon. Should increased wind disturbances be one consequence of underlying global climate change, then the forest carbon dynamics of old-growth forests have indeed been altered but not destroyed by the changing climate.

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