



**HAL**  
open science

## Odour-impact assessment around a landfill site from weather-type classification, complaint inventory and numerical simulation

Charles Chemel, Caroline Riesenmey, Mireille Batton-Hubert, Hervé Vaillant

► **To cite this version:**

Charles Chemel, Caroline Riesenmey, Mireille Batton-Hubert, Hervé Vaillant. Odour-impact assessment around a landfill site from weather-type classification, complaint inventory and numerical simulation. *Journal of Environmental Management*, 2012, Volume 93 (Issue 1), pp.85-94. 10.1016/j.jenvman.2011.08.016 . emse-00637877

**HAL Id: emse-00637877**

**<https://hal-emse.ccsd.cnrs.fr/emse-00637877>**

Submitted on 14 Nov 2011

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

# Odour-impact assessment around a landfill site from weather-type classification, complaint inventory and numerical simulation

C. Chemel<sup>a,\*</sup>, C. Riesenmey<sup>b</sup>, M. Batton-Hubert<sup>b</sup>, H. Vaillant<sup>b</sup>

<sup>a</sup>*National Centre for Atmospheric Science, Centre for Atmospheric & Instrumentation Research, University of Hertfordshire, College Lane, Hatfield, Herts AL10 9AB, UK*

<sup>b</sup>*Centre SITE, ENS des Mines de Saint-Etienne, 158 cours Fauriel, 42023 Saint-Etienne, Cedex 2, France*

---

## Abstract

Gases released from landfill sites into the atmosphere have the potential to cause olfactory nuisances within the surrounding communities. Landfill sites are often located over complex topography for convenience mainly related to waste disposal and environmental masking. Dispersion of odours is strongly conditioned by local atmospheric dynamics. Assessment of odour impacts needs to take into account the variability of local atmospheric dynamics. In this study, we discuss a method to assess odour impacts around a landfill site located over complex terrain in order to provide information to be used subsequently to identify management strategies to reduce olfactory nuisances in the residential neighbourhoods. A weather-type classification is defined in order to identify meteorological conditions under which olfactory nuisances are to be expected. A non-steady state Gaussian model and a full-physics meteorological model are used to predict olfactory nuisances for both the winter and summer scenarios that lead to the majority of complaints in neighbourhoods surrounding the landfill site. Simulating representative scenarios rather than full years make a high resolution simulation of local atmospheric dynamics in space and time possible. Results underline the key role of local atmospheric dynamics in driving the dispersion of odours. The odour concentration simulated by the full-physics meteorological model is combined with the density of the population in order to calculate an average population exposure for the two scenarios. Results of this study are expected to provide helpful information to develop technical solutions for an effective management of landfill operations, which would reduce odour impacts within the surrounding communities.

*Keywords:* Landfill site; Olfactory nuisances; Complex terrain; Data classification; Numerical simulation

---

\* Corresponding author. Tel.: 441707286143; fax: 441707284208.  
*E-mail address:* c.chemel@herts.ac.uk (C. Chemel).

## 24 1. Introduction

25 Municipal solid waste (MSW) is often disposed into landfill sites. A fraction of the landfill gases are  
26 emitted into the atmosphere. These gases are either originally present in the waste or formed during its  
27 decomposition process. Organic matter in the waste decomposes, while producing methane ( $\sim 60\%$ ),  
28 carbon dioxide ( $\sim 40\%$ ), and non-methane volatile organic compounds, referred to as VOCs (see for  
29 instance Brosseau and Heitz, 1994). Beyond major concerns related to environmental management, the  
30 greenhouse effect and health hazards, waste disposal units are potential sources of olfactory nuisances  
31 in residential neighbourhoods. Although biogas is usually collected and treated, and soil is covered to  
32 avoid emission, some of the landfill gases diffuse into the atmosphere, especially from the working face  
33 (typically  $\sim 20\%$ ) (Spokas *et al.*, 2006). Among the gases released in the atmosphere, undesirable trace  
34 VOCs contribute to a large degree to poor air quality (*e.g.* Allen *et al.*, 1997; Kim *et al.*, 2005). In addition,  
35 biological heat may significantly modify the energy balance of the working face and may thus lead to an  
36 increase in the net energy flux (Bendz and Bengtsson, 1996).

37 Although odour pollution events may be attributed to increases in the emission of landfill gases (for  
38 instance due to special manipulation of the waste), these events are usually associated with ‘stagnant’  
39 meteorological conditions with limited vertical mixing and low wind speeds, for which dispersion of odours  
40 is reduced. Landfill sites are often located over complex topography for convenience mainly related to  
41 waste disposal and environmental masking. Meteorological conditions are difficult to predict over complex  
42 terrain, and odour impacts are correspondingly difficult to assess.

43 Several studies have been conducted to identify and characterize relationships between meteorologi-  
44 cal conditions and impaired air quality episodes. Principal Component Analysis (PCA) and clustering  
45 techniques are commonly used to provide representative synoptic meteorological scenarios for air qual-  
46 ity studies (*e.g.* Eder *et al.*, 1994; Greene *et al.*, 1999; Kim Oanh *et al.*, 2005). However, only a few studies  
47 discussed the application of such classification methods over complex terrain. As pointed out for instance  
48 by Berman *et al.* (1995), synoptic weather-type classifications only give the atmospheric conditions under  
49 which a local-scale study should be further conducted to understand local-scale dispersion of pollutants.  
50 Nanni *et al.* (2004) used a local-scale approach to sort data into predefined weather-type classes in an  
51 alpine region. Brulfert *et al.* (2006) used a similar technique in two alpine valleys, while proceeding from  
52 local- to synoptic-scale weather-type classes to sort data.

53 Dispersion models have become a common tool to evaluate the impacts of odour sources for given me-  
54 teorological conditions (Yang and Hobson, 2000; McIntyre, 2000; Stuetz and Frechen, 2001). Most of the  
55 models that have been applied to odour-impact assessment are Gaussian models (Sarkar *et al.*, 2003). These

56 models are usually not designed to account for the variable characteristics of the dispersion process as well  
57 as for complex terrain (*e.g.* Ormerod, 2001). Even if the near-field dispersion is for the most part driven  
58 by the meandering behaviour of the plume and not so much by turbulent processes, the unsteady turbu-  
59 lent behaviour of the atmosphere needs to be considered appropriately (see for instance Aubrun and Leidl,  
60 2004). Non-steady state models were used to overcome this issue (*e.g.* Mussio *et al.*, 2001; Schaubberger  
61 *et al.*, 2001; Tagaris *et al.*, 2003; De Melo Lisboa *et al.*, 2006). In addition, some non-steady state Gaussian  
62 models make it possible to deal with complex terrain, such as the Atmospheric Dispersion Modelling Sys-  
63 tem (ADMS) (Carruthers *et al.*, 1994). However, the Gaussian approach is generally limited by simplified  
64 physics as well as a poor representation of the meteorological forcing.

65 The distinct objectives of our work are (*i*) to identify relationships between regional- and local-scale  
66 atmospheric dynamics and odour pollution events from a landfill site located over complex terrain, and (*ii*)  
67 to evaluate the value of both a non-steady state Gaussian model and a full-physics meteorological model  
68 in predicting population exposure to olfactory nuisances around the landfill site. The present study aims  
69 to provide relevant information to develop effective control or warning strategies with respect to olfactory  
70 nuisances in the nearby neighbourhood.

71 The outline of the paper is as follows. The landfill site and experimental data are presented in § 2. In § 3,  
72 PCA and clustering techniques are applied to both regional- and local-scale data in order to identify the  
73 weather types that favour odour pollution events around the landfill site. An overview of the Gaussian and  
74 meteorological models is given in § 4. In § 5, results from the models are discussed for the two weather  
75 types that lead to the majority of complaints in the vicinity of the landfill site. Conclusions and suggestions  
76 for further work are given in § 6.

## 77 **2. Observational site and experimental data**

### 78 *2.1. Site description*

79 The landfill site is located North of the French Alps. It is surrounded westwards by the Massif Central  
80 mountain range and eastwards by the Rhône corridor. The site is embedded within a complex terrain at the  
81 foothill of the Pilat Regional Nature Park, which reaches an altitude of 1432 m above ground level (a.g.l.)  
82 (see Fig. 1). The waste is heaped up into a small valley. Three major towns (with populations in the range  
83 10, 000–25, 000 inhabitants) spread around the landfill site within a 5-km radius, and are denoted by RLM,  
84 LCF and FIR in the present study. The landfill site is one of the five largest French disposal facilities. The  
85 site receives more than 500 ktons of waste (mainly MSW) every year. The waste is composed of 50 % of  
86 solid household refuse, 40 % of non dangerous industrial waste, and 10 % of sewage sludge. The filling

87 of the site started approximately 20 years ago and is expected to finish in about 15 years. The compacted  
88 waste is covered with a soil-covering, except the working face (*i.e.* the open cell), which encompasses an  
89 area of about 5000 m<sup>2</sup>. Landfill leachate is collected and discharged into a collection and treatment system.  
90 Biogas is collected and burned to produce electricity. Since residential areas are located close to the site,  
91 manipulations of fresh waste over the open cell may lead to olfactory nuisances within the surrounding  
92 communities.

## 93 2.2. *Equipment and collection of data*

94 A ground meteorological monitoring station is located in the landfill site area (see Fig. 1). Data from  
95 this station was recorded from 2002 and 2004, and includes pressure, temperature, relative humidity, wind  
96 speed and direction, and precipitation. The operational mode continuously samples these variables, using  
97 a 30-min acquisition cycle. A detailed complaint inventory (consisting of date, duration and location of  
98 olfactory nuisances) is available from 2002 to 2004 and contains a total of 71 complaints. Complaints were  
99 reported within a radius of about 5 km around the landfill site.

100 In order to identify the trace VOCs emitted by the landfill site and to quantify their emission rates,  
101 a field sampling was undertaken by TERA Technologies during workdays from 23 to 25 August 2005.  
102 The sampling site was located a few meters from the working face and the air was sampled at about 2 m  
103 above the ground. These days were typical of clear-sky summertime anticyclonic conditions. Sampling was  
104 carried out using sorbent materials (Tenax collectors). Each sample consists of an adsorptive tube which  
105 was loaded for 4 min using a pump at a rate of 0.1 L min<sup>-1</sup>, then desorbed and analyzed for 10 min. A  
106 total of 125 samples were analyzed continuously in time (one sample every 14 min). VOC analysis was  
107 performed using an automated system including a preconcentrator. Samples were thermally-desorbed at  
108 220 °C and transferred to a Gas Chromatography/Mass Spectrometry (GC/MS) system. The separation  
109 of the compounds was performed using either an OV-1 (polymethylsiloxane) or a Poraplot Q (styrene-  
110 divinylbenzene) GC column coupled with the MS, which covered the mass range 35–250 amu. A standard  
111 semi-quantitative analysis was carried out afterwards by comparing the compound mass spectra with those  
112 of a reference database.

113 Several VOC species were identified including including alkanes (*e.g.* heptane, decane), terpenes (*e.g.*  
114  $\alpha$ -pinene, limonene), aromatic compounds (*e.g.* toluene, xylene isomers), and chlorinated compounds (*e.g.*  
115 trichloroethylene, tetrachloroethylene). The trace VOCs emitted by the landfill site are comparable to those  
116 reported in previous studies (see for instance Allen *et al.*, 1997). Among these VOCs, toluene was selected  
117 as a typical trace VOC that is representative of the source under investigation and has a high emission rate  
118 (see also Davoli *et al.*, 2003). Assuming that emissions are fairly homogeneous over the area of the working

119 face (assumption to be tested in future research), the toluene emission rate was calculated by multiplying  
120 the concentration of toluene by the volume flow rate. A ‘generic’ daily emission profile was derived by  
121 averaging toluene emissions over the period of the field sampling. The emission profile was smoothed  
122 using a 1-h running average and normalized by the maximum emission. The resulting emission profile (see  
123 Fig. 2) is used in § 5 to investigate the dispersion of odours around the landfill site.

Fig. 2

### 124 3. Weather-type classification

#### 125 3.1. Regional-scale approach

126 As a first attempt to characterize typical atmospheric conditions around the landfill site, we have clas-  
127 sified the weather types using synoptic-scale criteria. The data was retrieved from twice daily operational  
128 radiosoundings in Lyon (France). Lyon is located approximately 100 km North-East of the landfill site.  
129 The data were extracted in the range 500–850 hPa and includes the extrema of the potential temperature  
130 gradient at 00 Coordinated Universal Time (UTC), 12 UTC, and 24 UTC to characterize regional atmo-  
131 spheric stability, the difference between air and dew-point temperatures at both 500 and 850 hPa at 12 UTC  
132 to characterize precipitation, and the wind speed and direction at both 500 hPa and 850 hPa and averaged  
133 between 00 UTC and 24 UTC. A PCA algorithm (see for instance Lebart *et al.*, 1997) was applied to the  
134 data from 2002 to 2004. A standard multivariate statistical method was used to identify the linearly inde-  
135 pendent components, which explain data variability (*e.g.* Kim Oanh *et al.*, 2005). Missing data was not  
136 taken into account in order to prevent artificial data from being included in the pooling procedure. All the  
137 factors deduced from the PCA algorithm were found to be significant and were thus retained for the data  
138 classification. Ascending hierarchical classification and K-mean cluster analysis were applied to divide the  
139 dataset into classes.

Table 1

140 The resulting regional weather classes are reported in Table 1. Eleven classes (weather types) were ob-  
141 tained. In order to discuss the relevance of these weather types with respect to odour nuisances, the number  
142 of complaints associated with each class is also indicated in Table 1. Two weather types induce more than 15  
143 complaints per 100 days, whereas the other ones generate less than 7 complaints per 100 days. Interpreting  
144 the physical meaning of these results may be premature although two weather types represent the majority  
145 of the recorded complaints. Precipitation and wind speed seem to be the most explanatory variables. An  
146 undeniable feature is that precipitation alleviates the emission of odours and cleans the atmosphere. Also,  
147 the presence of wind partly determines the dispersion of odours, while no wind favours odour stagnation.  
148 Interestingly, the regional atmospheric stability of the air mass does not provide any additional information  
149 since complaints occur for all stability criteria. It is very likely that this parameter differs significantly at

150 the landfill site and 100 km North.

151 This regional-scale approach suggests that one need to focus on clear-sky and calm-weather conditions,  
152 which favour odour pollution events. As pointed out for instance by Brulfert *et al.* (2006), local atmospheric  
153 dynamics determines to a large extent air quality over complex terrain. As a result, a weather-type classifica-  
154 tion would need to be developed by taking into account local-scale variables. In addition, the regional-scale  
155 approach is found insufficient since it does not provide any information about the location and the time pe-  
156 riod of the odour pollution events. Indeed, we found that the complaints that were recorded are somewhat  
157 evenly distributed during the day (Riesenmey, 2008). So, the weather-type classification needs to be refined  
158 by turning to a local-scale approach.

### 159 3.2. *Local-scale approach*

160 Atmospheric dynamics around the landfill site results from the combined effects of synoptic-scale dy-  
161 namics and local-scale dynamics induced by complex terrain. Local features induced by the topography  
162 such as valley and slope winds, and frequent temperature inversions (long-lasting during wintertime) have  
163 indeed a strong impact on air quality. Local-scale phenomena may either dilute or confine the air mass  
164 within the landfill site area. A methodology similar to the one at the regional scale was applied to data  
165 from the ground monitoring station. 9 local weather types were identified. Only the two most represen-  
166 tative classes that lead to the majority of complaints are retained for the discussion. These classes are  
167 characterized by high-pressure systems with no wind in both winter and summer. The spatial and temporal  
168 repartition of the complaints for both local-weather types is displayed in Fig. 3. In winter most of the odour  
169 pollution events occur in the evening in FIR, whereas in summer they occur in the morning in LCF. Note  
170 that no complaint was recorded at night (between 2300 UTC and 0600 UTC). The results show only a weak  
171 contribution of the average temperature to odour formation. Conversely, no wind conditions often lead to  
172 olfactory nuisances. This is consistent with the regional-scale weather-type classification. Nonetheless, the  
173 two local-scale classes contain less complaints per 100 days (about 8.9 in summer and 7.9 in winter) than  
174 the two regional-scale classes, which contain the majority of complaints.

175 While regional data gives useful information about the probability of odour events and complaints, local  
176 data gives essential information about the location of the complaints in space and time. Moreover, we  
177 found that regional and local data has to be treated separately. Indeed, it turned out that an analysis using  
178 both regional- and local-scale data leads to ill-defined weather classes (Riesenmey, 2008). In fact, these  
179 classes do not provide as much information about the probability of odour events and complaints as the  
180 regional classes and about the location of the complaints in space and time as the local classes. This section  
181 underlines the important role of local (thermally-driven) winds on odour dispersion around the landfill

182 site. Therefore, we may conclude that one must account for small-scale atmospheric processes to achieve  
183 accurate prediction of odour events in the vicinity of the landfill site.

#### 184 4. Numerical tools

185 The data classification presented above enables us to assign each day to a class representative of a weather  
186 type. Hereafter we focus on the winter and summer classes from the local-scale approach. A representative  
187 day of each of the two classes was identified as the closest to the barycentre of the point cloud points in  
188 the input variable space (Tirabassi and Nasseti, 1999; Sfetsos *et al.*, 2005). The days thus obtained are: 19  
189 October 2002 for the winter class and 18 August 2002 for the summer class.

190 In the following, we evaluate the performance of both a non-steady state Gaussian model and a full-  
191 physics meteorological model to predict olfactory nuisances around the landfill site for the two days rep-  
192 resentative of the winter and summer scenarios. The prediction of the frequency of odour pollution events  
193 is very challenging since odour is a subjective information, which depends on human perception. The  
194 nonlinear relationship between odour intensity and odour concentration is usually described by empirical  
195 psychophysical laws (see for instance Sarkar and Hobbs, 2002). Nicell (2003) suggested dose-response re-  
196 lationships, which can be used to estimate contours of probability of response and degree of annoyance. In  
197 our study, we simply assume a bijective linear relationship between odour intensity and pollutant (or tracer)  
198 concentrations. As pointed out for instance by Termonia and Termonia (1999), it is unrealistic to explain  
199 the complex interaction and interference between all the compounds, which lead to the formation of odours.  
200 However, we may assess the impacts of odours that can be associated with a tracer of these odours. The  
201 dispersion of odours can thus be easily predicted using a dispersion model or a full-physics meteorological  
202 model. Thereafter, toluene is considered as a passive tracer for waste odour. The dominant tropospheric loss  
203 process is by reaction with the OH radical. The lifetime of toluene due to reaction with the OH radical is  
204 in the order of 2 days (Atkinson, 2000), so that it can be considered as a passive tracer on the time scale of  
205 a day or so. As indicated by Tagaris *et al.* (2003), for odorous species with high reactivity or short lifetime  
206 in the atmosphere, a chemistry-transport model should be used. The normalized emission profile of toluene  
207 displayed in Fig. 2 was used as a reference emission profile for both scenarios, even though it is represen-  
208 tative of the summer class only. Odour emissions from the working face of the landfill site were assumed  
209 to cover one grid cell and were considered as area source emissions in both modelling systems.

#### 210 4.1. The Gaussian model

211 Numerical simulations were conducted using the ADMS model (see Carruthers *et al.*, 1994, for a de-  
212 scription of the model). This non-steady state Gaussian model has been extensively used to investigate  
213 several case studies under various meteorological conditions as well as over complex terrain. The effect of  
214 complex terrain on the wind flow was taken into account by using the complex terrain module (Carruthers  
215 *et al.*, 1994). The model has already been applied to simulate the dispersion of odours (*e.g.* Hobbs *et al.*,  
216 2000). The computations were performed on a 5 km  $\times$  5 km domain with a 100-m horizontal resolution.  
217 Such a high spatial resolution is required to faithfully account for dispersion at the local scale. The grid was  
218 centered over the landfill site area. The domain covers most of the area displayed in Fig. 1.

219 With respect to input requirements, data includes source characteristics (toluene molecular weight and  
220 molar heat capacity, emission rate and temperature), terrain properties (location, topography and rough-  
221 ness), and hourly surface meteorological data (pressure, temperature, relative humidity, wind speed and  
222 direction and precipitation). ADMS also requires as a bare minimum cloud cover, time of day and time  
223 of year. Note that the sky was clear during the two scenarios that were simulated, so that the cloud cover  
224 was set to zero. The model provides hourly ground surface concentrations over the whole computational  
225 domain. Note that ADMS does not provide any result when the wind speed is lower than 0.75 m s<sup>-1</sup>,  
226 leading to a stagnation of odours in the model under such conditions.

#### 227 4.2. The meteorological model

228 The ARW (Advanced Research core of the Weather Research and Forecasting (WRF) model, Skamarock  
229 *et al.* 2008) model (simply referred to as WRF thereafter) was used to simulate the dispersion of toluene  
230 around the landfill site. We used the WRF/Chem add-on to the WRF model, which provides a capability  
231 for the modelling of the dispersion of tracers. The model was run on multiple grids using one-way nests  
232 down to a horizontal resolution of 100 m. Five nests using horizontal resolutions of 16 km, 4 km, 1 km, 300  
233 m, and 100 m were used. The inner grid encompasses most of the domain in Fig. 1. The computations were  
234 made on 28 vertical levels up to 50 hPa. The grid was stretched along the vertical axis to accommodate a  
235 high resolution ( $\sim$  30 m on average) close to the ground surface. The averaged vertical grid spacing was  
236 about 500 m. For the two inner-most domains (using a horizontal resolution of 300 m and 100 m) we used  
237 high-resolution digital elevation, soil type, and landcover data at  $\sim$  10-m resolution. For the other domains  
238 and the other characteristics of the soil and the ground surface (*e.g.* monthly surface albedo), static data was  
239 derived from the default geographical data that is provided with the WRF preprocessing system.

240 Initial and lateral boundary conditions of the coarser domain were derived from the ECMWF (European

241 Centre for Medium-range Weather Forecasts) gridded analyses available every 6 h with a horizontal res-  
242 olution of  $0.5^\circ$  on operational pressure levels up to 50 hPa for vertically distributed data, and surface and  
243 soil levels for surface and deep-soil data. A grid nudging technique (see for instance Stauffer and Seaman,  
244 1990) was employed for the coarser domain during the first 6 h in order to constrain the model towards the  
245 analyses and to shorten the spin-up time. A relaxation zone covering 5 grid cells around each domain was  
246 employed to smooth gradients near the lateral boundaries.

247 For the simulations using a horizontal resolution greater than or equal to 1 km, we used the YSU non-  
248 local boundary-layer parameterization scheme (Hong *et al.*, 2006) for which sub-grid scale (SGS) mixing  
249 is classically parameterized within the scheme. This scheme assumes that there is a clear scale separation  
250 between the sub-grid and resolved scales, so that they can be treated separately. This assumption is not  
251 warranted as grid sizes approach a few hundred meters or less. Hence, for the finer-resolved domains using  
252 a horizontal resolution of 300 m and 100 m, we used a fully 3D local SGS parameterization scheme,  
253 namely the level-1.5 SGS parameterization scheme by Deardorff (1980). The Monin-Obukhov surface  
254 layer scheme was used to provide surface forcing in terms of momentum, heat, and moisture fluxes. The  
255 land-surface energy budget was calculated by the Noah soil-vegetation model (Ek *et al.*, 2003).

256 Other physics options that we used include the CAM3 radiation package (Collins *et al.*, 2006), the  
257 microphysical scheme by Thompson *et al.* (2004, 2006), and the ensemble cumulus scheme introduced by  
258 Grell and Dévényi (2002) for the coarser grids with a horizontal resolution larger than 4 km. Note that for  
259 the finer-resolved grids with a horizontal resolution of 1 km and less, convection was explicitly resolved  
260 (*i.e.* the cumulus scheme was switched off).

## 261 5. Results and discussion

### 262 5.1. Winter and summer scenarios

263 The meteorological model WRF was evaluated using observational data across the different nests. Re-  
264 sults of the simulations were compared with both ground surface and vertically-distributed measurements.  
265 This evaluation is not detailed here since it is not the focus of the present paper. The nudging technique  
266 that was used did constrain the model to remain close to observational data. Figure 4 shows that the WRF  
267 model is able to capture very well the temporal variations in wind speed and direction at the location of the  
268 ground meteorological monitoring station for the summer and winter scenarios. In the present study, the  
269 strategy for further analysis was to select appropriate times of the day to optimally characterize dispersion  
270 around the landfill site. Hereafter, we decided to use 1000 UTC, 1400 UTC, and 1800 UTC. These times  
271 are representative of morning, afternoon, and evening conditions, respectively. The concentration of toluene

272 at the ground surface, normalized by its daily maximum value, is displayed at these times in Fig. 5 for the  
273 winter scenario and Fig. 6 for the summer scenario.

274 The dispersion of any atmospheric constituent is driven essentially by the wind field and atmospheric  
275 stability. It is noteworthy that the footprint of toluene emissions is clearly different in the ADMS and  
276 WRF simulations, especially at 1400 UTC and 1800 UTC. This suggests that the observations from the  
277 ground monitoring station that were used as input to ADMS might not be so representative for the area  
278 of interest. This can be explained by the station being located somewhat on the slope on the side of the  
279 small valley in which waste is piled up (see Fig. 1). At the location of the landfill site, a typical valley wind  
280 system develops. The wind is typically directed up the valley during daylight hours (*i.e.* daytime) and down  
281 the valley otherwise (*i.e.* nighttime). Up-valley winds are usually stronger than down-valley winds. This  
282 idealized picture of the flow does explain the location of the complaints for both scenarios. Indeed, most of  
283 the complaints occurred either in the early morning or in the evening (see Fig. 3).

284 The record of complaints reported by the community was analyzed along with results from both models.  
285 In that respect, WRF was found to outperform ADMS in predicting the location and timing of complaints.  
286 In particular, the plume of toluene simulated by ADMS does not move down the valley to FIR in the  
287 evening in both winter and summer, for which complaints were recorded (see Fig. 3). As mentioned above,  
288 this might be attributed to the observations from the ground monitoring station that were used as input to  
289 ADMS being not so representative for this area over complex terrain. Therefore, we decided to discard the  
290 results from ADMS to calculate a population exposure for the two scenarios.

## 291 5.2. Population exposure

292 Environmental risk assessment is commonly applied in waste management (Pollard *et al.*, 2006; Butt  
293 *et al.*, 2008). Risks associated with potential health impacts from exposures to landfill gas and particulate  
294 matter receive increasing attention (*e.g.* Macleod *et al.*, 2006). Relating simulated odour concentrations to  
295 population exposure is key to assessing odour impacts around a landfill site (*e.g.* Sarkar *et al.*, 2003).

296 The concentration of toluene at the ground surface simulated by WRF was combined with the density  
297 of the population. We defined population exposure as the product of the concentration and the population  
298 density, integrated over a period of time. This average exposure is calculated by dividing the integrated ex-  
299 posure by the time period of integration (see for instance Monn, 2001). The whole population was assigned  
300 to the built areas. The fraction of built area in each grid cell of the inner domain is displayed in Fig. 7. Most  
301 of the built areas lie in the valleys. Only a few habitations are located North-East of the landfill site.

302 The time integration periods coincide with the time periods used to optimally map the temporal reparti-  
303 tion of the complaints for the two scenarios in Fig. 3, namely from 0600 UTC to 1000 UTC (~ morning),

Fig. 4

Fig. 5

Fig. 6

Fig. 7

304 from 1000 UTC to 1500 UTC (~ afternoon), and from 1500 UTC to 2300 UTC (~ evening). Note that  
305 since perception of odour correspond to a much shorter time scale (on the order of a few seconds), the  
306 average exposure would considerably dilute short-term impact.

307 The average population exposure to odour pollution, normalized by its daily maximum value, is dis-  
308 played for the abovementioned time integration periods in Fig. 8 for both the winter and summer scenarios.  
309 Consistent with initial indications from the WRF simulations, complaints are mainly to be expected in both  
310 winter and summer in the evening in FIR. Population exposure is much lower in the morning and in the  
311 afternoon during wintertime while being moderate in LCF in the morning and in FIR in the afternoon  
312 during summertime.

Fig. 8

## 313 6. Summary and conclusions

314 The main goal of this study was to assess odour impacts around a landfill site located over complex  
315 terrain in order to provide information that could be subsequently used to identify management strategies  
316 to prevent olfactory nuisances in the residential neighbourhoods. This study consisted of several steps, as  
317 follows.

- 318 • Odour sources were identified during field sampling. The sampling was carried out on site in the vicinity  
319 of the working face to characterize the temporal fluctuations in emissions. Toluene was selected as a  
320 typical trace VOC that is representative of the source under investigation and has a high emission rate  
321 (see also Davoli *et al.*, 2003). A ‘generic’ daily emission profile (see Fig. 2) was derived by averaging  
322 toluene emissions over the period of the field sampling.
- 323 • Relationships between regional- and local-scale atmospheric dynamics and odour pollution events were  
324 identified by classifying weather types for the years 2002 to 2004. Data from operational radiosoundings  
325 was used at the regional scale while data from a ground meteorological monitoring station was used  
326 at the local scale. Data classification consisted of standard ascending hierarchical classification and K-  
327 mean cluster analysis to divide the dataset into classes (weather types). Our approach to weather-type  
328 classification is similar to the one used by Greene *et al.* (1999) and Kim Oanh *et al.* (2005). While being  
329 not surprising, we found that local data gives more information about the location of the complaints  
330 in space and time than regional data. As a result, small-scale atmospheric processes would need to be  
331 considered so as to achieve accurate prediction of odour events in the vicinity of the landfill site. The  
332 two most representative weather types that lead to the majority of complaints were characterized by  
333 high-pressure systems with no wind in both winter and summer. A representative day of each of the two  
334 weather types was identified in order to be simulated by both a non-steady state Gaussian model and a

335 full-physics meteorological model.

336 • Two days that are representative of the winter and summer scenarios were simulated with the ADMS and  
337 WRF models. The normalized emission profile of toluene was used as a reference emission profile for  
338 both scenarios. The footprint of toluene emissions was found to be significantly different in the ADMS  
339 and WRF simulations, especially in the afternoon and evening. The record of complaints reported by  
340 the community was analyzed along with results from both models. In that respect, WRF was found  
341 to outperform ADMS in predicting the location and timing of complaints. This might be attributed to  
342 the observations from the ground monitoring station that were used as input to ADMS being not so  
343 representative for this area over complex terrain. Indeed, the station is located somewhat on the slope on  
344 the side of the small valley in which waste is landfilled (see Fig. 1). As a consequence, we decided to  
345 discard the results from ADMS and use WRF to calculate a population exposure for the two scenarios.  
346 One has to bear in mind that advanced meteorological models (such as WRF) are usually not designed to  
347 be used in a friendly way by the user community. The methodology that we proposed based on a limited  
348 number of scenarios might be an efficient alternative to anticipate odour pollution events.

349 • The concentration of toluene at the ground surface simulated by WRF was combined with the density of  
350 the population in order to calculate a population exposure for the winter and summer scenarios. Since no  
351 census of the population was available at the local scale, we assumed it to be proportional to the density  
352 of built area. Consistent with initial indications from the WRF simulations, we found that complaints  
353 are mainly to be expected in both winter and summer in the evening in FIR. The ‘risk’ of complaint was  
354 much reduced in winter in the morning and in the afternoon. A moderate exposure was found in summer  
355 in LCF in the morning and in FIR in the afternoon. While the days that were simulated are representative  
356 of the winter and summer scenarios, they do not represent the whole range of meteorological conditions  
357 which can occur in these seasons. Still, results of this study are expected to provide helpful information  
358 to develop technical solutions for an effective management of landfill operations, which would reduce  
359 odour impacts within the surrounding communities.

360 One has to be aware of the limitations of the approach to odour assessment that we used. Odours are the  
361 result of a complex combination of several compounds. In our work we selected a single indicator com-  
362 pound (namely toluene) that is representative of the source under investigation and has a high emission rate.  
363 However, this might lead to an underestimation of the impact of odours since we tracked only one com-  
364 pound. It is unwise to superimpose the odour strength of several compounds with known odour strength in  
365 a multicomponent mixture. Indeed, individual odour components can mutually reinforce, weaken, or mask  
366 each other. The processes that drive the odour strength of the mixture require further investigation. Numer-

367 ical simulations using a binary mixture might be a first step to refine and expand the present investigation.  
368 Source apportionment methods could also be used in order to analyze the contribution of various emission  
369 categories to population exposure.

370 The population exposure that we computed is integrated over time. However, short-term peak exposures  
371 can be significantly higher. One could also take into account characteristics of the microenvironments in  
372 which people spend their time by using information on their activities. In considering population activities,  
373 one could differentiate the population into age groups, young and elderly people being more sensitive to  
374 air pollution, or take into account geographical factors such as the location of schools or the proximity of  
375 hospitals. Another route to refine our analysis is to translate odour concentration into odour perception by  
376 using empirical laws such as those discussed by Sarkar and Hobbs (2002).

### 377 **Acknowledgments**

378 This work was partly supported by the SATROD company, subsidiary of SITA France, in charge of the  
379 activities of the landfill site. Results of this study may not necessarily reflect the views of the SATROD  
380 company and those of SITA France, and no official endorsement should be inferred. We thank TERA  
381 Technologies for collecting data and for helpful discussions on techniques for VOC measurement. Cal-  
382 culations with the ADMS model were realized while CR was visiting the Integrated Waste Management  
383 Centre, School of Applied Sciences, Cranfield University, UK. We would like to thank P. J. Longhurst for  
384 having hosted this visit and for fruitful discussions. Finally, we thank the four anonymous reviewers for  
385 their enlightened reviews and helpful suggestions, which have led to an improved manuscript.

### 386 **References**

- 387 Allen, M. R., Braithwaite, A., Hills, C. C., 1997. Trace organic compounds in landfill gas at seven U.K. waste disposal sites. *Envi-*  
388 *ron. Sci. Technol.* 31, 1054–1061.
- 389 Atkinson, R., 2000. Atmospheric chemistry of VOCs and NO<sub>x</sub>. *Atm. Environ.* 34, 2063–2101.
- 390 Aubrun, S., Leitzl, B., 2004. Unsteady characteristics of the dispersion process in the vicinity of a pig barn. Wind tunnel experiments  
391 and comparison with field data. *Atm. Environ.* 38, 81–93.
- 392 Bendz, D., Bengtsson, L., 1996. Evaporation from an active, uncovered landfill. *J. Hydrol.* 182, 143–155.
- 393 Berman, N. S., Boyer, D. L., Brazel, A. J., Brazel, S. W., Chen, R.-R., Fernando, H. J. S., Fitch, M. J., 1995. Synoptic classification  
394 and physical model experiments to guide field studies in complex terrain. *J. Appl. Meteor.* 34, 719–730.
- 395 Brosseau, J., Heitz, M., 1994. Trace gas compound emissions from municipal landfill sanitary sites. *Atm. Environ.* 28, 285–293.
- 396 Brulfert, G., Chemel, C., Chaxel, E., Chollet, J.-P., Jouve, B., Villard, H., 2006. Assessment of 2010 air quality in two Alpine valleys  
397 from modelling: Weather type and emission scenarios. *Atm. Environ.* 40, 7893–7907.

- 398 Butt, T. E., Lockley, E., Oduyemi, K. O. K., 2008. Risk assessment of landfill disposal sites – state of the art. *Waste Management* 28,  
399 952–964.
- 400 Carruthers, D. J., Holroyd, R. J., Hunt, J. C. R., Weng, W.-S., Robins, A. G., Apsley, D. D., Thompson, D. J., Smith, F. B., 1994.  
401 UK-ADMS: A new approach to modelling dispersion in the earth's atmospheric boundary layer. *J. Wind Eng. Ind. Aerodyn.* 52,  
402 139–153.
- 403 Collins, W. D., Rasch, P. J., Boville, B. A., Hack, J. J., McCaa, J. R., Williamson, D. L., Briegleb, B. P., Bitz, C. M., Lin, S.-J., Zhang,  
404 M., 2006. The formulation and atmospheric simulation of the Community Atmosphere Model Version 3 (CAM3). *J. Climate* 19,  
405 2144–2161.
- 406 Davoli, D., Gangai, M. L., Morselli, L., Tonelli, D., 2003. Characterisation of odorants emissions from landfills by SPME and GC/MS.  
407 *Chemosphere* 51, 357–368.
- 408 De Melo Lisboa, H., Guillot, J.-M., Fanlo, J.-L., Le Cloirec, P., 2006. Dispersion of odorous gases in the atmosphere – Part I: Modeling  
409 approaches to the phenomenon. *Sci. Tot. Env.* 361, 220–228.
- 410 Deardorff, J. W., 1980. Stratocumulus-capped mixed layers derived from a three-dimensional model. *Boundary-Layer Meteorol.* 18,  
411 495–527.
- 412 Eder, B. K., Davis, J. M., Bloomfield, P., 1994. An automated classification scheme designed to better elucidate the dependence of  
413 ozone on meteorology. *J. Appl. Meteor.* 33, 1182–1199.
- 414 Ek, M. B., Mitchell, K. E., Lin, Y., Rogers, E., Grunmann, P., Koren, V., Gayno, G., Tarpley, J. D., 2003. Implementation of Noah land  
415 surface model advances in the National Centers for Environmental Prediction operational mesoscale Eta model. *J. Geophys. Res.*  
416 108, D22/8851.
- 417 Greene, J. S., Kalkstein, L. S., Ye, H., Smoyer, K., 1999. Relationships between synoptic climatology and atmospheric pollution at 4  
418 US cities. *Theor. Appl. Clim.* 62, 163–174.
- 419 Grell, G. A., Dévényi, D., 2002. A generalized approach to parameterizing convection combining ensemble and data assimilation  
420 techniques. *Geophys. Res. Lett.* 121, D14/1693.
- 421 Hobbs, S. E., Longhurst, P., Sarkar, U., Sneath, R. W., 2000. Comparison of dispersion models for assessing odour from municipal  
422 solid wastes. *Waste Manage. Res.* 18, 420–428.
- 423 Hong, S.-Y., Noh, Y., Dudhia, J., 2006. A new vertical diffusion package with an explicit treatment of entrainment processes.  
424 *Mon. Weath. Rev.* 134, 2318–2341.
- 425 Kim, K.-H., Choi, Y. J., Jeon, E. C., Sunwoo, Y., 2005. Characterization of malodorous sulfur compounds in landfill gas. *Atm. Environ.*  
426 39, 1103–1112.
- 427 Kim Oanh, N. T., Chutimon, P., Ekbordin, W., Supat, W., 2005. Meteorological pattern classification and application for forecasting  
428 air pollution episode potential in a mountain-valley area. *Atm. Environ.* 39, 1211–1225.
- 429 Lebart, L., Morineau, A., Piron, M., 1997. *Statistique Exploratoire Multidimensionnelle (Ouvrage référence de SPAD)*. Dunod, Paris,  
430 France, 439 pp.
- 431 Macleod, C., Duarte-Davidson, R., Fisher, B., Ng, B., Willey, D., Shi, J. P., Martin, I., Drew, G., Pollard, S., 2006. Modeling human  
432 exposures to air pollution control (APC) residues released from landfills in England and Wales. *Atm. Environ.* 32, 500–509.
- 433 McIntyre, A., 2000. Application of dispersion modelling to odour assessment: a practical tool or a complex trap? *Wat. Sci. Tech.* 41,  
434 81–88.
- 435 Monn, C., 2001. Exposure assessment of air pollutants: a review on spatial heterogeneity and indoor/outdoor/personal exposure to

- 436 suspended particulate matter, nitrogen dioxide and ozone. *Atm. Environ.* 35, 1–32.
- 437 Mussio, P., Gnyp, A. W., Henshaw, P. F., 2001. A fluctuating plume dispersion model for the prediction of odour-impact frequencies  
438 from continuous stationary sources. *Atm. Environ.* 35, 2955–2962.
- 439 Nanni, A., Brusasca, G., Calori, G., Finardi, S., Tinarelli, G., Zublena, M., Agnesod, G., Pession, G., 2004. Integrated assessment of  
440 traffic impact in an Alpine region. *Sci. Tot. Env.* 334-335, 465–471.
- 441 Nicell, J. A., 2003. Expressions to relate population responses to odor concentration. *Atm. Environ.* 37, 4955–4964.
- 442 Ormerod, R., 2001. Improving odour assessment by using better dispersion models: some examples. *Wat. Sci. Tech.* 44, 149–156.
- 443 Pollard, S. J. T., Smith, R., Longhurst, P. J., Eduljee, G. H., Hall, D., 2006. Recent developments in the application of risk analysis to  
444 waste technologies. *Environ. Int.* 32, 1010–1020.
- 445 Riesenmey, C., 2008. Assessment of the impact of a municipal solid waste landfill site on the air quality in surrounding communities:  
446 approach using statistical analysis and determinist modelling. PhD Thesis No. 473 SGE, Ecole Nationale Supérieure des Mines de  
447 Saint-Etienne, Saint-Etienne, France.
- 448 Sarkar, U., Hobbs, S. E., 2002. Odour from municipal solid waste (MSW) landfills: a study on the analysis of perception. *Environ. Int.*  
449 27, 655–662.
- 450 Sarkar, U., Hobbs, S. E., Longhurst, P., 2003. Dispersion of odour: a case study with a municipal solid waste landfill site in North  
451 London, United Kingdom. *J. Env. Manage.* 68, 153–160.
- 452 Schaubberger, G., Piringer, M., Petz, E., 2001. Separation distance to avoid odour nuisance due to livestock calculated by the Austrian  
453 odour dispersion model (AODM). *Agr. Ecosyst. Environ.* 87, 13–28.
- 454 Sfetsos, A., Vlachogiannis, D., Gounaris, N., Stubos, A. K., 2005. On the identification of representative samples from large data sets,  
455 with application to synoptic climatology. *Theor. Appl. Clim.* 82, 177–182.
- 456 Skamarock, W. C., Klemp, J. B., Dudhia, J., Gill, D. O., Barker, D. M., Duda, M. G., Huang, X.-Y., Wang, W., Powers, J. G., 2008. A  
457 description of the Advanced Research WRF Version 3. NCAR Technical Note NCAR/TN-475+STR, NCAR, Boulder, CO, USA,  
458 125 pp.
- 459 Spokas, K., Bogner, J., Chanton, J. P., Morcet, M., Aran, C., Graff, C., Moreau-Le Golvan, Y., Hebe, I., 2006. Methane mass balance  
460 at three landfill sites: What is the efficiency of capture by gas collection systems? *Waste Management* 26, 516–525.
- 461 Stauffer, D. R., Seaman, N., 1990. Use of Four-Dimensional Data Assimilation in a limited-area mesoscale model. Part I: Experiments  
462 with synoptic-scale data. *Mon. Weath. Rev.* 118, 1250–1277.
- 463 Stuetz, R., Frechen, F.-B. (Eds.), 2001. *Odours in Wastewater Treatment: Measurement, Modelling and Control*. IWA Publishing,  
464 London, UK, 456 pp.
- 465 Tagaris, E., Sotiropoulou, R.-E. P., Pilinis, C., Halvadakis, C. P., 2003. A methodology to estimate odors around landfill sites: the use  
466 of methane as an odor index and its utility in landfill siting. *J. Air & Waste Manage. Asso.* 53, 629–634.
- 467 Termonia, A., Termonia, M., 1999. Characterisation and on-site monitoring of odorous organic compounds in the environment of a  
468 landfill site. *Int. J. Environ. Anal. Chem.* 73, 43–57.
- 469 Thompson, G., Field, P. R., Hall, W. D., Rasmussen, R. M., 2006. A new bulk microphysical parameterization for WRF (& MM5).  
470 In: *Proc. of the 8th WRF Users' Workshop*. 11–15 June 2007, Boulder, CO, USA.
- 471 Thompson, G., Rasmussen, R. M., Manning, K., 2004. Explicit forecasts of winter precipitation using an improved bulk micro-physics  
472 scheme. Part I: Description and sensitivity analysis. *Mon. Weath. Rev.* 132, 519–542.
- 473 Tirabassi, T., Nassetti, S., 1999. The representative day. *Atm. Environ.* 33, 2427–2434.

474 Yang, G., Hobson, J., 2000. Odour nuisance – advantages and disadvantages of a quantitative approach. *Wat. Sci. Tech.* 41, 97–106.

## Tables

Table 1. Description of the regional classes (weather types) and repartition of the complaints within the classes

Regional atmospheric stability	Description of the regional classes			Number of		
	Risk of precipitation	Wind speed / direction		Days	Complaints (□)	Complaints per 100 days
		Above 1500 m	Below 1500 m			
Moderately stable	Very low	Very low / NA	Very low / NA	121	19 (2)	15.7
Unstable	Very low	Very low / NA	Very low / NA	89	14 (1)	15.7
Neutral	Low	Very low / NA	Very low / NA	182	12 (7)	6.6
Very stable	Low	Very low / NA	Very low / NA	113	7 (2)	6.2
Very stable	Moderate	Moderate / N	Moderate / S	105	5 (0)	4.8
Moderately stable	Moderate	Moderate / E	Strong / N-W	77	3 (2)	3.9
Moderately stable	High	Strong / N	Low / NA	104	4 (2)	3.8
Moderately stable	High	Moderate / S-S-E	Moderate / N-N-W	82	3 (1)	3.7
Moderately stable	Moderate	Moderate / S	Low / NA	72	2 (0)	2.8
Moderately stable	Moderate	Strong / E	Low / NA	81	2 (0)	2.5
Very stable	Moderate	Low / N	Low / NA	44	0 (0)	0.0

(□) *Complaints of olfactory nuisances that were identified as being related to special manipulations of waste (e.g. opening of a new cell).*

## Figure Captions

Fig. 1. Orography of the landfill site area and its surroundings. The attached grey scale indicates altitude in meters above mean sea level (a.m.s.l.). Solid lines corresponds to major highways. The three major towns around the landfill site and the ground meteorological monitoring station located on the site are marked as ★ and ○, respectively. The stipple-filled area represents the exploited cells.

Fig. 2. Normalized daily profile of odour emission (—) derived from toluene concentration measurements close to the working face, averaged over the period of the field sampling. The profile was smoothed by calculating a running average of the raw data (○) and by using a curve-fit procedure to get an analytical function.

Fig. 3. Same caption as Fig. 1. The clocks indicate the temporal repartition of the complaints for the local-scale winter (in white) and summer (in grey) classes, which contain the majority of complaints.

Fig. 4. Time series of wind speed (upper panel) and direction (lower panel), observed (● symbols) and predicted by WRF (solid lines) at the location of the ground meteorological monitoring station for the summer (in red) and winter (in blue) scenarios.

Fig. 5. Color-filled contours of the concentration of toluene at the ground surface, normalized by its daily maximum value, simulated by ADMS (left-hand column) and WRF (right-hand column) in the inner domain, at 1000 UTC ((a) and (b)), 1400 UTC ((c) and (d)), and 1800 UTC ((e) and (f)) for the winter scenario. The 10-m horizontal wind vector observed at the location of the ground meteorological monitoring station is superimposed on the ADMS plots. The 10-m horizontal wind vectors predicted by WRF are superimposed on the WRF plots. Solid lines indicate the topography with 10-m interval contours. Note that the extent of the domain is slightly different in ADMS and WRF because of different projection systems. The three major towns around the landfill site are marked as ★.

Fig. 6. Same caption as Fig. 5 for the summer scenario.

Fig. 7. Raster representation of the fraction of built area in each grid cell of the inner domain used for the WRF simulations. Solid lines indicate the topography with 10-m interval contours.

Fig. 8. Color-filled contours of the average population exposure to odour pollution derived from the concentration of toluene at the ground surface simulated by WRF in the inner domain and population density, for the winter (left-hand column) and summer (right-hand column) scenarios, integrated from 0600 UTC to 1000 UTC ((a) and (b)), from 1000 UTC to 1500 UTC ((c) and (d)), and from 1500 UTC to 2300 UTC ((e) and (f)). The average population exposure is normalized by its daily maximum value. Solid lines indicate the topography with 10-m interval contours. The three major towns around the landfill site are marked as ★.

## Figures

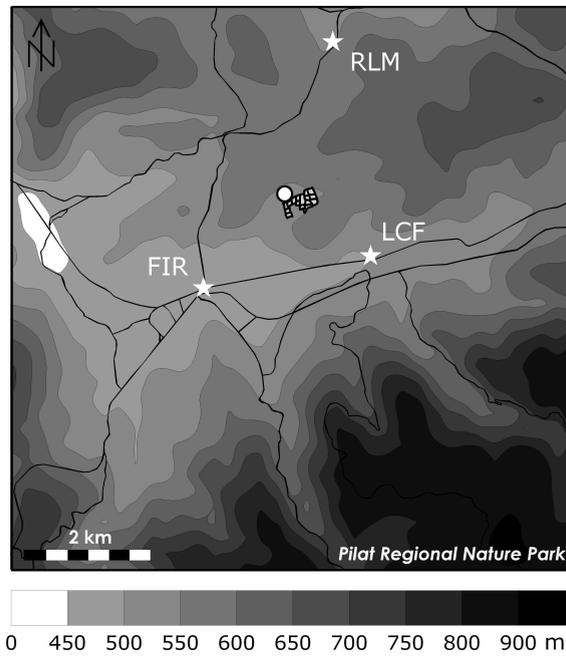


Fig. 1. Orography of the landfill site area and its surroundings. The attached grey scale indicates altitude in meters above mean sea level (a.m.s.l.). Solid lines corresponds to major highways. The three major towns around the landfill site and the ground meteorological monitoring station located on the site are marked as ★ and ○, respectively. The stipple-filled area represents the exploited cells.

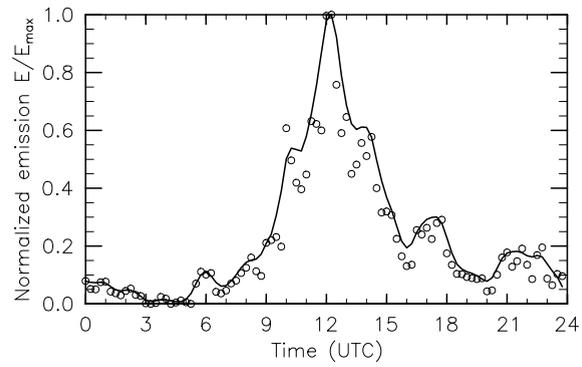


Fig. 2. Normalized daily profile of odour emission (—) derived from toluene concentration measurements close to the working face, averaged over the period of the field sampling. The profile was smoothed by calculating a running average of the raw data (o) and by using a curve-fit procedure to get an analytical function.

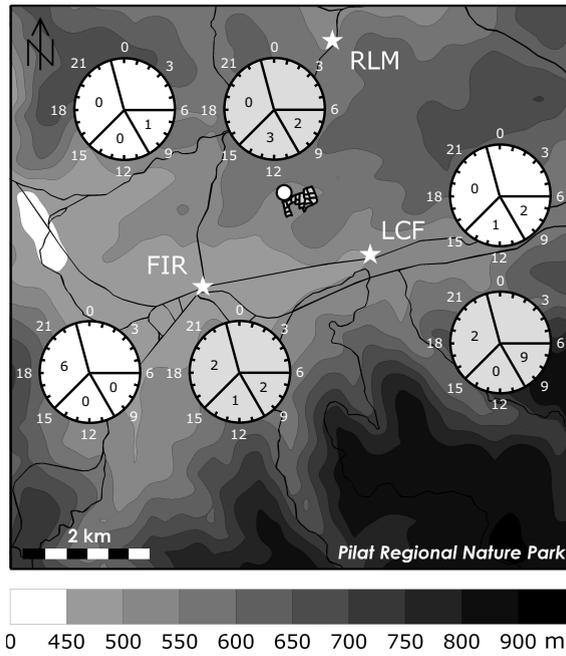


Fig. 3. Same caption as Fig. 1. The clocks indicate the temporal repartition of the complaints for the local-scale winter (in white) and summer (in grey) classes, which contain the majority of complaints.

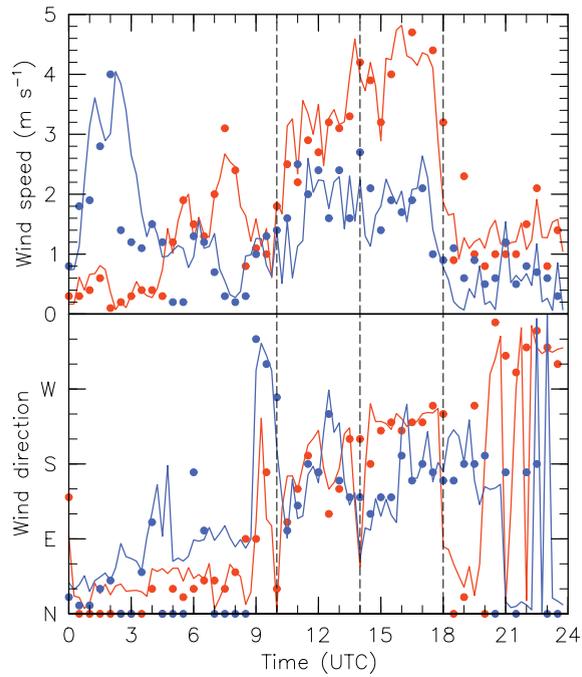


Fig. 4. Time series of wind speed (upper panel) and direction (lower panel), observed (● symbols) and predicted by WRF (solid lines) at the location of the ground meteorological monitoring station for the summer (in red) and winter (in blue) scenarios.

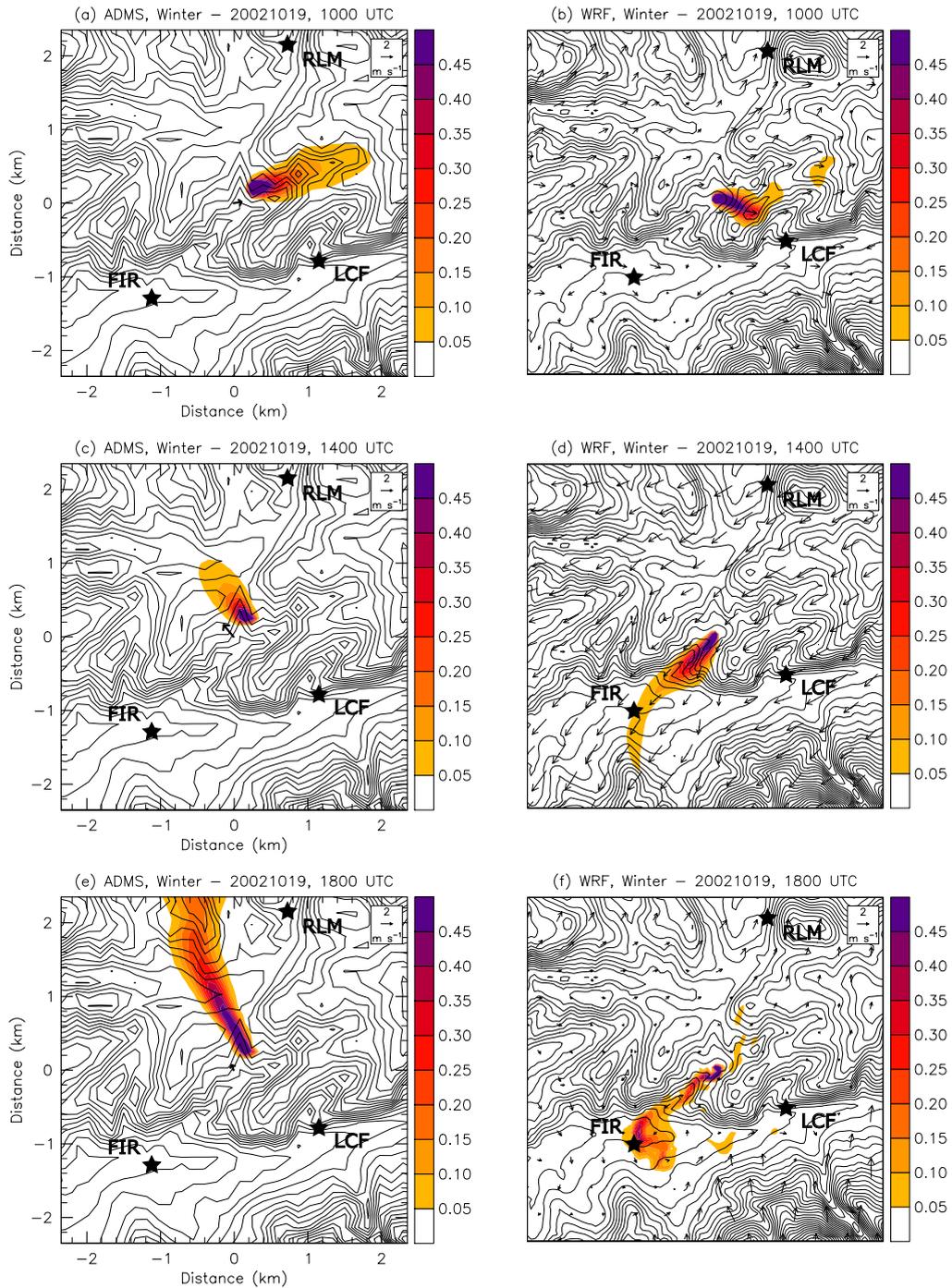


Fig. 5. Color-filled contours of the concentration of toluene at the ground surface, normalized by its daily maximum value, simulated by ADMS (left-hand column) and WRF (right-hand column) in the inner domain, at 1000 UTC ((a) and (b)), 1400 UTC ((c) and (d)), and 1800 UTC ((e) and (f)) for the winter scenario. The 10-m horizontal wind vector observed at the location of the ground meteorological monitoring station is superimposed on the ADMS plots. The 10-m horizontal wind vectors predicted by WRF are superimposed on the WRF plots. Solid lines indicate the topography with 10-m interval contours. Note that the extent of the domain is slightly different in ADMS and WRF because of different projection systems. The three major towns around the landfill site are marked as ★.

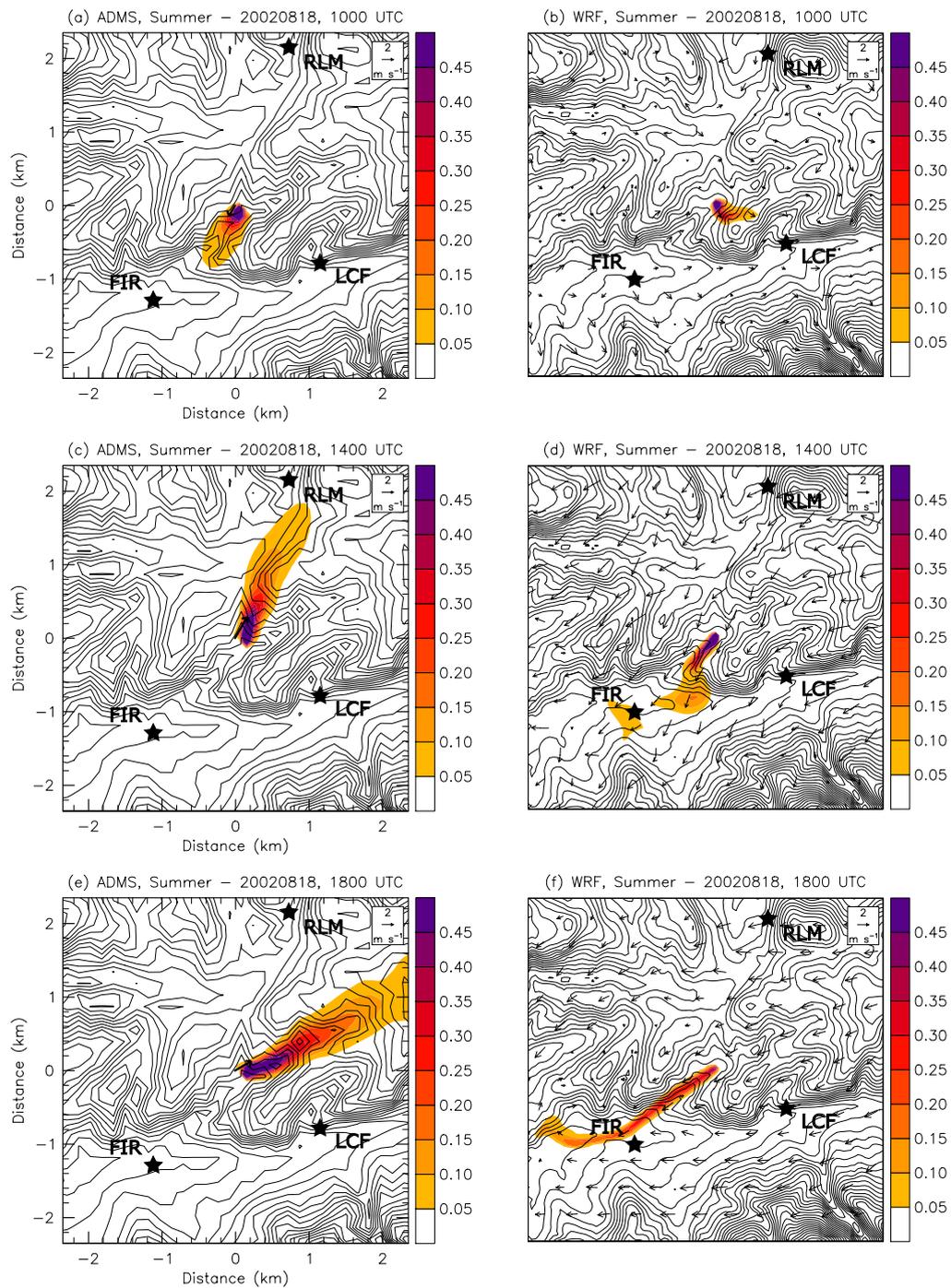


Fig. 6. Same caption as Fig. 5 for the summer scenario.

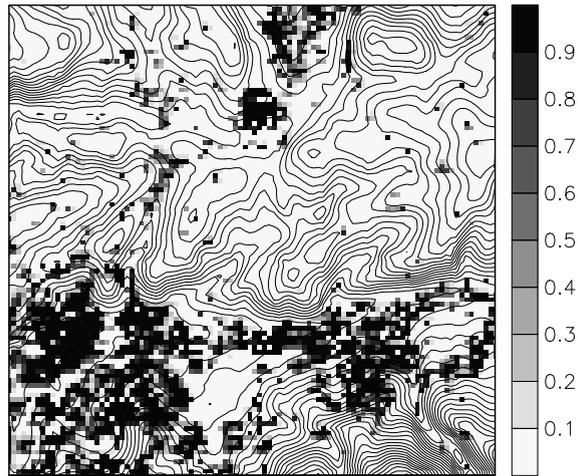


Fig. 7. Raster representation of the fraction of built area in each grid cell of the inner domain used for the WRF simulations. Solid lines indicate the topography with 10-m interval contours.

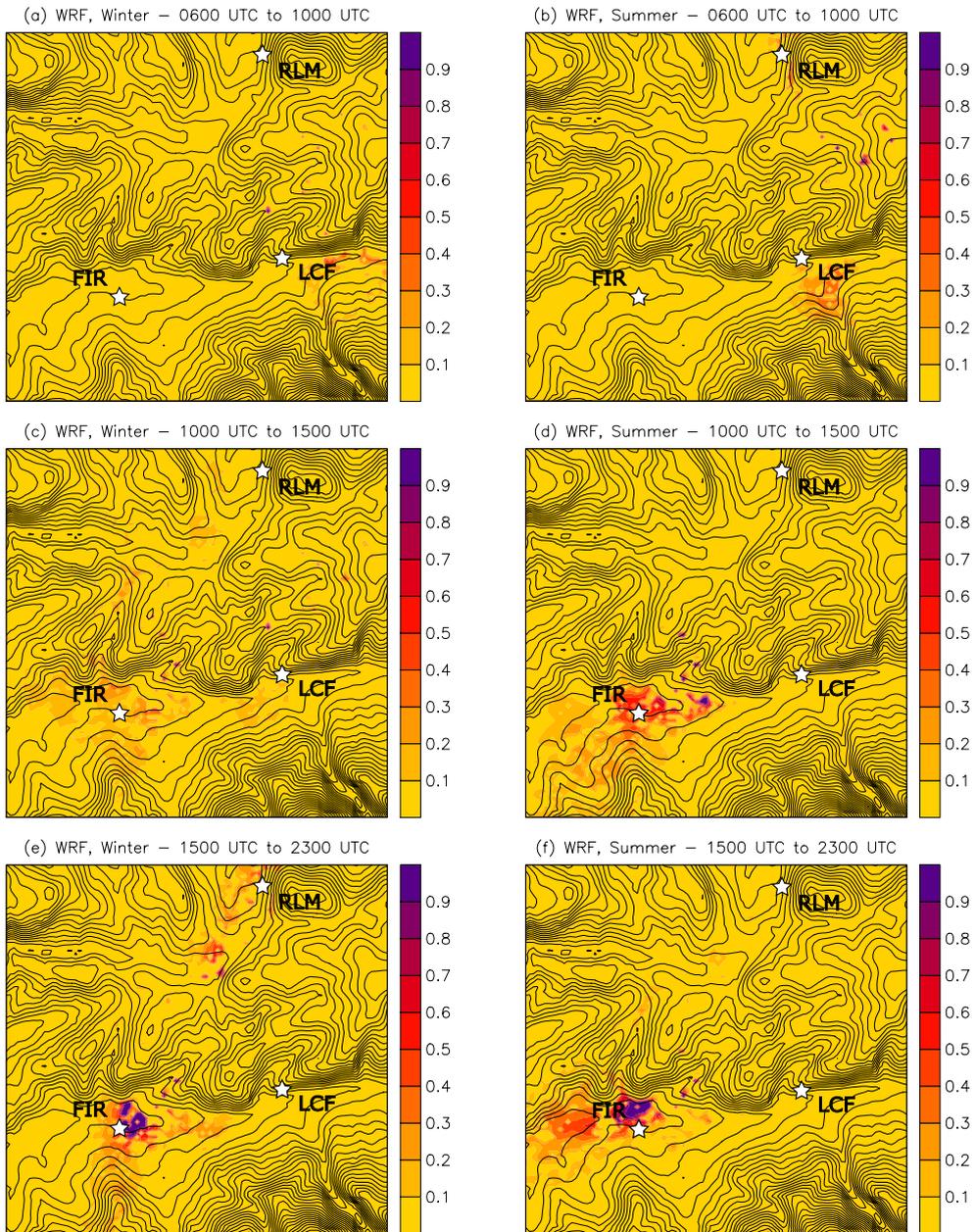


Fig. 8. Color-filled contours of the average population exposure to odour pollution derived from the concentration of toluene at the ground surface simulated by WRF in the inner domain and population density, for the winter (left-hand column) and summer (right-hand column) scenarios, integrated from 0600 UTC to 1000 UTC ((a) and (b)), from 1000 UTC to 1500 UTC ((c) and (d)), and from 1500 UTC to 2300 UTC ((e) and (f)). The average population exposure is normalized by its daily maximum value. Solid lines indicate the topography with 10-m interval contours. The three major towns around the landfill site are marked as ★.